



U.S. Army Corps of Engineers New England District

FINAL REMEDIAL INVESTIGATION REPORT AREA OF CONTAMINATION (AOC) 69W DEVENS, MASSACHUSETTS

VOLUME I OF II TEXT SECTIONS 1 THROUGH 10 AND FIGURES AND TABLES

CONTRACT DAAA-31-94-D-0061 DELIVERY ORDER NUMBER 0001

U.S. ARMY CORPS OF ENGINEERS NEW ENGLAND DISTRICT CONCORD, MASSACHUSETTS

August 1998

2798082 HLAP



RECEIVED AUG 2 4 1998

FINAL

REMEDIAL INVESTIGATION REPORT AREA OF CONTAMINATION (AOC) 69W DEVENS, MASSACHUSETTS

VOLUME I OF H TEXT SECTIONS 1 THROUGH 10 AND FIGURES AND TABLES

CONTRACT DAAA-31-94-D-0061 DELIVERY ORDER NUMBER.0001

Prepared for:

U.S. Army Corps of Engineers New England District Concord, Massachusetts

Prepared by:

Harding Lawson Associates Portland, ME Project No. 09144-03

August 1998

TABLE OF CONTENTS

Sect	ion		Title	-	Pag	e No.
EXI	ECUT	IVE SUN	IMARY			ES-1
1.0	INTR	ODUCT	ON			. 1-1
	1.1	PURPOS	E AND SCOPE			. 1-1
	1.2	REPORT	ORGANIZATION			1-3
	1.3	PROJEC	OBJECTIVES			1-4
	1.4	PROJEC	Г APPROACH			1-4
			Project Operations Plan			1-4
		1.4.2	Task Order Work Plan	•••	•••	1-6
2.0	INST	ALLATIC	N DESCRIPTION			2-1
	2.1	HISTORY	ζ			2-1
			L SETTING			2-3
			Climate			2-3
			Vegetation			2-4
			Ecology			2-5
		2.2.4				2-6
		2.2.5	Soils			2-7
		2.2.6	Surficial Geology			2-10
		2.2.7	Bedrock Geology			2-11
		2.2.8	Regional Hydrogeology	•••		2-13
3.0	ANAL	YTICAL	PROGRAM			. 3-1
0.00			NALYTICAL METHODS			
		3.1.1	Instrument Calibration			3-2
		3.1.2	Sample Preparation and Analysis			3-3
		3.1.3	Target Compound Concentrations Calculations			3-4
		3.1.4	Field Documentation Procedures			3-5
		3.1.5	Field Analytical Quality Control			3-5

TABLE OF CONTENTS (continued)

Title

		3.1.6 Method Detection Limits and Data Qualifiers 3-7
	3.2	OFF-SITE LABORATORY ANALYTICAL PARAMETERS
		3.2.1 Off-Site Laboratory Certification
		3.2.2 Off-Site Laboratory Methods Quality Control 3-9
		3.2.3 Data Reduction, Validation, and Reporting 3-9
		3.2.4 Data Reporting 3-11
		3.2.5 Field Quality Control Samples 3-11
		3.2.6 Off-Site Analytical Data Quality Evaluation 3-13
	3.3	CHEMICAL DATA MANAGEMENT
		3.3.1 Sample Tracking System 3-14
		3.3.2 Installation Restoration Data Management Information
		System
	3.4	1997 SUPPLEMENTAL ANALYSES 3-15
4.0	۸P	PLICABLE OR RELEVANT AND APPROPRIATE REQUIREMENTS
-1.0		ENTIFICATION
		CHEMICAL-SPECIFIC ARARS
	3.1	4.1.1 Groundwater
		4.1.2 Soil
		4.1.3 Massachusetts Contingency Plan
	47	LOCATION-SPECIFIC ARARS
		ACTION-SPECIFIC ARARS
	4.4	BACKGROUND CONCENTRATIONS
5.0		69W REMEDIAL INVESTIGATION
	5.1	SITE BACKGROUND AND CONDITIONS 5-1
	5.2	SUMMARY OF PREVIOUS INVESTIGATIONS
		5.2.1 ADL AREE 69 Evaluation (AREE 69W) 5-2

Section

9144-03

Page No.

TABLE OF CONTENTS (continued)

Section

Title

Page No.

	5.2.1.1 Surface Soil 5-3	
	5.2.1.2 Subsurface Soil	
	5.2.1.3 Groundwater 5-4	
	5.2.1.4 Surface Water and Sediment 5-5	
5.3	REMEDIAL INVESTIGATION PROGRAM 5-6	
	5.3.1 Technical Objectives 5-6	
	5.3.1.1 Geophysical Survey 5-6	
	5.3.1.2 TerraProbe SM and Soil Borings	
	5.3.1.3 Monitoring Wells 5-6	
	5.3.1.4 Piezometer Installation	
	5.3.1.5 Test Pitting 5-6	
	5.3.1.6 Field Laboratory Analysis 5-7	
	5.3.1.7 Off-Site Laboratory Analytical Sampling 5-7	
	5.3.1.8 Air Sampling 5-7	
	5.3.2 Data Quality Objectives 5-8	
5.4	SUMMARY OF REMEDIAL INVESTIGATION FIELD PROGRAMS 5-11	
	5.4.1 Surficial Geophysical Survey 5-13	
	5.4.2 TerraProbe sm 5-14	
	5.4.3 Test Pitting 5-15	
	5.4.4 Soil Borings and Sampling 5-15	
	5.4.5 Monitoring Well Installation 5-16	
	5.4.6 Monitoring Well Development 5-16	
	5.4.7 Groundwater Sampling 5-17	
	5.4.8 In-Situ Hydraulic Conductivity Testing and Water Level	
	Measurement 5-18	
	5.4.9 Sediment and Toxicity Test Sampling 5-18	
	5.4.10 Indoor Air Sampling 5-19	
	5.4.11 Equipment Decontamination	

TABLE OF CONTENTS (continued)

Secti	ion Title Page	No.
	5.4.12 Investigation-Derived Waste	5-20
	5.4.13 Location and Elevation Survey	5-20
	5.5 SUMMARY OF REMOVAL ACTION AT AOC 69W	5-21
6.0	SITE HYDROLOGY, GEOLOGY AND GROUNDWAT	ΓER
	CHARACTERIZATION	6-1
	6.1 SITE HYDROLOGY	6-1
	6.2 SITE GEOLOGY	6-2
	6.2.1 Overburden Soils	6-2
	6.2.2 Bedrock Geology	6-3
	6.2.3 Site Geology Interpretation Summary	6-3
	6.3 HYDROGEOLOGY	6-4
	6.3.1 Site Hydrogeology Interpretation Summary	6-5
7.0	NATURE AND DISTRIBUTION OF DETECTED S	ITE
	CONTAMINANTS	7-1
	7.1 APPROACH TO CONTAMINATION ASSESSMENT	7-1
	7.1.1 Tentatively Identified Compounds/Non-Project Analyte	
	List Compounds	7-2
	7.1.1.1 TICs Detected in Samples from AOC 69W	7-3
	7.1.2 Potential Laboratory and Sampling Contaminants	7-5
	7.1.3 Analytical Data Accuracy and Precision	7-7
	7.2 INVESTIGATIONS AT AOC 69W	7-11
	7.2.1 1994 ADL AREE 69W Investigation	7-11
		7-11
	7.2.1.2 Subsurface Soil	7-11
	7.2.1.3 Groundwater	7-12
	7.2.1.4 Surface Water and Sediment	7-13

TABLE OF CONTENTS (continued)

 7.2.2 AOC 69W RI and Removal Action Soils Results	
 7.2.2.1 RI Field Analytical Soil Results	
 7.2.2.2 RI Off-Site Analytical Soils Results	
 7.2.2.4 Summary of Soil Impacts	
 7.2.3 AOC 69W RI Groundwater Results	
 7.2.3.1 RI Field Analytical Groundwater Results 7.2.3.2 RI Groundwater Off-Site Laboratory Analytical Sample Results	7-20
 7.2.3.2 RI Groundwater Off-Site Laboratory Analytical Sample Results	7-21
Sample Results	7-21
Sample Results	
 7.2.3.3 Summary of Groundwater Impacts 7.2.4 AOC 69W RI Sediment Results 8.0 CONTAMINANT FATE AND TRANSPORT 8.1 COMPOUND PROPERTIES AND TRANSPORT PROCESSES 8.1.1 Physical and Chemical Properties Significant to Fate and Transport	7-22
 8.0 CONTAMINANT FATE AND TRANSPORT	7-26
 8.1 COMPOUND PROPERTIES AND TRANSPORT PROCESSES 8.1.1 Physical and Chemical Properties Significant to Fate and Transport	7-27
 8.1 COMPOUND PROPERTIES AND TRANSPORT PROCESSES 8.1.1 Physical and Chemical Properties Significant to Fate and Transport	
 8.1.1 Physical and Chemical Properties Significant to Fate and Transport	. 8-1
Transport	8-1
8.1.2 General Transport and Attenuation Processes 8.2 FATE AND TRANSPORT OF CONTAMINANTS DETECTED AT AOC	
8.1.2 General Transport and Attenuation Processes 8.2 FATE AND TRANSPORT OF CONTAMINANTS DETECTED AT AOC	8-2
	8-3
69W	
	8-7
8.3 SITE CONCEPTUAL MODEL	
9.0 BASELINE HUMAN HEALTH RISK ASSESSMENT	. 9-1
9.1 Overview	9-1
9.2 HUMAN HEALTH RISK ASSESSMENT	
9.2.1 Selection of Chemicals of Potential Concern	9-5
9.2.1.1 Identification and Selection of Analytical	
Data	. 9-5
9.2.1.2 Data Summary Procedures	10.00
9.2.1.3 Data Screening Procedures :	

TABLE OF CONTENTS

(continued)

-						
s	0	n	t 1		73	
0	c	•	Lł	U	11	
-	-	-	-		-	

Title

Page No.

	sure Assessment	9-12
9.2.2	.1 Exposure Pathways	9-12
9.2.2	.2 Estimation of Exposure	9-16
9.2.3 Toxic		9-22
	Characterization	9-25
9.2.4	.1 Risk Characterization Methods	9-25
9.2.4	.2 Risk Characterization Results	9-27
9.2.5 Evalu	ation of Uncertainty	9-30
		9-36
0.2 Ecological Disk A		9-37
	SSESSMENT	9-37
	Characterization	9-30
	.1 Vegetative Cover	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1
	.2 Wetland and Drainage Ditch Habitat	9-40
	.3 Wildlife Habitat Characterization	9-40
	.4 Rare, Threatened, and Endangered Species	9-40
	em Formulation	9-41
	.1 Identification of Receptors	9-41
	.2 Identification of Exposure Pathways	9-42
	.3 Identification of Endpoints	9-43
9.3.3 Haza	rd Assessment and Selection of Contaminants of	
Pote	ntial Concern (CPCs)	9-44
9.3.3	.1 Surface Soil	9-46
	.2 Upgradient Sediment	9-46
	.3 Downgradient Sediment	9-46
9.3.3	.4 Groundwater	9-46
	sure Assessment	9-46
the second se	.1 Calculation of Exposure Point Concentrations .	9-47

TABLE OF CONTENTS

(continued)

\sim			1.8			
	a	n	Ť٦	~	T1	
0		6	ы	U	n	

-		. 1		
	11	t i	И	0
x	x	L	Ŀ	-

Page No.

9.3.4.3 Terrestrial Plants and Invertebrates)-47)-50)-50)-51
	-50
9.3.4.4 Aquatic Receptors	
	-51
	-51
9.3.5.2 Terrestrial Plants and Invertebrates	-52
9.3.5.3 Aquatic Receptors	-52
	-54
	-54
9.3.6.2 Terrestrial Plants	-56
	-57
	-57
	-60
	-62
10.0 CONCLUSIONS AND RECOMMENDATIONS	0-1
	0-1
	0-6

GLOSSARY OF ACRONYMS AND ABBREVIATIONS

REFERENCES

APPENDICES

APPENDIX A	EXPLORATION LOCATIONS
APPENDIX B	MONITORING WELL AND PIEZOMETER
	CONSTRUCTION DIAGRAMS

W003971.080

TABLE OF CONTENTS

(continued)

Section	Title	Page No.		
APPENDIX C	GEOPHYSICAL INVESTIGATION D	ATA AND ANALYSIS		
APPENDIX D	QUALITY CONTROL RESULTS AN	D ASSESSMENT		
D-1	1995 ON-SITE AND OFF-SITE LABO	DRATORY DATA		
D-2	1996 ON-SITE LABORATORY DATA	A		
D-3	1996 OFF-SITE LABORATORY DAT	ΓA		
D-4	1997 OFF-SITE LABORATORY DAT	FA		
APPENDIX E	HYDROGEOLOGIC DATA			
E-1	IN-SITU HYDRAULIC CONDUCTIV	/ITY TESTING		
E-2	HYDRAULIC GRADIENT AND GRO	DUNDWATER FLOW		
	VELOCITY CALCULATIONS			
E-3	WILLOW BROOK HYDROLOGIC D	DATA		
APPENDIX F	WELL DEVELOPMENT RECORDS			
APPENDIX G	FIELD SAMPLE DATA RECORDS (GROUNDWATER)			
APPENDIX H	SURVEY DATA			
APPENDIX I	GEOTECHNICAL DATA (GRAIN SI	ZE DISTRIBUTION)		
APPENDIX J	PROJECT ANALYTE LIST			
APPENDIX K	CALCULATION OF BACKGROUND	CONCENTRATIONS		
APPENDIX L	OFF-SITE ANALYTICAL LABORAT	ORY DATA		
APPENDIX M	INDOOR AIR QUALITY REPORT			
APPENDIX N	SOIL REMOVAL ACTION MEMOR.	ANDUM		
APPENDIX O	HUMAN HEALTH RISK ASSESSME	ENT		
	O-1 VAPOR MIGRATION PATHW	VAY ANALYSIS		
	O-2 CALCULATIONS			
	O-3 HHRA SHORT TOXICITY PR	OFILES		
APPENDIX P	ECOLOGICAL RISK ASSESSMENT			
	P-1 EXPOSURE AND EFFECTS A	ASSUMPTIONS		
	P-2 ECOLOGICAL RISK CALCUL	LATIONS		
	P-3 STATISTICAL ANALYSIS OF	TOXICITY TESTING		

TABLE OF CONTENTS (continued)

Section	Title		Page No.
COLUMN TO THE			

(MIDGE GROWTH) APPENDIX Q RESPONSE TO COMMENTS ON DRAFT REMEDIAL INVESTIGATION REPORT FOR AREA OF CONTAMINATION (AOC) 69W, DATED APRIL 1998 APPENDIX R USEPA - INDOOR AIR SAMPLING STUDY

LIST OF FIGURES

Figure

Title

- ES-1 RI Sampling Locations
- 2-1 Location of Devens, MA
- 2-2 Location of AOC 69W
- 2-3 General Soils Map
- 2-4 Bedrock Geology
- 2-5 Regional Aquifer Transmissivities
- 2-6 Regional Overburden Groundwater Flow Map
- 2-7 Regional Bedrock Groundwater Flow Map
- 5-1 Location of AOC 69W
- 5-2 ADL AREE 69W Exploration Locations
- 5-3 RI Sampling Locations
- 5-4 Air Sampling Locations
- 5-5 Soil Removal Action Excavation and Soil Sample Locations
- 6-1 Orientation of Geologic Cross Section A-A'
- 6-2 Interpretive Geologic Cross Section A-A'
- 6-3 Interpretive Water Table Elevation Contours
- 7-1 Analytes in Soils (0'-5' bgs) 1995 Field Analytical Results
- 7-2 Analytes in Subsurface Soils (6'-10' bgs) 1995 Field Analytical Results
- 7-3 TPHC Concentrations in Subsurface Soils (4'-10' bgs) Field Analytical Results
- 7-4 Analytes in Subsurface Soils 1996 Field Analytical Results
- 7-5 Organic Analytes in Surface Soils (0'-2' bgs) 1995 Off-Site Analytical Results
- 7-6 Organic Analytes in Subsurface Soils 1995 Off-Site Analytical Results
- 7-7 (East) EPH/VPH Concentrations in Soil 1997 Off-Site Analytical Results
- 7-7 (West) EPH/VPH Concentrations in Soil 1997 Off-Site Analytical Results
- 7-8 EPH Concentrations in Subsurface Soils Post Soil Removal
- 7-9 Selected PAL VOCs in Groundwater 1995 Field Analytical Results

W003971.080

9144-03

х

LIST OF FIGURES

(continued)

Figure	Title	

- 7-10 EPH/VPH and TPHC Concentrations in Groundwater 1995 and 1996 Off-Site Analytical Results
- 7-11 EPH/VPH, PAH, and VOC Concentrations in Groundwater 1997 Off-Site Analytical Results
- 7-12 Analytes in Sediment
- 7-13 VOC Concentrations in Indoor Air 1997 Off-Site Analytical Results
- 8-1 Conceptual Site Model Diagram
- 9-1 RI Sampling Locations
- 9-2 Conceptual Site Model Diagram
- 9-3 Contaminant Pathway Model For Ecological Receptors Exposed to Surface Soil, Sediment and Groundwater
- 9-4 Ecological Contaminants of Potential Concern Selection Process

LIST OF TABLES

Table	Title
ES-1	Summary of Human Health Risk Assessment
ES-2	Summary of Ecological Risk Assessment
3-1	Summary of Target Compounds and Reporting Limits
3-2	USAEC Data Flags and Qualifiers
4-1	Federal ARARS and TBC Guidance - Groundwater and Surface Water
4-2	State ARARS and TBC Guidance - Groundwater
4-3	ARARS and TBC Guidance - Soil and Sediment
4-4	Potential Location-Specific ARARS at Devens
4-5	Potential Action-Specific Federal ARARS at Devens
5-1	Summary of Investigation Activities
5-2	ADL, AREE 69W Field Analytical Results
5-3	ADL, AREE 69W Soil Off-Site Analytical Results
5-4	ADL, AREE 69W Groundwater Off-Site Analytical Results
5-5	ADL, AREE 69W Surface Water Off-Site Analytical Results
5-6	ADL, AREE 69W Sediment Off-Site Analytical Results
5-7	Summary of Soil Borings
5-8	Monitoring Well Completion Details
6-1	Summary of Water Level Elevation Data
6-2	Summary of In-situ Hydraulic Conductivity Testing
7-1	Summary of Unknown Compounds Detected in Subsurface Soil
7-2	Summary of Analytes Detected in Method Blanks
7-3	Summary of Analytes Detected in Rinse Blanks
7-4	Summary of Analytes Detected in Trip Blanks
7-5	ADL, AREE 69W Field Analytical Results
7-6	ADL, AREE 69W Soil Off-Site Analytical Results
7-7	ADL, AREE 69W Groundwater Off-Site Analytical Results

W003971.080

LIST OF TABLES

(continued)

-		1. T	
1	2	hv.	6
ча.	a	LA.	10

Title

- 7-8 ADL, AREE 69W Surface Water Off-Site Analytical Results
- 7-9 ADL, AREE 69W Sediment Off-Site Analytical Results
- 7-10 Subsurface Soil Field Analytical Results
- 7-11 RI Soil Off-Site Analytical Results
- 7-12 EPH/VPH Subsurface Soil Off-Site Analytical Results
- 7-13 Soil Removal Action Off-Site Analytical Results
- 7-14 RI Groundwater Field Analytical Results
- 7-15 RI Groundwater Off-Site Analytical Results
- 7-16 EPH/VPH Groundwater Off-Site Laboratory Results
- 7-17 RI Sediment Off-Site Analytical Results
- 7-18 RI Air Off-Site Analytical Results
- 8-1 Chemical and Physical Properties of Compounds Detected
- 8-2 Mobilities of Inorganic Elements
- 9-1 Summary of Sample Locations Used in the Human Health Risk Assessment
- 9-2 Chemicals of Potential Concern
- 9-3 Summary of Potential Pathways for Human Health Risk Assessment
- 9-4 Exposure Parameters
- 9-5 Permeability Coefficients for CPCs in Surface Water
- 9-6 Oral Dose/Response Data For Carcinogenic Effects
- 9-7 Inhalation Dose/Response Information for Carcinogenic Effects
- 9-8 Oral Dose/Response Information for Noncarcinogenic Effects
- 9-9 Inhalation Dose/Response Data for Noncarcinogenic Effects
- 9-10 Dermal Dose/Response Information for Carcinogenic and Noncarcinogenic Effects
- 9-11 Quantitative Risk Summary
- 9-12 Potential Sources of Uncertainty

W003971.080

LIST OF TABLES

(continued)

Т	a	b	le

Title

- 9-13 Endpoints for Ecological Assessment
- 9-14 Ecological Contaminants of Potential Concern in Surface Soil
- 9-15 Ecological Contaminants of Potential Concern in Upgradient Sediment
- 9-16 Ecological Contaminants of Potential Concern in Downgradient Sediment
- 9-17 Ecological Contaminants of Potential Concern in Groundwater
- 9-18 Ecological Receptors Evaluated at AOC 69W
- 9-19 Model for Estimation of Contaminant Exposures for Representative Wildlife Species
- 9-20 Estimation of Bioaccumulation Factors
- 9-21 Results of Sediment Toxicity Testing
- 9-22 Results of Food-Web Modeling for Surface Soil and Sediment
- 9-23 Summary of Ecological Risk for Plants and Invertebrates in Surface Soil
- 9-24 Comparison of Downgradient Sediment Exposure Concentrations with Toxicity Benchmark Values
- 9-25 Comparison of Groundwater Exposure Concentrations with Toxicity Benchmark Values
- 9-26 Potential Sources of Uncertainty in Ecological Risk Assessment

Harding Lawson Associates (HLA), formerly ABB Environmental Services, Inc. (ABB-ES) prepared this Remedial Investigation (RI) Report on Area of Contamination (AOC) 69W to support Task Order 001 and Modification No. 001 of Contract DAAA-31-94-D-0061 under the oversight of the U.S. Army Corps of Engineers - New England District (CENAE). This RI Report details the results of the RI and previous investigations at AOC 69W, which were completed in accordance with relevant U.S. Environmental Protection Agency (USEPA), U.S. Army Environmental Center (USAEC), and CENAE guidance.

Fort Devens was identified for cessation of operations and closure under Public Law 101-510, the Defense Base Realignment and Closure Act of 1990, and was officially closed in September 1996. Portions of the property formerly occupied by Fort Devens were retained by the Army and renamed the Devens Reserve Forces Training Area (RFTA). Areas not retained as part of the Devens RFTA were, or are in the process of being transferred to new owners for reuse and redevelopment. AOC 69W is located in an area designated for transfer to the Massachusetts Government Land Bank for reuse as a recreation area and open space. An existing school building within AOC 69W may be re-opened in the future.

AOC 69W is located on the northern portion of the former Main Post near the northeast corner of the intersection of MacArthur Avenue and Antietam Street. AOC 69W is comprised of the former Fort Devens Elementary School (Building 215) and the associated parking lot and adjacent lawn extending approximately 300 feet northwest to Willow Brook (Figure ES-1).

AOC 69W was identified for investigation based upon reports of fuel oil releases which occurred in 1972 and 1978. Each release was reported to have been between 7,000 and 8,000 gallons. In response to the 1972 fuel oil release an oil recovery system was installed consisting of a subsurface pipe leading from the school foundation to a 250 gallon UST located adjacent to Willow Brook (Figure ES-1).

RI ACTIVITIES

In general, the efforts associated with this RI have defined two areas of soil contamination at AOC 69W. The larger of the two areas was attributed to the 1972

Harding Lawson Associates

W003971.080

EXECUTIVE SUMMARY

fuel oil release and extended from the new boiler room to approximately 300 feet northwest. Detected contaminants included primarily TPHC and PAHs at approximately 6 to 10 feet below ground surface (bgs) adjacent to the school and 0 to 4 feet bgs downgradient near the wooded area. Contaminant distributions indicated that the underground recovery system piping had acted as a conduit for contaminant migration. Based on a review of the soil and groundwater contaminant data, a removal action was undertaken in the winter of 1997 and 1998 to remove contaminated subsurface soil in the vicinity of the 1972 fuel oil spill and the associated recovery system. Confirmatory sample results indicate that concentrations of PAHs, VPH, and EPH still exceed MCP S-1/GW-1 standards immediately adjacent to the school, but that concentrations of fuel-related contaminants in other portions of the site are generally low. The soil removal action has removed the majority of the source area contamination.

The other identified area of soil contamination is attributed to the 1978 fuel oil release and is located adjacent to the school building outside the old boiler room and beneath the paved parking lot. Contaminants are primarily TPHC at depths of 4 to 7 feet bgs. The observed contamination does not appear to be migrating downgradient based upon analytical sample results from numerous downgradient TerraProbeSM explorations and monitoring wells. In addition, the area of the release is covered by the school building and the paved parking lot which inhibit groundwater recharge. No further migration is expected based upon the age of the spill.

Fuel related VOCs, SVOCs, TPHC, and inorganics comprise the majority of the observed groundwater contaminants at AOC 69W. Groundwater contamination identified by field and off-site analytical data was coincident with the observed soil contamination and fuel recovery system. Soil and groundwater contaminants appear to be predominantly located in the vicinity of the water table.

Review of contaminant distribution in sediment samples from Willow Brook suggests that the petroleum contamination may have contributed SVOCs to Willow Brook sediments as deep as 2 feet bgs. However, the stream bed is lined with chunks of asphalt which may be acting as an alternative source of SVOCs. Pesticides, PCBs, and TPHCs were detected primarily in upgradient samples suggesting that their source is not site related.

Harding Lawson Associates

W003971.080

Human Health Risk Assessment

Possible health risks were evaluated for the current and anticipated future land uses at AOC 69W. Since the former elementary school at AOC 69W is presently closed, exposures and risks for current site use were evaluated for a site maintenance worker (possible exposure to surface soil), and a trespasser ages 6 through 18 (possible exposure to sediment and groundwater discharge to surface water). The possible health risks associated with future site use were evaluated assuming that the school will be re-opened, and included evaluation of a pupil ages 6 through 18 (possible exposure to surface soil, sediment, groundwater discharge to surface water, and indoor air), and an excavation worker (possible exposure to surface soil and subsurface soil). In addition, future use of the groundwater as a potable water source was evaluated. Since groundwater at and beneath AOC 69W is not presently used as a source of drinking or industrial water, evaluation of potable use represents a *hypothetical worst-case* evaluation of potential exposures and risks.

The risk assessment evaluated post-removal action conditions for surface soil and subsurface soil, and pre-removal action conditions for groundwater, sediment, and indoor air. Chemicals of potential concern (CPCs) were identified in surface soil, subsurface soil, sediment, groundwater, and indoor air, and included metals, semivolatile organic compounds (SVOCs), volatile organic compounds (VOCs), and petroleum-related compounds including total petroleum hydrocarbons (TPHC), extractable petroleum hydrocarbons (EPH), volatile petroleum hydrocarbons (VPH), and polynuclear aromatic hydrocarbons (PAH). Among these chemicals of potential concern, only the petroleum-related compounds are directly associated with the release of fuel oil at AOC 69W. The presence of the constituents detected in the building air is likely due to the presence of numerous ambient sources within and outside the building, and not due to fuel-related constituents that may be present in soil and groundwater beneath the building. Nevertheless, circumstantial evidence indicated that a possible association between low level detections of xylenes, ethylbenzene, and methylbeptane detected in air samples and soil samples collected in the vicinity of the northwestern portion of the building could not be ruled out. However, potential exposures to the maximum detected concentrations of these three analytes, as well as all others detected in indoor air samples collected from the building rooms, were not associated with risks above USEPA threshold risk levels. This indicates that pupils and school staff members who may hypothetically be

Harding Lawson Associates

W003971,080

EXECUTIVE SUMMARY

exposed to these constituents in building air would not incur health risks considered to be of concern by the USEPA.

Possible health risks were quantified for carcinogenic and non-carcinogenic effects, for both reasonable maximum and central tendency exposure assumptions. Estimated cancer and non-cancer risks associated with possible current and future land use exposures to surface soil, subsurface soil, sediment, groundwater discharge to surface water and indoor air were within acceptable levels established the USEPA (Table ES-1). Estimated cancer and non-cancer risks associated with hypothetical exposures to AOC 69W groundwater used as a residential drinking water source exceeded levels generally considered acceptable by USEPA. However, these risks are primarily due to the presence of arsenic in the groundwater, which is not interpreted to be directly related to the release of fuel-oil at AOC 69W.

Based on the conclusions of the risk assessment, there are no unacceptable human health risks associated with surface soil, subsurface soil, groundwater discharge, sediment, or indoor air at AOC 69W. Future use of the site as a school is not associated with any unacceptable risks. Moreover, the soil removal action at AOC 69W significantly reduced fuel oil contamination in soil thereby mitigating possible exposures to fuel-related CPCs in soil, reducing the CPC concentrations in groundwater, and eliminating a possible source of fuel-related vapors. Therefore, the risk estimates presented in this risk assessment are worst-case estimates that are unlikely to be exceeded under conceivable future land use conditions.

Ecological Risk Assessment

Potential risks for ecological receptors were evaluated for chemicals detected in surface soil, sediment, and groundwater at AOC 69W. CPCs that were identified in these media included metals, pesticides, PCBs, SVOCs, VOCs, and petroleum-related compounds including TPHC, EPH/VPH, and PAHs.

Exposure point concentrations (i.e., the concentration in environmental media to which ecological receptors may be exposed) in soil reflect conditions after the soil removal action occurred, whereas exposure point concentrations in groundwater and sediment reflect conditions before the soil removal action occurred. However, the soil removal action has mitigated the reducing conditions that has likely mobilized

Harding Lawson Associates

certain metals in groundwater. Risks to ecological receptors were determined based on a conservative comparison of exposure point concentrations or exposure doses with appropriate toxicological benchmarks. If an exposure dose or concentration in site media exceeded a toxicological benchmark, then risks to ecological receptors might occur, and the uncertainties associated with those risks are discussed.

The following exposure pathways were evaluated in the ERA:

- small mammal and bird, predatory mammal, terrestrial plant, and soil invertebrate exposures to surface soil,
- small mammal and bird, predatory mammal, and aquatic receptor exposures to sediment in Willow Brook, and
- aquatic receptors exposures to groundwater that seasonally discharges to Willow Brook.

The ERA for aquatic receptors is highly conservative as Willow Brook is only seasonally inundated and is generally characterized as a degraded ditch habitat (i.e., the ditch is bordered by maintained grass, and is lined with asphalt in the upper portions).

The risk results of the AOC 69W ERA are summarized in Table ES-2. In general, there are no risks to ecological receptors except in few cases where negligible risks were estimated (i.e., exposure point concentrations or doses only slightly exceeded the toxicity benchmark). Risks to terrestrial plants may occur at one surface soil sample location (ZWS-95-42X) due to the presence of lead. However, the presence of lead at this location may be associated more with road run-off or lawn mower maintenance than from the fuel oil release, which occurred approximately 300 feet to the southeast. Risks to plants would be localized, and are not likely to result in population-level effects.

Risks to aquatic organisms were also identified for certain metals; however, the soil removal action has likely mitigated the reducing conditions in the subsurface soils that may have mobilized the metals in groundwater. Adverse effects were observed for aquatic organisms exposed to sediment in toxicity tests; however, these adverse

Harding Lawson Associates

effects are likely related to the poor habitat and substrate quality, rather than the presence of site-related chemicals. This is supported by the fact that exposure point concentrations for chemicals detected in sediment only slightly exceeded sediment benchmarks.

Based on the conclusions of the ERA, there are no unacceptable risks associated with site-related fuel oil contamination at AOC 69W.

RECOMMENDATIONS

Based upon the results and interpretation of the RI, supplemental sampling, and the soil removal action it is recommended that AOC 69W be proposed for limited action consisting of long term monitoring of downgradient groundwater quality with no additional investigation or remedial action. This course of action is supported by the following:

- The soil removal significantly reduced surficial and subsurface contaminants that were acting as a source for groundwater contamination
- There are no unacceptable risks posed to human health or the environment from site related fuel contamination
- Estimated risks associated with hypothetical exposure to groundwater used as a residential drinking water source exceeded levels generally considered acceptable by the USEPA. However, groundwater at AOC 69W is not used as a source of drinking or industrial water. Estimated risks for groundwater as a drinking water source are primarily due to the presence of arsenic. The presence of the arsenic is attributed to the mobilization of naturally occurring arsenic by reducing conditions in the aquifer brought on by the aerobic degradation of fuel related contaminants. The soil removal will act to lessen reducing conditions in the aquifer and therefore decrease arsenic concentrations in the groundwater.
- Because the soil removal eliminated the majority of source area contaminants, estimated risks and interpretations represent worst-case estimates that are unlikely to be exceeded under future land use conditions.

Harding Lawson Associates

W003971.080

1.0 INTRODUCTION

This Remedial Investigation (RI) Report (Data Item A009) for Area of Contamination (AOC) 69W at Devens in north central Massachusetts was prepared by Harding Lawson Associates (HLA) formerly ABB Environmental Services, Inc. (ABB-ES), as a component of Task Order 001 of Contract DAAA31-94-D-001 (ABB-ES, 1996a) with the U.S. Army Corps of Engineers, New England District (CENAE), (ABB-ES, 1996b). This report details the results of the RI program at AOC 69W, which was completed in accordance with relevant USACE, U.S. Army Environmental Center (USAEC), and U.S. Environmental Protection Agency (USEPA) guidance.

1.1 PURPOSE AND SCOPE

The scope of work for the RI at AOC 69W was specified by the Army based on previous studies and investigations, and by the USEPA and Massachusetts Department of Environmental Protection (MADEP) comments on prior investigations at this AOC.

Prior investigations included a supplemental site investigation (SSI) activities completed in 1994 under the Areas Requiring Environmental Evaluation (AREE) program. The investigation phase was undertaken to establish the nature and distribution of site-related contaminants in groundwater and subsurface soil at then AREE 69W. Based on the findings of the SSI it was recommended that further investigation of groundwater and subsurface soil contamination should be conducted. In compliance with the Interagency Agreement (IAG) AREE 69W was administratively transferred to the RI/FS process and redesignated AOC 69W. The following activities were included in the RI investigation:

- Background research of historical records, personnel interviews, areal photographic interpretation, and literature search.
 - A geophysical survey to determine if any additional site-related contaminant source areas were present, and to clear intrusive investigation activities.

Harding Lawson Associates

W003971.080

- Sediment and toxicity test sampling in nearby Willow Brook.
- Sampling and field analysis from TerraProbeSM points and soil borings to further define the horizontal and vertical distribution of soil and groundwater contamination.
- Installation of groundwater monitoring wells, and groundwater sampling from new and existing monitoring wells for off-site laboratory analyses.
- Installation of piezometers and a surface water measurement station in nearby Willow Brook.
- Aquifer testing.
- Ecological survey and wetland investigations.
- Vertical and horizontal location surveys.

Based upon the RI data, supplemental sampling was conducted in 1996 to address identified data gaps. The supplemental sampling consisted of:

- Air quality sampling within Building 215 (the Elementary School) to assess impacts to indoor air from subsurface soil contamination;
- Installation of soil borings and collection of soil samples within both of the boiler rooms and courtyard to identify potential contaminant source areas;
- Installation of monitoring wells;
- Two rounds of groundwater sampling for off-site analysis;
- Aquifer testing;
- Vertical and horizontal location surveys; and

Harding Lawson Associates

W003971.080

Additional air sampling and groundwater sampling in 1997.

1.2 REPORT ORGANIZATION

Preparation of this RI Report consisted of characterizing the geologic and hydrogeologic conditions at the site, and assessing the distribution, migration, and potential effects of identified chemicals on human and ecological receptors. The content and presentation of this report relies heavily upon figures and tables which present the data in the context of exploration locations on site maps. The text within the report supports the figures, and provides detail, interpretation, and analysis that cannot be presented in figures.

After acquiring and evaluating the field and off-site laboratory data and identifying chemical-specific Applicable or Relevant and Appropriate Requirements (ARARs), HLA prepared this RI Report for AOC 69W in accordance with USEPA and USAEC guidance. The report describes the field methods employed, and presents, summarizes, and evaluates the relevant background information, field and laboratory data, results and conclusions from previous investigations, and assesses the potential human health and ecological risks.

Section 2.0 of this report describes the history and physical setting of the Devens area. Section 3.0 summarizes the RI analytical program, including the field procedures, QA and QC, and data management. Section 4.0 presents potential ARARs and background concentrations of inorganic analytes in soil and groundwater. Section 5.0 of this report summarizes the AOC 69W background and physical conditions, previous investigations, technical objectives of the RI, and RI sampling and investigatory techniques. Section 6.0 presents the interpretation of geologic and hydrogeologic conditions at AOC 69W. Section 7.0 presents the nature and distribution of detected site contaminants, Section 8.0 outlines the fate and transport of the detected site contaminants, Section 9.0 presents the human health and ecological baseline risk assessment, and Section 10.0 presents the conclusions and recommendations for AOC 69W.

In accordance with the Federal Facilities Agreement (FFA), this RI report will be presented in a Draft version, and after regulatory review, a Final version.

Harding Lawson Associates

1.3 PROJECT OBJECTIVES

The objective of the project at AOC 69W was to perform an RI in accordance with all relevant MADEP and USEPA guidance and in compliance with applicable Army approved field methods and procedures. The purpose of the RI conducted at AOC 69W was to further define the site contaminants detected in the soil and groundwater during the AREE 69W SSI, and to determine whether remediation is warranted.

1.4 PROJECT APPROACH

To meet the project objectives, a significant amount of effort was focused on the production of several RI planning documents. The planning documents were developed in compliance with the appropriate regulatory guidance for remedial investigations, regulatory and Army comments, and results of previous investigations.

The project plans were designed to answer data gaps identified from the previous investigation and gather additional data on the physical conditions of the AOC, the nature and distribution of site-related contaminants, and assess the risks to human and ecological receptors.

1.4.1 Project Operations Plan

The principal planning document was the Fort Devens Project Operations Plan (POP) (ABB-ES, 1995a), which provides detailed descriptions and discussions of the elements essential to conducting field investigation activities. The purpose of this plan was to define responsibilities and authorities for data quality, and to define requirements such that the field investigation activities undertaken by HLA at Devens would be planned and executed in a manner consistent with Army quality assurance (QA) program objectives. The Fort Devens POP (ABB-ES, 1995a) includes the specified elements of a Sampling and Analysis Plan (SAP) and Health and Safety Plan (HASP). The Fort Devens POP (ABB-ES, 1995a) also includes the essential elements of the Quality Assurance Project Plan (QAPP) and the Field Sampling Plan. USEPA has prepared guidance on the preparation of a Fort Devens POP (ABB-ES, 1995a) in "Guidance for Preparation of Combined Work/Quality Assurance Project Plans for Environmental Monitoring"; (USEPA, 1984). The guidance was designed to eliminate the necessity for preparation of multiple,

Harding Lawson Associates

W003971.080

redundant documents.

The requirements of the Fort Devens POP (ABB-ES, 1995a) were applied to HLA and subcontractor activities related to the collection of environmental data at Devens. The Fort Devens POP (ABB-ES, 1995a) adheres to the requirements and guidelines contained in the "USAEC QA Program, January 1990" for collection and analysis of samples and the USAEC "Geotechnical Requirements for Drilling, Monitoring Wells, Data Acquisition, and Reports, March 1987" for the installation of test pits, soil borings and monitoring wells, collection of soil, surface water, sediment, and groundwater samples, and for land survey location. In addition, the Fort Devens POP (ABB-ES, 1995a) meets guidelines of USAEC chain-of-custody (COC) procedures.

The Fort Devens POP (ABB-ES, 1995a) provides guidance and specifications to ensure that samples are obtained under controlled conditions using appropriate, documented procedures; and that samples are identified uniquely and controlled through sample tracking systems and COC protocols. The Fort Devens POP (ABB-ES, 1995a) also includes specifications to ensure that field determinations and laboratory analytical results are of known quality and are valid, consistent, and compatible with the USAEC chemical data base through the use of certified methods, preventive maintenance, calibration, and analytical protocols, quality control (QC) measurements, review, correction of out-of-control situations, and audits. The Fort Devens POP (ABB-ES, 1995a) also specifies the methods and procedures to be used to ensure that calculations and evaluations are accurate, appropriate, and consistent throughout the projects; generated data are validated and their use in calculations is documented; and records are retained as documentary evidence of the quality of samples, applied processes, equipment, and results.

The HASP was prepared as an integral element of the Fort Devens POP (ABB-ES, 1995a) in accordance with the same schedule and review requirements (ABB-ES, 1995b, Appendix A). The HASP complies with USAEC's EM 385-1-1, AMC-R-385-100, and Devens RFTA safety requirements, as well as Occupational Safety and Health Administration (OSHA) Regulations 29 CFR 1910.120. The HASP development was based on appropriate information contained in previous investigation documents from Devens. The HASP portion of the Fort Devens POP (ABB-ES, 1995a) ensures that health and safety procedures are maintained by requiring inclusion of the health and safety staff function in the project organization.

Harding Lawson Associates

1.4.2 Task Order Work Plan

The background, rationale, and specific scope for the RI are set forth in companion planning documents, the Task Order Work Plan, Task Order Work Plan Addendum, and specific sampling approach documents. The Revised Final Task Order Work Plan for AOC 69W was prepared under Contract DAAA31-94-D-0061 Task Order No. 001 (ABB-ES, 1996a) and the Work Plan Addendum was prepared under Modification No. 1 of the aforementioned contract (ABB-ES, 1996b). The Task Order Work Plans were developed to comply with the Massachusetts Contingency Plan (MCP) (310 Code of Massachusetts Regulations [CMR] 40.000); the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA), as amended by the Superfund Amendments and Reauthorization Act (SARA) of 1986; the corrective action provisions of the Hazardous and Solid Waste Amendments; and the Toxic Substances Control Act. Work conducted under the Task Order Work Plans was performed in accordance with the provisions of the Federal Facility Agreement (FFA) (USEPA and U.S. Army, 1991) and Army guidelines.

The background information provided in the Revised Final Task Order Work Plan for AOC 69W was based largely on information in the Master Environmental Plan (MEP), review of installation documents, observations made during site visits conducted by HLA, interviews with installation personnel, and the AREE 69W SSI. Summaries of each of these activities and discussions of specific field activities to be conducted under Task Order 001 were included in the Revised Final Task Order Work Plans. The discussions focused specifically on the objectives and scope of proposed RI activities.

Harding Lawson Associates

W003971.080

2.0 INSTALLATION DESCRIPTION

Devens is located in the towns of Ayer and Shirley (Middlesex County) and Harvard and Lancaster (Worcester County), approximately 35 miles northwest of Boston, Massachusetts. It lies within the Ayer, Shirley, and Clinton map quadrangles (7^{1/2}-minute series). Devens and the RFTA occupy approximately 9,260 acres and are divided into the Former North Post, the Main Post, and South Post (Figure 2-1).

Fort Devens was identified for cessation of operations and closure under Public Law 101-510, the Defense Base Realignment and Closure Act of 1990, and was officially closed in September 1996. Portions of the property formerly occupied by Fort Devens were retained by the Army for reserve forces training and renamed the Devens Reserve Forces Training Area (RFTA). Areas not retained as part of the Devens RFTA were, or are in the process of being, transferred to new owners for reuse and redevelopment. AOC 69W is located in an area planned for transfer to the Massachusetts Government Land Bank for recreation/open space reuse.

Over 6,000 acres at Fort Devens were used for training and military maneuvers, and over 3,000 acres were developed for housing, buildings, and other facilities; the installation has been reported as the largest undeveloped land holding under a single owner in north-central Massachusetts (United States Fish and Wildlife Service [USFWS], 1992).

The South Post is located south of Massachusetts Route 2 and is largely undeveloped. The Main Post and North Post primarily contain developed lands, including recreational areas (e.g., a golf course and Mirror Lake), training areas, and an airfield. AOC 69W is located on the Main Post (Figure 2-2).

The following subsections describe the history and physical setting of Devens.

2.1 HISTORY

Camp Devens was created as a temporary cantonment in 1917 for training soldiers from the New England area. It was named after Charles Devens -- a Massachusetts Brevet Major General in the Union Army during the Civil War who later became

Harding Lawson Associates

2-1

W003971.080

SECTION 2

Attorney General under President Rutherford Hayes. Camp Devens served as a reception center for selectees, as a training facility, and, at the end of World War I, as a demobilization center (Marcoa Publishing Inc., 1990). At Camp Devens the 1918 outbreak of Spanish influenza infected 14,000 people, killed 800, and caused the installation to be quarantined (McMaster et al., 1982). Peak military strength during World War I was 38,000. After World War II, Camp Devens became an installation of the U.S. Army Field Forces, CONARC in 1962, and the U.S. Army Forces Command in 1973 (Biang et al., 1992).

In 1921, Camp Devens was placed in caretaker status. During summers from 1922 to 1931, it was used as a training camp for National Guard troops, Reserve units, Reserve Officer Training Corps cadets, and the Civilian Military Training Corps. In 1929, Dr. Robert Goddard used Fort Devens to test his early liquid-fuel rockets, and there is a monument to him on Sheridan Road near Jackson Gate (Fort Devens Dispatch, 1992).

In 1931, troops were again garrisoned at Camp Devens. It was declared a permanent installation, and in 1932 was formally dedicated as Fort Devens. During the 1930s, there was a limited building program, and beautification projects were conducted by the Works Progress Administration (WPA) and Civilian Conservation Corps.

In 1940, Fort Devens became a reception center for New England draftees. It expanded to more than 10,000 acres. Approximately 1,200 wooden buildings were constructed, and two 1,200-bed hospitals were built. In 1941, the Army Airfield was constructed by the WPA in a period of 113 days (Fort Devens Dispatch, 1992). In 1942, the Whittemore Service Command Base Shop for motor vehicle repair (Building 3713) was built, and at the time it was known as the largest garage in the world (U.S. Army, 1979). The installation's current wastewater treatment plant was also constructed in 1942 (Biang et al., 1992).

During World War II, more than 614,000 inductees were processed. Fort Devens' population reached a peak of 65,000. Three Army divisions and the Fourth Women's Army Corps trained at Fort Devens, and it was the location of the Army's Chaplain School, the Cook & Baker School, and a basic training center for Army nurses. A prisoner of war camp for 5,000 German and Italian soldiers was operated from 1944 to 1946. At the end of the war, Fort Devens again became a demobilization center, and in 1946 it reverted to caretaker status.

Harding Lawson Associates

W003971.080

Fort Devens was reactivated in July 1948 and again became a reception center during the Korean Conflict. It has been an active Army facility since that time.

Most recently, the mission at Fort Devens was to command and train its assigned duty units, operate the South Boston Support Activity in Boston, Massachusetts, the Sudbury Training Annex, and the Hingham U.S. Army Reserve (USAR) Annex, and support the 10th Special Forces Group (A), the U.S. Army Intelligence School, Fort Devens, the U.S. Army Reserves, Massachusetts Army National Guard, and Reserve Officer Training Programs. No major industrial operations occur at Fort Devens, although several small-scale industrial operations were performed under the Directorate of Plans, Training, and Security; the Directorate of Logistics; and the Directorate of Engineering and Housing. The major waste-producing operations performed by these groups was photographic processing and maintenance of vehicles, aircraft, and small engines. Past artillery fire, mortar fire, and waste explosive disposal at Fort Devens are potential sources for explosives contamination (USAEC, 1993).

2.2 PHYSICAL SETTING

The climate, vegetation, ecology, physiography, soils, surficial and bedrock geology, and regional hydrogeology of Devens are described in the subsections that follow.

2.2.1 Climate

The climate of Devens is typical of the northeastern United States, with long cold winters and short hot summers. Climatological data were reported for Fort Devens by U.S. Department of the Army (1979), based in part on a 16-year record from Moore Army Airfield (MAAF).

The mean daily minimum temperature in the coldest months (January and February) is 17 degrees Fahrenheit (°F), and the mean daily maximum temperature in the hottest month (July) is 83°F. The average annual temperature is 58°F. There are normally 12 days per year when the temperature reaches or exceeds 90°F and 134 days when it falls to or below freezing.

Harding Lawson Associates

W003971.080

The average annual rainfall is 39 inches. Mean monthly precipitation varies from a low of 2.3 inches (in June) to a high of 5.5 inches (in September). The average annual snowfall is 65 inches, and snowfall has been recorded in the months of September through May (falling most heavily from December through March).

Wind speed averages 5 miles per hour (mph), ranging from the highest monthly average of 7 mph (March-April) to the lowest monthly average of 4 mph (September).

Average daytime relative humidities range from 71 percent (January) to 91 percent (August), and average nighttime relative humidities range from 46 percent (April) to 60 percent (January).

2.2.2 Vegetation

The Main and North Posts at Devens are primarily characterized by urban and developed cover types. Approximately 56 percent of these areas are covered by developed lands, the golf course, the airfield, and the wastewater infiltration beds. Early successional forest cover types (primarily black cherry-aspen hardwoods) cover approximately 2 percent of the area, mixed oak-red maple hardwoods approximately 20 percent, and white pine-hardwood mixes approximately 11 percent. The rest of the North and Main Posts are characterized by various coniferous species, shrub habitat, and herbaceous cover types.

Much of the South Post is undeveloped forested land. The area includes approximately 8 percent early successional forest (black cherry, red birch, grey birch, quaking aspen, red maple); 26 percent mixed oak hardwoods; and 9 percent coniferous forest (white pine, pitch pine, red pine). Four percent of the area comprises a mixed shrub community. The 200-acre Turner Drop Zone is maintained as a grassland that represents a "prairie" habitat. Vegetative cover in the large "impact area" of the central South Post has not been mapped in detail. It is dominated by fire-tolerant species such as pitch pine and scrub oak.

Extensive sandy glaciofluvial soils are found in the Nashua River Valley, particularly in the South and North Post areas of Devens. Extensive accumulations of these soils are unusual in Massachusetts outside of Cape Cod and adjacent areas of southeastern

Harding Lawson Associates

W003971.080

Massachusetts, and they account for some of the floral and faunal diversity at the installation.

2.2.3 Ecology

The former Fort Devens encompasses numerous terrestrial, wetland, and aquatic habitats in various successional stages. Floral and faunal diversity is strengthened by the installation's close proximity to the Nashua River; the amount, distribution, and nature of wetlands; and the undeveloped state and size of the South Post (USFWS, 1992). Much of Fort Devens was formerly agricultural land and included pastures, woodlots, orchards, and cropped fields. Existing habitat types reflect this agrarian history, ranging from abandoned agricultural land to secondary growth forested regions. Devens is generally reverting back to a forested state.

There are 1,313 acres of wetlands at Devens. The wetlands are primarily palustrine, although riverine and lacustrine types are also found. Forested palustrine floodplain wetlands associated with the Nashua River and its tributary Nonacoicus Brook are located on Devens' Main and North Posts. These include 191 acres of flooded areas, emergent marsh, and shrub wetlands. Also present are 245 acres of isolated regions of palustrine wetlands and lacustrine systems. On the South Post, there are 877 acres of wetlands, consisting of deciduous forested wetlands, deciduous shrub swamps, emergent marsh, open lacustrine waters in ponds, and open riverine waters.

Approximately half of Devens' land area abuts the northern boundary of the Oxbow National Wildlife Refuge (NWR), a federal resource administered as part of the Great Meadows NWR (USFWS, 1992).

Devens supports an abundance and diversity of wildlife. Identified taxa include 771 vascular plant species, 538 species of butterflies and moths, eight tiger beetle species, 30 vernal pool invertebrates, 15 amphibian species (six salamanders, two toads, seven frogs), 19 reptile species (seven turtles, 12 snakes), 152 bird species, and 42 mammal species. The status of fish populations in Devens aquatic systems has not been fully defined.

Rare and endangered species at Devens include the federally listed (threatened) bald eagle and peregrine falcon (both occasional transients); the state-listed (endangered) upland sandpiper, ovoid spike rush, and Houghton's flatsedge; the state-listed

Harding Lawson Associates

2-5

(threatened) Blanding's turtle, cattail sedge, pied-billed grebe, and northern harrier; and the state-listed (special concern) blue-spotted salamander, grasshopper sparrow, spotted turtle, wood turtle, water shrew, blackpoll warbler, American bittern, Cooper's hawk, sharp-shinned hawk, and Mystic Valley amphipod. Also state-listed as rare or endangered are three Lepidoptera (butterfly and moth) species identified at Devens.

The Massachusetts Natural Heritage Program has developed Watch Lists of unprotected species that are uncommon or rare in Massachusetts. From the Watch Lists, 14 plant species, two amphibian species, and 15 bird species have been observed at Devens.

2.2.4 Physiography

Devens is in a transitional area between the coastal lowland and central upland regions of Massachusetts. All of the landforms are products of glacial erosion and deposition on a crystalline bedrock terrain. Glacial erosion was superimposed on ancient bedrock landforms that were developed by the erosional action of preglacial streams. Generally, what were bedrock hills and ridges before the onset of Pleistocene glaciation were only moderately modified by glacial action, and they remain bedrock hills and ridges today. Similarly, preglacial bedrock valleys are still bedrock valleys. In post-glacial time, streams have locally modified the surficial glacial landforms but generally have not affected bedrock.

The predominant physiographic (and hydrologic) feature in the Devens area is the Nashua River (see Figure 2-1). It forms the eastern installation boundary on the South Post, where its valley varies from a relatively narrow channel (at Still River Gate), to an extensive floodplain with a meandering river course and numerous cutoff meanders (at Oxbow National Wildlife Sanctuary). The Nashua River forms the western boundary of much of the Main Post, and there its valley is deep and comparatively steep-sided with extensive bedrock outcroppings on the eastern bank. The river flows through the North Post in a well-defined channel within a broad forested floodplain.

Terrain at Devens falls generally into three types. The least common is bedrock terrain, where rocks that have been resistant to both glacial and fluvial erosion remain as topographic highs, sometimes thinly veneered by glacial deposits.

Harding Lawson Associates

W003971.080

Shepley's Hill on the Main Post is the most prominent example.

A similar but more common terrain at Devens consists of materials (tills) deposited directly by glaciers as they advanced through the area or as the ice masses wasted (melted). These landforms often conform to the shape of the underlying bedrock surface. They range from areas of comparatively low topographic relief (such as near Lake George Street on the Main Post) to elongated hills (drumlins) whose orientations reflect the direction of glacier movement (such as Whittemore Hill on the South Post).

The third type of terrain was formed by sediment accumulations in glacial-meltwater streams and lakes (glaciofluvial and glaciolacustrine deposits). This is the most common terrain, comprising most of the North and South Posts and much of the Main Post. Its form bears little or no relationship to the shape of the underlying bedrock surface. Landforms include extensive flat uplands such as the hills on which MAAF and the wastewater infiltration beds are located on the North Post. Those are large remnants of what was once a continuous surface that was later incised and divided by downcutting of the Nashua River. Another prominent glacial meltwater feature is the area around Cranberry Pond and H-Range on the South Post. This is classic kame-and-kettle topography formed by sand and gravel deposition against and over large isolated ice blocks, followed by melting of the ice and collapse of the sediments. The consistent elevations of the tops of these ice-contact deposits are an indication of the glacial-lake stage with which they are associated. Mirror Lake and Little Mirror Lake on the Main Post occupy another conspicuous kettle.

2.2.5 Soils

Devens and the RFTA lies within Worcester County and Middlesex County in Massachusetts (see Figure 2-1). The soils of Worcester County have been mapped by the Soil Conservation Service (SCS) of the U.S. Department of Agriculture (USDA) (SCS, 1985). Mapping of the soils of Middlesex County has not been completed. However, an interim report (SCS, 1991), field sheet #19 (SCS, 1989), and an unpublished general soil map (SCS, undated) are available.

Soil mapping units ("soil series") that occur together in intricate characteristic patterns in given geographic areas are grouped into soil "associations." Soils in the Worcester County portions of Devens consist generally of three associations. Three

Harding Lawson Associates

W003971.080

associations also have been mapped in the Middlesex County portions of Devens. Although the mapped associations are not entirely the same on both sides of the county line, the differences reflect differences in definition and the interim status of Middlesex County mapping. The general distributions of the soil associations are shown in Figure 2-3, and descriptions of the soil series in those associations are provided below.

WORCESTER COUNTY (SCS, 1985)

Winooski-Limerick-Saco Association:

<u>Winooski Series</u>. Very deep; moderately well-drained; slopes 0 to 3 percent; occurs on floodplains; forms in silty alluvium.

<u>Limerick Series</u>. Very deep; poorly drained; slopes 0 to 3 percent; occurs on floodplains; forms in silty alluvium.

<u>Saco Series</u>. Very deep; very poorly drained; slopes 0 to 3 percent; occurs on floodplains; derived mainly from schist and gneiss.

Hinckley-Merrimac-Windsor Association:

<u>Hinckley Series</u>. Very deep; excessively drained; slopes 0 to 35 percent; occurs on stream terraces, eskers, kames, and outwash plains.

<u>Merrimac Series</u>. Very deep; excessively drained; slopes 0 to 25 percent; occurs on stream terraces, eskers, kames, and outwash plains.

Windsor Series. Very deep; moderately well-drained; slopes 0 to 3 percent; occurs on floodplains.

Paxton-Woodbridge-Canton Association:

<u>Paxton Series</u>. Very deep; well-drained; slopes 3 to 35 percent; occurs on glacial till uplands; formed in friable till overlying firm till.

Woodbridge Series. Very deep; moderately well-drained; slopes 0 to

Harding Lawson Associates

15 percent; occurs on glacial till uplands; formed in firm till.

<u>Canton Series</u>. Very deep; well-drained; slopes 3 to 35 percent; occurs on glaciated uplands; formed in friable till derived mainly from gneiss and schist.

MIDDLESEX COUNTY (SCS, 1991)

<u>Hinckley-Freetown-Windsor Association</u>: The soils at AOC 69W are comprised of this soil type (See Figure 2-3). (This is a continuation of the Hinckley-Merrimac-Windsor Association mapped in Worcester County):

<u>Hinckley Series</u>. Deep; excessively drained; nearly level to very steep; occurs on glacial outwash terraces, kames, and eskers; formed in gravelly and cobbley coarse textured glacial outwash.

<u>Freetown Series</u>. Deep; very poorly drained; nearly level, organic; occurs in depressions and on flat areas of uplands and glacial outwash plains.

<u>Windsor Series</u>. Deep; excessively drained; nearly level to very steep; occurs on glacial outwash plains, terraces, deltas, and escarpments; formed in sandy glacial outwash.

Quonset-Carver Association:

<u>Quonset Series</u>. Deep; excessively drained; nearly level to very steep; occurs on glacial outwash plains, terraces, eskers, and kames; formed in water-sorted sands derived principally from dark phyllite, shale, or slate.

<u>Carver Series</u>. Deep; excessively drained; nearly level to steep; occurs on glacial outwash plains, terraces, and deltas; formed in coarse, sandy, water-sorted material.

<u>Winooski-Limerick-Saco Association</u>: (This is a continuation of the same association mapped along the Nashua River floodplain in Worcester County).

Harding Lawson Associates

2.2.6 Surficial Geology

Devens and the RFTA lie in three topographic quadrangles: Ayer, Clinton, and Shirley. The surficial geology of Devens has been mapped only in the Ayer quadrangle (Jahns, 1953) and Clinton quadrangle (Koteff, 1966); the Shirley quadrangle is unmapped.

Unconsolidated surficial deposits of glacial and postglacial origin comprise nearly all of the exposed geologic materials at Devens. The glacial units consist of till, deltaic deposits of glacial Lake Nashua, and deposits of glacial meltwater streams.

The superficial geology at AOC 69W can be placed in the following geologic setting. The till ranges from unstratified gravel to silt, and it is characteristically bouldery. Jahns (1953) and Koteff (1966) recognize a deeper unit of dense, subglacial till, and an upper, looser material that is probably a slightly younger till of englacial or superglacial origin. Till is exposed in ground-moraine areas of the Main Post (such as in the area of Lake George Street) and on the South Post at and south of Whittemore Hill. It also underlies some of the water-laid deposits (Jahns, 1953). Till averages approximately 10 feet in thickness but reaches 60 feet in drumlin areas (Koteff, 1966).

Most of the surficial glacial units in the Nashua Valley are associated with deposition in glacial Lake Nashua, which formed against the terminus of the Wisconsinan ice sheet as it retreated northward along the valley. Successively lower outlets were uncovered by the retreating glacier, and the lake level was correspondingly lowered. Koteff (1966) and Jahns (1953) recognize six lake levels (stages) in the Devens area, distinguished generally by the elevations and distribution of their associated deposits. The stages are, in order of development: Clinton Stage; Pin Hill Stage; Old Mill Stage; Harvard Stage; Aver Stage; and Groton Stage.

The glacial lake deposits consist chiefly of sand and gravelly sand. Coarser materials are found in topset beds of deltas built out into the lakes and in glacial stream beds graded to the lakes. Delta foreset beds are typically composed of medium to fine sand, silt, and clay. Lake-bottom deposits, which consist of fine sand, silt, and clay, are mostly covered by delta deposits and are seldom observed in glacial Lake Nashua deposits. One of the few known exposures of glacial lake-bottom sediments in the region is on the South Post near A- and C-Ranges. There, a section of more than

Harding Lawson Associates

14 feet of laminated clay was mined for brick-making in the early part of this century (Alden, 1925, pp. 70-71). The general physical characteristics of glacial lake deposits are the same regardless of the particular lake stage in which the deposits accumulated (Koteff, 1966; Jahns, 1953). Although glaciofluvial and glaciolacustrine sediments are typically well stratified, correlations between borings are difficult because of laterally abrupt changes characteristic of these generally high-energy depositional environments.

Postglacial deposits consist mostly of river-terrace sands and gravels; fine alluvial sands and silts beneath modern floodplains; and muck, peat, silt, and sand in swampy areas.

Jahns (1953) also observed a widespread veneer of windblown sand and ventifacts above the glacial materials (and probably derived from them in the brief interval between lake drainage and the establishment of vegetative cover).

2.2.7 Bedrock Geology

Devens and the RFTA is underlain by low-grade metasedimentary rocks, gneisses, and granites. The rocks range in age from Late Ordovician to Early Devonian (approximately 450 million to 370 million years old). Devens is situated approximately 2 miles west of the Clinton-Newbury-Bloody Bluff fault zone, that developed when the ancestral European continental plate collided with and underthrust the ancestral North American plate. The continents reseparated in the Mesozoic to form the modern Atlantic Ocean. Devens is located on the very eastern edge of the ancestral North American continental plate. A piece of the ancestral European continental plate to the very eastern edge of the ancestral North American continental plate.

Preliminary bedrock maps (at scale 2,000 feet/inch) are available for the Clinton quadrangle (Peck, 1975 and 1976) and Shirley quadrangle (Russell and Allmendinger, 1975; Robinson, 1978). Bedrock information for the Ayer quadrangle is from the Massachusetts state bedrock map (at a regional scale of 4 miles/inch) (Zen, 1983) and in associated references (Robinson and Goldsmith, 1991; Wones and Goldsmith, 1991). Among these sources, there is some disagreement about unit names and stratigraphic sequence; however, there is general agreement about the distribution of rock types.

Harding Lawson Associates

In contrast to the high metamorphic grade and highly sheared rocks of the Clinton-Newbury zone, the rocks in the Devens area are low grade metamorphics (generally below the biotite isograd) and typically exhibit less brittle deformation. Major faults have been mapped, however, including the Wekepeke fault exposed west of Devens (in an outcrop 0.25 mile west of the old Howard Johnson rest stop on Route 2).

Figure 2-4 is a generalized summary of the bedrock geology of Devens. It is compiled from Peck (1975), Robinson (1978), Russell and Allmendinger (1975), and Zen (1983), and it adopts the nomenclature of Zen (1983). Because of limited bedrock exposures, the locations of mapped contacts are considered approximate, and the mapped faults are inferred. Rock units strike generally northward to northeastward but vary locally. The bedrock units underlying Devens and the RFTA are as follows:

- DSw WORCESTER FORMATION (Lower Devonian and Silurian) Carbonaceous slate and phyllite, with minor metagraywacke to the west (Zen, 1983; Peck, 1975). Bedding is typically obscure due to a lack of compositional differences. It is relatively resistant to erosion and forms locally prominent outcrops. The abandoned Shaker slate quarry on the South Post is in rocks of the Worcester Formation. The unit corresponds to the "DSgs" and "DSs" units of Peck (1975) and the "e3" unit of Russell and Allmendinger (1975).
- So OAKDALE FORMATION (Silurian) Metasiltstone and phyllite. It is finegrained and consists of quartz and minor feldspar and ankerite, and it is commonly deformed by kink banding (Zen, 1983; Peck, 1975; Russell and Allmendinger, 1975). In outcrop it has alternating layers of brown siltstone and greenish phyllite. The Oakdale Formation crops out most visibly on Route 2 just east of the Jackson Gate exit. It corresponds to the "DSsp" unit of Peck (1975), the "e2" unit of Russell and Allmendinger (1975), and "ms" unit of Robinson (1978). The bedrock at AOC 69W has been identified as part of this formation.
- Sb **BERWICK FORMATION** (Silurian) Thin- to thick-bedded metamorphosed calcareous metasiltstone, biotitic metasiltstone, and fine-grained metasandstone, interbedded with quartz-muscovite-garnet schist and feldspathic quartzite (Zen, 1983; Robinson and Goldsmith, 1991). In areas

Harding Lawson Associates

northwest of Fort Devens, cataclastic zones have been observed (Robinson, 1978).

Dcgr CHELMSFORD GRANITE (Lower Devonian) Light-colored and gneissic, even and medium-grained, quartz-microcline-plagioclase-muscovite-biotite, pervasive ductile deformation visible in elongate quartz grains aligned parallel to mica. It intrudes the Berwick Formation and Ayer granite (Wones and Goldsmith, 1991).

AYER GRANITE

- Sacgr Clinton facies (Lower Silurian) Coarse-grained, porphyritic, foliated biotite granite with a nonporphyritic border phase; it intrudes the Oakdale and Berwick Formations and possibly the Devens-Long Pond Facies (Zen, 1983; Wones and Goldsmith, 1991).
- SOad Devens-Long Pond facies (Upper Ordovician and Lower Silurian) Gneissic, equigranular to porphyoblastic biotite granite and granodiorite. Its contact relationship with the Clinton facies is unknown (Wones and Goldsmith, 1991). Observations of mapped exposures of this unit on Fort Devens indicate that it may not be intrusive.

Bedrock is typically unweathered to only slightly weathered at Devens. Glaciers stripped away virtually all of the preglacially weathered materials, and there has been insufficient time for chemical weathering of rocks in the comparatively brief geologic interval since glacial retreat.

2.2.8 Regional Hydrogeology

Devens is in the Nashua River drainage basin, and the Nashua River is the eventual discharge locus for all surface water and groundwater flow at the installation.

The water of the Nashua River has been assigned to Class B under Commonwealth of Massachusetts regulations. Class B surface water is "designated for the uses of protection and propagation of fish, other aquatic life and wildlife, and for primary and secondary contact recreation" (314 CMR 4.03).

Harding Lawson Associates

The principal tributaries of the north-flowing Nashua River at Devens are Nonacoicus Brook and Walker Brook on the North Post; Cold Spring Brook (which is a tributary of Nonacoicus Brook) on the Main Post; and Spectacle Brook and Ponakin Brook (tributaries of the North Nashua River), Slate Rock Brook, and New Cranberry Pond Brook on the South Post (see Figure 2-5).

There are two ponds on Devens' South Post that are called Cranberry Pond. For the purpose of the SIs, the isolated kettle pond located east of H-Range is referred to as Cranberry Pond, and the pond impounded in the 1970s 0.5-mile west of the Still River gate is referred to as New Cranberry Pond.

Glacial meltwater deposits constitute the primary aquifer at Devens. Measured hydraulic conductivities in meltwater deposits were comparatively high - typically 10^3 to 10^{-2} centimeters per second (cm/sec) or 2.8 to 28 feet per day (ft/day). In till and in clayey lake-bottom sediments, measured hydraulic conductivities were lower and ranged generally from 10^{-6} to 10^{-4} cm/sec or 2.8×10^{-3} to 0.28 ft/day. Groundwater also occurs in the underlying bedrock; however, flow is limited because the rocks have no primary porosity and water moves only in fractures and dissolution voids.

Groundwater in the surficial aquifer at Devens has been assigned to Class I under Commonwealth of Massachusetts regulations. Class I consists of groundwaters that are "found in the saturated zone of unconsolidated deposits or consolidated rock and bedrock and are designated as a source of potable water supply" (314 CMR 6.03).

The transmissivity of an aquifer is the product of its hydraulic conductivity and saturated thickness, and as such it is a good measure of groundwater availability. Figure 2-5 shows aquifer transmissivities at Devens, based on the regional work of Brackley and Hansen (1977). Transmissivities in the meltwater deposits range from 10 square feet per day (ft²/day) to more than 4,000 ft²/day. Aquifer transmissivities between 10 and 1,350 ft²/day correspond to potential well yields generally between 10 and 100 gallons per minute (gpm); transmissivities from 1,350 to 4,000 ft²/day typically yield from 100 to 300 gpm; and where transmissivities exceed 4,000 ft²/day, well yields greater than 300 gpm can be expected. (Most domestic wells in the area are drilled 100 to 200 feet into bedrock and yield less than 10 gpm. Higher yields are associated with deeper bedrock wells.)

Harding Lawson Associates

In Figure 2-5, the zones of highest transmissivity are found in areas of thick glacial meltwater deposits on the North and Main Posts, and these encompass the Sheboken, Patton, and McPherson production wells and the largely inactive Grove Pond well-field. AOC 69W is located in the area of the installation with a moderate transmissivity (1,350 to 4,000 ft^2/day) (see Figure 2-5). The zones of lowest transmissivity are associated with exposed till and bedrock and are located on the Main Post surrounding Shepley's Hill and between Jackson Gate and the parade ground and on the South Post at Whittemore Hill and isolated areas to the north and west.

A regional study of water resources in the Nashua River basin was reported by Brackley and Hansen (1977). A digital model of groundwater flow at Devens is available in a draft final report by Engineering Technologies Associates, Inc. (ETA) (1992).

According to ETA (1992), in the absence of pumping or other disturbances, groundwater recharge occurs in upland areas (e.g., the high ground on the Main Post between Queenstown, Givry, and Lake George Streets, and on the South Post the area around Whittemore Hill). The groundwater flows generally from the topographic highs to topographic lows. It discharges in wetlands, ponds, streams, and directly into the Nashua River. Groundwater discharge maintains the dry-weather flow of the rivers and streams. Figures 2-6 and 2-7, respectively, present ETA's regional overburden and bedrock groundwater flow maps (ETA, 1992).

Harding Lawson Associates

W003971.080

3.0 ANALYTICAL PROGRAM

Based on data obtained from previous investigations summarized in the Final Task Order Work Plan for AOC 57, 63AX, and 69W (ABB-ES, 1996a) and Task Work Plan Addendum for AOC 69W (ABB-ES, 1996b), an analytical program for the RI was established to identify contaminants that were potentially present at AOC 69W due to historical activities. Fuel hydrocarbons and chlorinated solvent contaminants were discovered in past investigations at AOC 69W. The purpose of the following subsection is to outline only those analytical procedures used during the RI program. Analytical results generated during the previous AREE 69W SSE will be included in the RI; however, the analytical program is not described in this document. Information on the AREE 69W SSE is referenced in Section 5.0 and Section 7.0.

The AOC 69W RI analytical program included field analysis as well as off-site laboratory analyses for a predetermined set of organic and inorganic analytes. The specific analyses implemented for these investigations are outlined in Subsection 3.1 for the on-site methods and Subsection 3.2 for the off-site analytical program. The following subsections describe the field and off-site analytical programs implemented for both the RI and Supplemental RI completed by ABB-ES at AOC 69W.

3.1 FIELD ANALYTICAL METHODS

Samples were analyzed in the field during the RI investigation conducted in 1995 and 1996 to provide real-time chemical data. Soil and groundwater samples were analyzed for selected volatile organic compounds (VOCs) and total petroleum hydrocarbons (TPHC), gasoline range organics (GRO), and diesel range organics (DRO) during the 1995 field program. During the 1996 program, samples were analyzed for VOCs and TPHC. Data were primarily used to evaluate the distribution of benzene, toluene, ethylbenzene, and xylene (BTEX), chlorinated solvents, and petroleum related contamination in groundwater and soil at AOC 69W. A discussion of field analytical procedures, data quality objectives, field documentation procedures, and quality control steps are outlined in Subsection 4.6 of the Fort Devens POP (ABB-ES, 1995a) and Appendix D. Target compounds and detection limits for onsite field analysis compounds are outlined in Table 3-1.

Harding Lawson Associates

W003971.080

TPHC analyses using a Miran Fixed Filter Infrared Spectrophotometer (IR) was the primary field method for evaluating semivolatile petroleum hydrocarbons in soil samples. This method is similar to USEPA Method 418.1. A soil microextraction sample preparation technique was developed for use in a field laboratory. This method provides qualitative data on the presence and absence, and relative concentration, of hydrocarbons. During the 1995 program DRO gas chromatography (GC)/Flame Ionization Detector (FID) analysis was also conducted on a subset of soils to provide semiquantitative data on medium molecular weight range petroleum hydrocarbons. DRO analysis was conducted for a subset of samples that exhibited hydrocarbon characteristics on the VOC analysis.

A Hewlett Packard 5890 Series II GC, in series with a Tekmar 3000 purge and trap concentrator, was used to measure concentrations of VOCs. Target analytes included BTEX, chlorinated solvents, and GRO to measure the volatile petroleum hydrocarbons. Several detectors were used in conjunction with the GC during the 1995 and 1996 field programs. Detectors included a FID, photoionization detector (PID), electron capture detector (ECD) and/or Electrolytic Conductivity Detector (ELCD).

During the 1995 program, detectors included ECD/PID/FID in series. During the 1996 program, detectors included PID/ELCD as specified in USEPA Method 8021 for VOCs.

3.1.1 Instrument Calibration

For analysis of samples for target compounds using a GC, an initial calibration was established. The initial calibration was accomplished through the analysis of three to five different concentrations of working standards. The response of the instrument to each standard was plotted versus the concentrations of standards to establish a calibration curve. The range of standards used to create the calibration curve was determined by the anticipated range of VOC contamination. Once all points were established on the calibration curve, the linearity was measured using linear regression analysis. The r^2 value, which provided a measure of this linearity, was required to be a minimum of 0.95 for all target analytes.

Prior to analysis of samples, a continuing calibration check standard was analyzed each day to ensure that the response of the instrument had not changed from the

Harding Lawson Associates

initial calibration. The concentration of the check standard was at mid-level in the calibration curve. The initial calibration remained valid if concentrations obtained for the target analytes were no greater than 30 percent different from values obtained from the initial calibration. If greater than two target compounds for multianalyte analysis for BTEX and chlorinated compounds were outside the 30 percent difference, a new initial calibration was created.

3.1.2 Sample Preparation and Analysis

Sample preparation for the total petroleum hydrocarbon procedure (IR analysis) is detailed in Subsection 4.6.2 of the Fort Devens POP (ABB-ES, 1995a). The IR analysis was used for gross hydrocarbon measurements and to indicate the presence or absence of contamination. A freon-113 extraction was used to prepare samples. Samples were analyzed by USEPA Method 418.1 (USEPA, 1983).

Sample preparation techniques for GC VOCs and GRO were adapted from protocols outlined in USEPA Method 5030 (USEPA, 1986). Soil samples were prepared for field analysis by the measurement of 5 grams into a soil sparger. For water samples, the amount used was 5 milliliters (mL). Both soil and groundwater samples were analyzed by the purge and trap concentration and GC separation. A DB-624 capillary column was installed in the gas chromatograph for the purpose of separating out the various compounds. A retention time window of +/-3 percent was used for the identification of target compounds.

For DRO analysis sample preparation, techniques were adapted from USEPA Method 3550 and Method 8015 (USEPA, 1986). The extraction procedure required the measurement of 2 grams of soil into a test tube with the addition of 2 grams of sodium sulfate and 2 mL of methylene chloride (solvent). The supernatant was then transferred to an injection vial labeled with the sample identification. The GC, equipped with an autosampler for a 2 microliters (μ L) sample volume injection, then analyzed the sample for identification and quantitation of DRO concentration. Additional solvent may have been added if a sufficient volume of supernatant was not initially achieved. If additional solvent was added, a dilution factor was incorporated during sample quantitation.

Harding Lawson Associates

W003971.080

3.1.3 Target Compound Concentrations Calculations

Target VOC concentrations were determined from comparisons of responses of compounds in samples versus responses from standards in the initial calibration curves described in Subsection 3.1.1. Soil compound concentrations were reported on a dry weight basis. Solid fraction data was used to calculate final VOC, GRO, DRO, and IR concentrations. Dilutions performed on both water and soil samples also were used to calculate final VOC, GRO, DRO, and IR concentrations. Dilution factors were calculated for any analyses where sample amounts were modified due to high concentrations of chemicals present in samples. Final sample results were calculated by dividing original unadjusted sample results by fraction of solid and multiplying results by any dilution factors.

Based on secondary data reviews conducted by HLA's Quality Assurance Officer and project chemist, possible data bias was identified in the GRO and DRO data set. The possible data bias is discussed below for GRO and DRO.

The results of the GRO analyses contain a possible positive bias which overestimated the measured concentration by approximately 20 percent of the true value. The bias was introduced during the preparation of the stock standard for the GRO analysis. The density of GRO was approximated as the density of benzene (0.88 grams per milliliter [g/mL]), however, according to information in the Installation Restoration Program Toxicology Guide (U.S. Air Force) gasoline has a density of approximately 0.73 g/mL.

The results of the DRO analyses should be considered estimated. Possible impacts on quantitation of hydrocarbons was introduced during the set-up of the GC analytical run program. The instrument conditions used for DRO analyses caused the loss of approximately 25 percent of the light end hydrocarbons within the diesel hydrocarbon range. The primary purpose of the DRO analysis was to estimate concentrations of fuel oils or waste oils at the site. The analytical run would effectively detect the medium to heavy molecular weight fraction of oil products; however, concentrations should be considered estimated within approximately 0.5 to 2 times the reported concentration.

Harding Lawson Associates

3.1.4 Field Documentation Procedures

Instrument logbooks were completed for each instrument used during each of the field analytical programs. A log of all chromatography runs was recorded in these logbooks. The logbooks recorded the concentrations for all calibration standards used, sample run number, sample identification, date, standard preparation records, instrument maintenance records, percent solid determination data, sample volume or weight, and any additional comments or observations of the field chemist. In addition, the results from each GC run were saved into a computerized database.

At the conclusion of the RI field effort, raw data from the GC analyses and instrument logbooks were transferred for storage at HLA's Portland, Maine office. Raw data includes chromatograms, quantitation reports, and instrument and notebook records to document analyses.

3.1.5 Field Analytical Quality Control

A QC program for the field analytical results was established prior to commencement of the RI on-site laboratory analysis. This program was developed to ensure that the data generated at the field laboratory was of sufficient quality to be considered satisfactory for its intended use. QC parameters for the RI field analytical program included initial and daily calibration check standard runs, mid-level calibration check standards after every ten samples, low-level and mid-level method blanks, cleaning blanks, and field or laboratory duplicates. QC objectives for the on-site laboratory analyses are outlined in the Fort Devens POP (ABB-ES, 1995a) and Appendix D of this report. QC sample results for the on-site laboratory are assessed in Appendix D.

Method blanks were analyzed daily to document that the analytical system was free of contamination. Samples were not run if the there were any target compounds detected above the Practical Quantitation Limit (PQL) in the method blank. In addition to the low-level method blank, a mid-level method blank was run in instances where methanol extractions were necessary. One hundred μ L of methanol were added to deionized water and analyzed to ensure that it was free of contamination.

During VOC GC analyses, cleaning blanks were run at the beginning of each day to show that the analytical system was clean. They were also run after particularly

Harding Lawson Associates

W003971.080

heavily contaminated samples were run through the GC.

For VOC analyses, a surrogate was added to every sample to determine if the matrix was having an effect on the recovery of the target compounds. The surrogate used for all field investigations was 4-Bromofluorobenzene. This surrogate was used because it is chemically similar to the target compounds and responds well on the detectors selected for the field programs. Surrogate recovery goals of 30 percent to 170 percent were established for the screening program. Data usability goals for the screening samples in the field program allowed for a wider range of accuracy because providing real time Level II quality data was the goal. The surrogate recovery limits were based on the need to analyze a large number of screening samples with a 24-hour turnaround to provide Level II data quality. Samples for which the surrogate did not meet this criteria were reanalyzed and/or qualified.

Field duplicate samples were also analyzed to determine the precision of sampling and analytical techniques. Reported concentrations of target compounds for each sample and associated duplicate pair were compared by calculating the relative percent difference (RPD) of the results. RPDs were compared to criteria from USEPA (hazardous site evaluation division) Region I laboratory data validation functional guidelines for evaluating organics analyses to evaluate the precision of measurements. Duplicate results for the RI are presented in Appendix D.

In some instances, data qualifiers were used to address data quality issues associated with a particular sample. The following qualifiers were used during the RI at AOC 69W:

- J Denotes target compound concentrations that are estimated.
- E Denotes target compound concentrations that exceed the highest standard of the calibration curve.
- U Denotes sample concentrations that are less than PQLs.
- N Denotes a value that is a possible false positive due to method blank contamination.

Harding Lawson Associates

Results of the on-site sample analyses are presented in discussions of the nature and distribution of site contaminants, in Subsection 7.0 and Appendix L of this report.

3.1.6 Method Detection Limits and Data Qualifiers

Method Detection Limits (MDLs) were established during the RI for the ELCD, PID, and the FID detector target compounds. The MDL study was completed for all VOC target compounds to provide data to verify the reporting limit (RL) concentrations used for the various field programs. RL goals of 1-2 μ g/L were established based on project needs and regulatory considerations prior to each field MDLs were calculated based on procedures published in CFR program. Appendix B, Part 136, vol. 49, no. 209. The MDL study provides an estimation of the lower concentration limit of what the detectors were able to measure. The MDLs obtained during the RI field analytical program are presented in Appendix D. With the exception of 1.1-dichloroethene (1,1-DCE), MDLs calculated for all target analytes were less three times the reporting limits indicating RL goals specified for the project were achievable. During the 1995 program, 1,1-Dichloroethane was not observed at the $1 \mu g/L$ reporting limit concentration with reliability, therefore its RL was set at 5 μ g/L. The lower standard concentration used for initial calibrations was equivalent to the RL for all target compounds.

3.2 OFF-SITE LABORATORY ANALYTICAL PARAMETERS

Soil and groundwater samples collected during the 1995 and 1996 RIs from AOC 69W were analyzed at an off-site laboratory for chemical parameters on the Devens Project Analyte List (PAL). Off-site laboratory analyses for PAL organics and inorganics were considered equivalent to USEPA definitive data (USEPA, 1993). The Devens PAL and off-site laboratory methods are presented in the Fort Devens POP (ABB-ES, 1995a) and Appendix D of this report. A summary of analytical methods requested for each sample collected at AOC 69W is contained in Section 5.0, Table 5-2. Supplemental analyses were conducted in 1997 on groundwater and air samples as discussed in Subsection 3.4.

During the 1995 and 1996 investigations, off-site laboratories performing the analytical work for all Devens investigations were required to implement the 1990 U.S. Army Toxic and Hazardous Materials Agency (USATHAMA, now USAEC) QA

Harding Lawson Associates

Program (USATHAMA, 1990). All method performance demonstrations, data management, and oversight for previous USATHAMA analytical procedures were performed by the USAEC.

The off-site laboratory contracted to implement the analytical program for the RI at AOCs 69W was Environmental Science and Engineering, Inc. (ESE) of Gainesville, Florida. This laboratory was approved to complete analyses using USATHAMA and USEPA methods. Analyses were completed while implementing the 1990 USATHAMA QA Program. Specific performance demonstration and QC components of the 1990 USATHAMA QA Plan are detailed in Subsection 3.2.3 of this report.

The following subsection describes the procedures implemented to achieve the objectives of the QA program and any additional quality control processes implemented during the RI.

3.2.1 Off-Site Laboratory Certification

In accordance with the 1990 USATHAMA OA Program, laboratories were required to demonstrate competency by performance demonstration of the PAL analytical methods conducted in association with field investigations. The USAEC requires that a laboratory demonstrate proficiency in performing USAEC methods for specific analytes. Analytical methods are based on USEPA procedures (USEPA 1983; 1986). Laboratories demonstrate proficiency by submitting data from runs of precertification calibration standards. Performance samples are then sent for analysis to the laboratory by the USAEC. The true concentrations of the analytes in the performance samples are unknown by the laboratory. The data obtained from the analyses of these samples are then sent to the USAEC to determine the laboratory's precision and accuracy. Qualifications to perform USAEC methods are awarded to laboratories based on this performance. Certified Reporting Limits (CRLs) are also determined through this process based on calculations and certification standard analyses outlined in the USATHAMA QA Program (USATHAMA, 1990). A method code associated with each USAEC analysis and laboratory is then assigned and reported with the results. Listings of USAEC certified analytical methods used during the RI, target analytes, and CRLs are presented in Appendix D-1, Table D-1 for the 1995 Field Investigation and in Appendix D-3, Table D-1 for the 1996 Field Investigation.

Harding Lawson Associates

Some standard USEPA methods such as hardness, total organic carbon (TOC), TPHC, and total suspended solids (TSS) have no associated USAEC certification. The USAEC recognizes standard USEPA protocols or internal laboratory methods for these analyses. Listings of USEPA analytical methods used during the RI and project reporting limits are presented in Appendix D-1, Table D-1 for the 1995 Field Investigation and in Appendix D-3, Table D-1 for the 1996 Field Investigation.

3.2.2 Off-Site Laboratory Methods Quality Control

All field samples sent to the laboratory were organized into lots which were assigned a three digit code using letters of the alphabet. Each lot consisted of the maximum number of samples, including QC samples, that can be processed through the rate limiting step of the method during a single time period (not exceeding 24 hours). Associated with each lot were laboratory control samples. Control samples were spikes of high and low concentrations of specific analytes that help monitor ESE's precision and accuracy. The recoveries of these spikes were plotted on control charts generated by ESE and submitted to the USAEC. Data generated during the performance demonstration process were used to calculate a mean of the recoveries. Control and warning limits were statistically generated by the USAEC Chemistry Branch to help measure laboratory data quality. Control charts are generated with each lot providing a continuous benchmark for trend evaluation of laboratory performance.

Method blanks were also analyzed at ESE to evaluate the potential for target analytes to be introduced during the processing and analysis of samples. One method blank was included with each analytical lot. Because analytical lots included samples from several areas, method blank results are presented and discussed for all AOCs investigated during the RI.

3.2.3 Data Reduction, Validation, and Reporting

Initial responsibility for accuracy and completeness of Devens analytical data packages rested with ESE. Secondary data quality reviews were completed by the USAEC Chemistry Branch and HLA. The USAEC Chemistry Branch conducts reviews on control chart data from low and high concentration blank spikes provided with each data lots submitted by the ESE. The processes for reporting control chart data and determining data accuracy is outlined in the USATHAMA QA Program

Harding Lawson Associates

(USATHAMA, 1990). Complete evaluations of QC results including blank data, surrogate and spike recovery data, field duplicate data, and tentatively identified compounds is completed by HLA, and data quality interpretations are provided in Section 7.0 and Appendix D.

All laboratory submissions to the USAEC first underwent a review process, including checks on the data quality, which evaluated completeness of the ESE data, accuracy of reporting limits, compliance with QC limits and holding times, and correlation of ESE data to associated laboratory tests.

The following items were validated by ESE before submission to the USAEC:

- COC records;
- instrument printouts for agreement with handwritten results;
- calibration records to ensure a particular lot is associated with only one calibration;
- chromatograms and explanations for operator corrective actions (such as manual integration);
- standard preparation and documentation of source;
- calculations on selected samples;
- notebooks and sheets of paper to ensure all pages were dated and initialed, and explanations of procedure changes;
- GC/mass spectrometer (MS) library search of unknown compounds; and
- transfer files and records to ensure agreement with analysis results.

Harding Lawson Associates

W003971.080

3.2.4 Data Reporting

After review and validation by ESE, the data were encoded for transmission into the USAEC's Installation Restoration Data Management Information System (IRDMIS) as Level 1 Data. IRDMIS, a computerized data management system used by the USAEC, is described in detail in Subsection 3.3. Once the data were entered into the system, a group and records check was completed. Data were then transferred to USAEC's data management contractor. During this phase, the data were elevated to Level 2. Another group and records check was performed and the data were reviewed by the USAEC Chemistry Branch. When errors were identified, the data were returned to ESE for correction. Control charts were produced by ESE that plotted recoveries of high and low concentrations of laboratory control spikes of the target analytes. The control charts provided the USAEC with information about the accuracy of the analytical methods performed by ESE. Once data were reviewed by the USAEC Chemistry Branch, the determination was made on a lot-by-lot basis whether the data were acceptable. Qualifiers may be added to results to identify quality issues related to data quality. Two types of qualifiers are used for data entered into the IRDMIS data base. Qualifiers include flagging codes which are entered by the subcontract laboratory and data qualifiers which are entered by USAEC Chemistry Branch during the secondary review process described in Subsection 3.2.2. Flagging codes and data qualifiers codes used on the IRDMIS are described in Table 3-2. The data that were accepted were then elevated to Level 3 and made available to USAEC personnel and ABB-ES by modem to a main frame computer. Off-site results are presented in Subsection 7.0 and Appendix L-2.

3.2.5 Field Quality Control Samples

Field QC samples which were collected routinely during the RI included MS/matrix spike duplicates (MSDs), field duplicate samples, rinse blanks, and trip blanks.

Before field investigations were initiated, a sample of water, collected from the source, was used for sampling equipment decontamination. The water source for the RI at AOC 69W was the South Post Water Point (Well D-1). For the purpose of off-site laboratory QC, this was identified as the field blank (source water sample). The field blank data were sent to the USAEC Chemistry Branch where approval was granted for the use of this water in decontamination procedures.

Harding Lawson Associates

As specified in the Fort Devens POP (ABB-ES, 1995a), MS/MSDs were spiked and analyzed for PAL inorganics, and pesticides/polychlorinated biphenyl (PCBs), as well as several USEPA Methods for hardness, total petroleum hydrocarbons, (TPH) by USEPA Methods 9071 and 8105, TOC, total phosphate, nitrate and nitrite-nitrogen, and kjeldahl-nitrogen. HLA personnel made the determination of which samples were to be designated as MS/MSDs. This was noted on the COC forms submitted to ESE.

Samples designated as MS/MSDs were spiked at the off-site laboratory with specified concentrations of analytes to determine matrix effects based on USAEC and USEPA method guidelines. MS/MSD data were also used to assess the accuracy of the analyses used. MS/MSD samples were collected at a rate of one set per 20 samples. During the RI field investigation, samples were collected from AOCs 69W, 57 and 63AX simultaneously. Therefore, assessments of MS/MSD data, contained in Appendix D, were made for these AOCs collectively.

Field duplicate samples were collected at a rate of one per 20 field samples. The purpose of duplicate sample analysis was to assess the sampling and off-site laboratory precision for particular methods. AOCs were investigated simultaneously during the RI field effort, and field duplicates were collected for each media sampled. Duplicate data were assessed collectively for the RIs. Duplicate sample results are presented in Appendix D.

Rinse blanks were collected and analyzed for PAL analytes and TPH by USEPA Methods 418.1 and 8015. Rinse blanks consisted of previously analyzed deionized water which was poured over sampling equipment. Analysis of this water provided information used to evaluate the potential for sample contamination during sample collection. The results were also used to assess decontamination procedures for the sampling equipment. As specified in the Fort Devens POP (ABB-ES, 1995a), rinse blanks were collected at a rate of one per 20 samples. Rinse blank results from the RIs are included in the data quality reports in Appendix D. Discussions regarding rinse blank contamination are relevant to both AOCs investigated during the RIs.

For every shipment of VOC samples to ESE, trip blanks accompanied the samples. The purpose of analyzing trip blanks was to determine if there was any VOC cross contamination during the shipment and handling of samples. The trip blanks consisted of previously analyzed deionized water that was bottled at ESE. Trip

Harding Lawson Associates

blanks were shipped in sealed containers to the job site. As needed, trip blanks were then included with shipments of VOC field samples. Trip blank data are associated with AOCs 69W, 57, and 63AX simultaneously. Data were included for trip blanks sent with samples from all AOCs investigated during the RI. Trip blank data are presented in Appendix D.

3.2.6 Off-Site Analytical Data Quality Evaluation

Off-site data quality reviews were conducted by the project chemist for results generated during the RI. In addition to USAEC laboratory data reviews described in Subsection 3.2.2, precision and accuracy of results were assessed by reviewing MS/MSD results, field duplicate results, and surrogate recovery. QC sample results were compared to goals outlined in the Fort Devens POP (ABB-ES, 1995a) and USEPA Region I validation guidelines (USEPA, 1988; USEPA, 1989). Detailed discussions of these reviews are contained in Appendix D. Conclusions on the precision and accuracy of analytical measurements and possible sample contaminants are summarized in Subsection 7.1.2.

Off-site laboratory data collected during the AOCs 69W, 57 and 63AX RIs at Devens were evaluated for possible off-site laboratory or sampling-related contamination. QC blank results were evaluated to assess the potential for sample contamination during sample collection or at the off-site laboratory. This evaluation did not include validation according to USEPA guidelines. Sample results reported and discussed in this report were not adjusted for reported analytes that were also detected at similar concentrations in blanks associated with that sample; action levels were not established, and the 10X rule was not applied to compounds considered by the USEPA to be common laboratory contaminants. Examples of these contaminants include the VOCs acetone, methylene chloride, and the phthalate semivolatile organic compounds (SVOCs). Likewise, action levels for other analytes using the 5X rule application were not established. Analytes that would have been below these action levels were not removed from the data as they would have been in the USEPA validation process.

General trends relating to blank and sample contamination were examined. Comparison of blank data with results from the entire data set are discussed as a data assessment. Assessments are made based on analyte detection in blanks, the frequency of the detection and the concentrations of these analytes. A summary of

Harding Lawson Associates

blank contamination is presented in Subsection 7.1.2 of this report. Some analytes are interpreted to represent non-site related contamination in the contamination assessments presented in Section 7.0

3.3 CHEMICAL DATA MANAGEMENT

Chemical data were managed by HLA's Sample Tracking System and the USAEC's IRDMIS. These systems are described in the following sections.

3.3.1 Sample Tracking System

HLA employed its computerized Sample Management System to track environmental samples from field collection to shipment to the off-site laboratory. HLA also tracked the status of analyses and reporting by the off-site laboratory.

Each day the field sampling teams carried computer-generated sample labels into the field that stated the sample control number, sample identification, size and type of container, sample preservation summary, analysis method code, and sample medium. The labels also provided space for sampling date, time, depth (if applicable), and the collector's initials to be added at the time of collection.

After collection in the field, the samples were stored on ice for transport back to the HLA field office. Samples were temporarily stored in the HLA field office refrigerator. They were checked-in on the field office computer, and the collector's initials and the sampling date and time were entered. The system would then indicate the sample status as "COLLECTION IN PROGRESS."

When the samples were prepared for shipment, they were "RELEASED" by the sample management system. Upon request, the system printed an Analysis Request Form (ARF) and a COC, which were signed and included with the samples in the shipment. The system would then indicate the sample status as "SENT TO LAB."

This system substantially reduced the time required for preparation of sample tracking documentation, and it provided an automated record of sample status.

Harding Lawson Associates

After shipment of samples to the off-site laboratory, ABB-ES continued to use the sample tracking system to track and record the status of the samples, including the date analyzed (to determine actual holding time), the date a transfer file was established by ESE, and the date the data were sent to IRDMIS (see Subsection 3.3.2).

3.3.2 Installation Restoration Data Management Information System

IRDMIS is an integrated system for collection, validation, storage, retrieval, and presentation of data of the USAEC's Installation Restoration and Base Closure Program. It uses personal computers (PCs), a UNIX-based minicomputer, printers, plotters, and communications networks to link these devices.

For each sample lot, HLA developed a "provisional" map file for the sample locations, which was entered into IRDMIS by Potomac Research, Inc. (PRI), USAEC's data management contractor.

Following analysis of the sample lot, ESE created chemical files using data codes provided by HLA and entered the analytical results (Level 1) on a PC in accordance with the User's Manual (PRI, 1993). For each sample lot, a hard copy was printed, reviewed, and checked by ESE's Laboratory Program Manager. ESE created a transfer file from accepted records which was sent to HLA (Level 2). HLA performed a group and record check and sent approved records in a chemical transfer file to PRI. PRI checked the data and, if accepted, entered it in to the IRDMIS minicomputer (Level 3). Level 3 chemical data are the data used for evaluating site conditions, and are the data used in this AOC 69W report.

3.4 1997 SUPPLEMENTAL ANALYSES

Additional samples were collected in the Fall 1997 for off-site laboratory analyses to supplement existing data sets. Groundwater samples were analyzed for volatile petroleum hydrocarbons (VPH) and extractable petroleum hydrocarbons (EPH) using Massachusetts methodologies (MADEP, 1995a; MADEP, 1995b), and USEPA Methods 8260B and 8270B (USEPA, 1996). Air samples were also collected and analyzed for VOCs using USEPA Method TO14. Laboratory data packages were reviewed by the HLA project chemists, and data were validated in accordance with

Harding Lawson Associates

W003971.080

USEPA Region I guidelines. Data validation reports are presented in Appendix D-4.

Harding Lawson Associates

4.0 APPLICABLE OR RELEVANT AND APPROPRIATE REQUIREMENTS IDENTIFICATION

CERCLA was enacted by Congress in 1980, establishing the Superfund program. The regulations implementing this program are found in 40 CFR Part 300, also known as the National Contingency Plan (NCP). CERCLA was amended in 1986 by SARA, which mandated that the level or standard of control specified in a remedial action be "at least that of any ARAR standard, requirement, criteria, or limitation under any federal environmental law, or any more stringent standard, requirement, criteria or limitation promulgated pursuant to a state environmental statute." SARA also established that the requirements of the NCP apply to federal facilities.

The purpose of the RI was to determine the nature and distribution of site-related soil and groundwater contamination at AOC 69W. In order to evaluate whether there is a potential threat to human health and the environment, preliminary ARARs are identified in this section and will then be compared to site-specific data. ARARs are federal and state human health and environmental requirements used to (1) evaluate the distribution of site impacts and the appropriate extent of site cleanup; (2) define and formulate remedial action alternatives; and (3) govern implementation and operation of the final remedy.

Identification and evaluation of ARARs is an iterative task, necessary throughout the remedial response process. Therefore, the preliminary lists of requirements identified for AOC 69W and their relevance may change as more information is obtained, as the preferred alternative is chosen, and as the design and approach to remediation becomes more refined.

<u>Applicable Requirements</u> - Applicable requirements are those cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations promulgated under federal or state law that specifically address a hazardous substance, pollutant, contaminant, remedial action, location, or other circumstance that have jurisdiction at a site. An example of an applicable requirement is the use of the Safe Drinking Water Act (SDWA) MCLs drinking water standards for a site where hazardous substances have caused water in a public water supply distribution system to become contaminated.

Harding Lawson Associates

<u>Relevant and Appropriate Requirements</u> - Relevant and appropriate requirements are cleanup standards, standards of control, and other substantive environmental protection requirements, criteria, or limitations that, while not "applicable" to a hazardous substance, pollutant, contaminant, remedial action, location or other circumstance at a site, address problems or situations sufficiently similar to those encountered at the site that their use is well-suited to the particular site. For example, MCLs for drinking water would be relevant and appropriate requirements at a site where hazardous substances are found in or could enter drinking water classified as a current or future drinking water source. When a requirement is found to be relevant and appropriate, it is complied with to the same degree as if it were applicable.

<u>To be Considered (TBC) Material</u>. Non-promulgated advisories or guidance issued by the federal and state government are not legally binding and do not have the status of potential ARARs. However, in many circumstances, TBCs will be considered along with ARARs as part of the site risk assessment, and may be used in determining the level of cleanup for protection of human health or the environment.

ARARs that pertain to the remedial response can be classified into three categories: chemical-, location-, and action-specific. The following subsections provide an overview of these ARARs.

4.1 CHEMICAL-SPECIFIC ARARS

Because of their site-specific nature, the identification of ARARs requires an evaluation of the federal, state, and local environmental regulations with respect to chemicals of concern and site characteristics. Chemical-specific ARARs generally involve health- or risk-based numerical values or methodologies that establish site-specific acceptable chemical concentrations or amounts. These values are used to develop action levels or cleanup concentrations.

4.1.1 Groundwater

Table 4-1 sets forth the federal chemical-specific ARARs and TBC information for groundwater. USEPA SDWA MCLs are legally applicable to contaminants found

Harding Lawson Associates

in public water systems that have at least 15 service connections or serve an average of at least 25 people daily at least 60 days per year. Even when not legally applicable, MCLs may be relevant and appropriate to groundwater remediation. Non-zero maximum contaminant level goals (MCLGs) are health-based goals at which no known or anticipated adverse effects on health will occur and could be relevant and appropriate standards if considered as cleanup levels. Table 4-1 also includes the current version of USEPA Region III risk-based concentrations (RBCs) which are commonly used as TBC information at CERCLA sites. The surface water criteria set forth in Table 4-1 are TBC information and will only be applicable if a discharge to surface water will be part of the groundwater remedial action.

Table 4-2 sets forth the state chemical-specific ARARs and TBC information for groundwater. The Commonwealth of Massachusetts has developed drinking water standard and guidelines, expressed in terms of maximum levels of contaminants allowed in drinking water. Groundwater data from AOC 69W will be applied to Massachusetts Maximum Contaminant Levels (MMCLs), Massachusetts Class I groundwater quality standards, and/or USEPA Region III RBCs for tap water.

4.1.2 Soil

Table 4-3 sets forth the soil screening levels (TBCs) from the current USEPA Region III RBC documents.

4.1.3 Massachusetts Contingency Plan

The NCP provides that CERCLA response actions must comply with environmental and public health laws and regulations to the extent they are substantive (i.e., pertain directly to actions or conditions in the environment), but do not need to comply with those that are administrative (i.e., mechanisms that facilitate the implementation of the substantive requirements).

The provisions of the MCP, 310 CMR 40.0000 (January 13, 1995) are mostly administrative in nature and, therefore do not have to be complied with in connection with the response action selected for AOC 69W. Further, the MCP contains a specific provision (310 CMR 40.0111) for deferring application of the MCP at CERCLA sites. As stated in the MCP, response actions at CERCLA sites are deemed adequately regulated for purposes of compliance with the MCP, provided

Harding Lawson Associates

the MADEP concurs in the CERCLA Record of Decision (ROD).

However, some provisions of the MCP contain substantive requirements that may be ARARs. Section 310 CMR 40.0940 sets forth three methods of risk characterization. Section 310 CMR 40.0942 provides that any of the three methods may be used, subject to certain specified limitations. MCP Method 1 establishes specific numerical standards for certain listed contaminants (see 310 CMR 40.0974.-0975). Since MCP Method 1 contains promulgated numerical standards, it may be an ARAR if this method is selected.

MCP Method 3 does not contain substantive numerical standards; rather it provides a risk characterization methodology to determine the appropriate cleanup level (see 310 CMR 40.0991.-0996). Because MCP Method 3 is a methodology and does not contain substantive standards, and because it defines protectiveness in a way which is inconsistent with the CERCLA NCP, Method 3 is not an ARAR which has to be met. Therefore, these standards of the MCP do not apply to the remedial response at AOC 69W.

4.2 LOCATION-SPECIFIC ARARS

Location-specific ARARs represent restrictions placed on the concentration of hazardous substances or the conduct of activities because of the location or characteristics of a site. These ARARs set restrictions relative to special locations such as wetlands, floodplains, sensitive ecosystems, as well as historic or archeological sites, and provide a basis for assessing existing site conditions. Table 4-4 lists location-specific federal and state requirements.

Some of the location-specific ARARs for areas such as wetlands and floodplains may or may not be applicable, or relevant and appropriate, depending on the remedial action selected because the regulations do not apply unless some activity is conducted in a certain defined area.

Harding Lawson Associates

4.3 ACTION-SPECIFIC ARARS

Action-specific ARARs involve design, implementation, and performance requirements that are generally technology- or activity-based. Action-specific ARARs, unlike location- and chemical-specific ARARs, are usually technology- or activity-based limitations that direct how remedial actions are conducted. After remedial alternatives are developed, the evaluation of action-specific ARARs is one criterion for assessing the feasibility and effectiveness of compliance with proposed remedial alternatives. The applicability of this set of requirements is directly related to the particular remedial activities selected for the site. Table 4-5 represents an overview of potential action-specific ARARs that may or may not ultimately be applicable to AOC 69W.

4.4 BACKGROUND CONCENTRATIONS

As a means to evaluate concentrations of inorganic analytes detected in samples collected as part of each phase of investigation, background concentrations were calculated for the Fort Devens installation. Background concentration calculations were based on analytical data results gathered from soil and groundwater samples collected throughout the Fort Devens installation, selected as representative of background (non-contaminated) conditions. Although most of the calculations include assumptions on both the distribution of chemical concentrations and on the selection of representative samples that are not statistically rigorous, the results are considered representative of actual background concentrations at Fort Devens.

For soil, chemical data gathered from 20 soil samples collected by Ecology & Environment, Inc. (E&E) as part of their Group 1A and 1B investigation activities were used. The samples were collected from the major soil associations throughout Fort Devens specifically to establish background concentrations of inorganic analytes in soil. The background soil samples were collected from locations that were visually undisturbed, at least 50 feet from any road, and 300 feet from any known SA.

The calculations were performed on 22 of the 23 PAL inorganic analytes (no data was available for thallium). For analytes that were not detected in the majority of soil samples, the detection limit for that analyte was selected as the background concentration. Sample location, data ranges, mean values, details of calculations, and

Harding Lawson Associates

calculated background concentrations are summarized in Appendix K.

For groundwater, HLA selected 10 representative groundwater samples collected from the Round I groundwater sampling events, completed in 1992, for Groups 2, 3, 5, 6 and 7 for the purpose of calculating background inorganic analyte concentrations in groundwater. Representative groundwater samples were selected from specific monitoring wells located upgradient of a SA, exhibiting low TSS and/or low aluminum concentrations. Aware that elevated TSS concentrations artificially elevate inorganic analyte concentrations, HLA selected samples that exhibited TSS concentrations on the same order of magnitude as the South Post Water Point (Well D-1). Because a close correlation between TSS concentrations and aluminum concentrations was observed in all the groundwater samples analyzed, the aluminum concentration was used as an alternate selection criterion in the absence of TSS data. The concentration values detected in the ten samples were calculated using the same assumptions on outliers and detection limits applied to the soils background concentration calculations. The statistical analysis calculations for groundwater inorganics, and the resulting background concentrations, data ranges, mean values, and details of the calculations are also provided in Appendix K.

Harding Lawson Associates

5.0 AOC 69W REMEDIAL INVESTIGATION

5.1 SITE BACKGROUND AND CONDITIONS

AOC 69W is located on the northern portion of the Main Post near the northeast corner of the intersection of MacArthur Avenue and Antietam Street (Figure 5-1). AOC 69W is comprised of the former Fort Devens Elementary School (Building 215), the associated parking lot, and adjacent lawn extending approximately 300 feet northwest to Willow Brook (Figure 5-2).

A reference in the 1971 Ayer Annual Town Report indicated that the Fort Devens Elementary School was constructed in 1951. Various historical site plans indicated that heating oil for the Elementary School was provided by a 10,000 gallon underground storage tank (UST) located in what is now the school courtyard. In 1972, an addition was added to the school resulting in the current building. As part of the addition a new boiler room was added to complement the existing boiler. The 10,000 gallon UST, located in what is now the courtyard, and associated piping were removed and a new 10,000 gallon UST was installed under the parking lot on the north side of the school. This UST serviced the school until the school's closure in 1993. A review of historical records and a series of personal interviews indicated that there have been two separate releases of fuel oil at AOC 69W, the first in 1972 and the second in 1978.

The 1972 fuel oil release was due to a crimp in the piping which ran from the new 10,000 gallon UST to the new boiler room. It has been estimated that approximately 7,000 to 8,000 gallons of fuel oil were released into soil and groundwater prior to repair of the piping. The exact location of the release is unknown; however, some visual evidence and contaminant distributions suggest that the release was in the vicinity of the boiler room.

As a result of the fuel oil release a "skimmer system" was installed next to the UST in either late 1972 or early 1973. The nature and exact location of the system was unclear; however, some evidence suggested that the system is little more than a french drain. It is known that the system was connected to, or possibly comprised of, a pipe buried approximately three feet below ground surface extending from the vicinity of the 10,000 gallon UST to a buried 250 gallon UST located approximately

Harding Lawson Associates

250 feet to the northwest. The 250 gallon UST collected oily water and was pumped out approximately every three months. Sometime after 1986 the UST was filled with crushed rock. The "skimmer system" was excavated and removed during the 1997-1998 soil removal action. Details and descriptions of the removal action, piping, and value are provided in Subsection 5.5.

The 1978 fuel oil release resulted from a failed piping joint from fuel oil pipes leading to the old boiler. Approximately 7,000 to 8,000 gallons of fuel oil were released into soil and groundwater during the 1978 incident. A large hole was excavated on the north side of the school adjacent to the loading dock in an attempt to locate the source of the release. Reports indicate that the excavation collected residual oil for one month before the damaged piping was found and replaced. Shortly after the release an oily sheen was reported in Willow Brook and the associated wetlands to the north of the school. Following the spill, 2,600 gallons of residual oil were pumped from the 250 gallon UST.

5.2 SUMMARY OF PREVIOUS INVESTIGATIONS

The following subsections discuss results of the AREE 69W site evaluation and investigation performed at AOC 69W by ADL. A brief summary of analytical data is presented to demonstrate the need for subsequent investigations at the site. A complete assessment of the analytical data is presented in Section 7.0 of this RI report. The scope of investigation activities performed at AOC 69W is summarized chronologically in Table 5-1.

5.2.1 ADL AREE 69 Evaluation (AREE 69W)

In July of 1993 ADL investigated the Fort Devens Elementary School, designated AREE 69W, as part of the basewide AREE 69 (Past Spill Sites) evaluation. The investigation focused on the 1978 fuel oil release and was comprised of a document review and site visit. The study concluded that there was a potential for fuel oil contamination in the soil and groundwater (ADL, 1995).

Further investigation was performed at AREE 69W from March through June of 1994. The investigation involved sampling, field screening, and laboratory analysis of surface soil, subsurface soil, groundwater, surface water, and sediment, and a

Harding Lawson Associates

geophysical survey to locate subsurface utilities.

5.2.1.1 Surface Soil. Six surface soil samples were collected from a depth of 0 to 1 foot below ground surface (bgs), from the grassy area north and northwest of the parking lot (Figure 5-2). The samples were analyzed by portable field instruments for BTEX and TPHC, using GC and NDIR techniques, respectively. The sample with the highest observed TPHC concentration was submitted for laboratory analysis of PAL VOCs, PAL SVOCs, TPHC, PAL inorganics, and TOC. AREE 69W field analytical soils data are presented in Table 5-2. AREE 69W off-site analytical soils data are presented in Table 5-3.

TPHC field screening concentrations obtained from the surface soil samples ranged from 9.5 parts per million (ppm) to a high of 131 ppm (observed at location HA-5, located just off the northwest corner of the paved area). No BTEX were detected in the screening results. Laboratory results from the single surface soil sample submitted for laboratory analysis revealed no compounds at concentrations exceeding Massachusetts Contingency Plan (MCP) Method 1, S-1/GW-1 Standards. Carcinogenic polycyclic aromatic hydrocarbons (cPAHs) detected in the surface soils at the site consisted of benzo(a)anthracene and chrysene at a combined concentration of 0.29 μ g/g.

5.2.1.2 Subsurface Soil. Subsurface soil samples were collected during the installation of groundwater monitoring wells and during the Geoprobe investigations. During the first round of Geoprobe sampling, subsurface samples were collected from 0 to 2 and 3 to 5 feet bgs at 16 locations (Figure 5-2) for field analysis of TPHC and BTEX. Of the 32 samples analyzed in the field, three samples exhibiting the highest TPHC concentrations and one sample with the lowest TPHC concentration were submitted for laboratory analysis of PAL VOC, PAL SVOCs, TPHC, PAL inorganics and TOC analysis. During the second Geoprobe sampling round, nine additional locations were investigated (Figure 5-2). Subsurface soil samples were collected from a depth of 3 to 5 feet bgs and field screened for TPHC. AREE 69W field analytical data are provided in Table 5-2. Off-site analytical soils data are provided in Table 5-3.

Subsurface soil samples were collected at depth intervals of 0 to 2 feet, 2 to 4 feet, and 11 to 13 feet bgs during the monitoring well installation effort. These samples were screened in the field for TPHC and BTEX. The samples from the 2 to 4 and

Harding Lawson Associates

4 to 6 foot depth intervals were submitted for laboratory analysis of TPHC, PAL VOCs, PAL SVOCs, PAL inorganics, and TOC analysis.

TPHC concentrations in soils collected with the Geoprobe and from monitoring well, soil borings ranged from 7.5 ppm to 15,500 ppm (GP-2 at 3 to 5 feet bgs). Concentrations of benzo(b)fluoranthene, benzo(a)anthracene, and chrysene detected in Geoprobe soils were above MCP Method 1, S-1/GW-1 standards.

Based on the field screening and laboratory analysis results, TPHC and cPAH soil contamination appears to be concentrated in the area of the existing UST (the presumed source area), and may have migrated downgradient towards Willow Brook.

5.2.1.3 Groundwater. Groundwater samples were collected from each Geoprobe location and from the six newly installed groundwater monitoring wells. Sixteen groundwater samples were collected during the first Geoprobe sampling round and field screened for TPHC and BTEX (GP-01 through GP-16). Filtered and non-filtered groundwater samples collected during the second Geoprobe sampling round were field screened for TPHC (GP-17 through GP-25).

Field screening results from the 25 Geoprobe groundwater samples indicated that TPHC was present in groundwater. BTEX was not detected. Five sample locations from the first Geoprobe sampling round exhibiting the highest field screening TPHC concentrations were resampled and submitted to the laboratory for analysis of PAL VOCs, PAL SVOCs, TPHC and water quality parameters. No samples from the second geoprobe sampling round were sent for laboratory analysis. Results indicated that TPHC, inorganic analytes (arsenic, lead, antimony, beryllium, chromium, and nickel), and organic compounds (1,1-dichloroethene, benzene, carbon tetrachloride, chloroform, tetrachloroethene, trichloroethene, 2-methyl naphthalene, and naphthalene) were detected at concentrations exceeding MCP Method 1 GW-1 Standards. Most of these exceedances occurred at locations GP-1, GP-2, GP-6 and GP-15, located in the area of the UST and downgradient of this location. No cPAHs were detected in the Geoprobe groundwater samples. AREE 69W field analytical results are provided in Table 5-2. Results of off-site analysis of groundwater samples are provided in Table 5-4.

Six monitoring wells installed at the site confirmed the results of the Geoprobe investigation. Groundwater samples were submitted for analysis of TPHC, PAL

Harding Lawson Associates

VOCs, PAL SVOCs, unfiltered inorganics and water quality parameters. Results indicated that TPHC, arsenic, beryllium, cadmium, chromium, lead, nickel, 2-methyl naphthalene, acenaphthene, and naphthalene were detected at concentrations exceeding MCP Method 1 GW-1 Standards. These exceedances occurred at monitoring wells 69W-94-10, 69W-94-11, 69W-94-13 and 69W-94-14 (Figure 5-2). No cPAHs were detected in the groundwater samples.

Groundwater sample results indicate that the area around the UST has the greatest number of compounds exceeding MCP Standards. Groundwater northwest of the UST was also found to have elevated concentrations of inorganics and TPHC, suggesting that contaminants have potentially migrated downgradient of the UST location.

5.2.1.4 Surface Water and Sediment. Surface water and sediment samples were collected from two locations in Willow Brook (Figure 5-2). One sample location (69W-94-16) was placed in line with the inferred plume migration pathway indicated by the Geoprobe survey, and the other (69W-94-15) was placed upstream of this area. Samples were analyzed for TPHC, PAL VOCs, PAL SVOCs, unfiltered inorganics, and water quality parameters. Surface water and sediment off-site analytical results are presented in Tables 5-5 and 5-6, respectively.

The results indicated the presence of cPAHs in both sediment samples, and TPHC in sample 69W-94-16. Specifically, the cPAHs benzo(a)anthracene, chrysene, benzo(b)fluoranthene, and benzo(k)fluoranthene were detected in the 69W-94-15 (upstream) sediment sample. In sediment sample 69W-94-16, TPHC, benzo(a)anthracene and chrysene were detected. Total cPAHs in the upstream sample barely exceeded 7.0 ppm. Total cPAHs in the downstream sample (69W-94-16) were an order of magnitude less than the clean-up values. Other PAHs and metals were detected in both samples.

TPHC and cPAHs were not detected in surface water samples.

Harding Lawson Associates

W003971.080

5.3 REMEDIAL INVESTIGATION PROGRAM

5.3.1 Technical Objectives

The following subsections present the technical objectives of the AOC 69W RI field investigations.

5.3.1.1 Geophysical Survey. The technical objective of the geophysical survey is to gather AOC-wide, non-intrusive data on subsurface features. Specific objectives included: confirming that the fuel oil UST and ancillary piping were removed from the school courtyard, clearing exploration locations for subsurface utilities, and providing information on subsurface geology.

5.3.1.2 TerraProbeSM and Soil Borings. The technical objective of the TerraProbeSM and soil boring program was to obtain representative soil and groundwater samples for conducting field screening analyses, off-site laboratory analyses, and grain size distribution. In addition, borings were intended to yield data to further define horizontal and vertical VOC and TPHC contaminant distribution and characterize the geologic setting at AOC 69W. This includes location of the site within the regional stratigraphic setting and identification of heterogeneous soil layers.

5.3.1.3 Monitoring Wells. The technical objective of the monitoring well installation program was to characterize local groundwater flow, and groundwater quality, and define contaminant distribution and source areas. This included collecting water level data to establish flow directions, vertical and horizontal gradients, and estimate the hydraulic conductivity of the geologic strata.

5.3.1.4 Piezometer Installation. Piezometers were installed to determine vertical hydraulic gradients in the vicinity of Willow Brook. Hydrogeologic data obtained from the piezometers helped to define Willow Brooks relationship to the local and regional hydrogeology.

5.3.1.5 Test Pitting. The technical objective of the test pitting was to investigate subsurface geophysical anomalies identified during the geophysical survey. The test pitting was also intended to help define subsurface stratigraphy.

Harding Lawson Associates

5.3.1.6 Field Laboratory Analysis. The technical objective of the field analytical program was to quickly generate USEPA Level II analytical data for previously identified site-related compounds (specifically TPHC and BTEX) allowing for preliminary identification of contaminant distribution. This information enabled more accurate placement of groundwater monitoring wells. In addition, field analytical data were used to support the results and findings of the contaminant assessment. Field analytical techniques employed as part of this RI are discussed in detail in Subsection 3.1 of this report. Results of the field analyses are discussed in detail in Section 7.0 of this report.

5.3.1.7 Off-Site Laboratory Analytical Sampling. The technical objective of the analytical sampling program was to supplement the existing field analytical data for subsurface soil and groundwater at AOC 69W, and to better characterize site conditions in surface soil, subsurface soil, groundwater, and sediment. This database was used as the foundation for the contamination assessment, fate and transport discussion, and risk assessment. Off-site laboratory analytical methods employed as part of this RI are discussed in detail in Subsection 3.2 of this report. Results of the analytical data are discussed in detail in Section 7.0 of this report.

5.3.1.8 Air Sampling. The objective of the air sampling was to determine if fuelrelated contaminants present in soil and groundwater beneath the school are impacting air quality within the school. Analytical data from the 1997 air sampling event are discussed in the contamination assessment, fate and transport discussion, and risk assessment. Analytical protocol are discussed in Subsection 3.3 of this report. Due to sample collection issues with the first air sampling event completed in 1996, air sample analytical data for 1996 are not discussed in this report.

Harding Lawson Associates

5.3.2 Data Quality Objectives

Data Quality Objectives (DQOs) are qualitative or quantitative statements developed by the data user to specify the quality of data needed from a particular activity to support specific decisions. DQOs are the starting point in the design of the RI. The DQO development process matches sampling and analytical capabilities to the data targeted for specific uses and ensures that the quality of the data does not underestimate project requirements. The procedures of the QA objectives presented in Section 3.0 of Volume I of the Fort Devens POP (ABB-ES, 1995a) were followed during the RI field programs at AOC 69W. This subsection includes a general scope of work, DQOs, and the QA/QC approach.

Analyses were conducted on samples collected from AOC 69W to evaluate the nature and distribution of the contaminants detected in the previous AREE 69W investigation. On-site field analysis conformed with the guidelines presented in Subsection 4.6 of Volume I of the Fort Devens POP (ABB-ES, 1995a). Off-site laboratory analytical procedures are presented in Section 7.0 of Volume I of the Fort Devens POP (ABB-ES 1995a), and the Laboratory QA Plan and the USAEC Certified Analytical Methods procedures are presented in Appendices B and C, respectively, in Volume II of the Fort Devens POP (ABB-ES, 1995a).

USEPA has identified five general levels of analytical data quality as being potentially applicable to field investigations conducted at potential hazardous waste sites under CERCLA (USEPA, 1987). The five levels were replaced by two levels including definitive data, and screening data with confirmation by the USEPA after the AOC 69W investigation was initiated (USEPA, 1993). For consistency with the Fort Devens POP (ABB-ES, 1995a) and historic planning documents, the five data quality levels are referenced in this RI. These levels are summarized as follows:

- <u>Level I</u> Field Screening. This level is characterized by the use of portable instruments that can provide real time data to assist in optimization of sampling point locations and for health and safety support. Data can be generated regarding the presence or absence of certain contaminants (especially volatiles) at sampling locations.
- Level II Field Analysis. This level is characterized by the on-site use of portable analytical instruments and mobile laboratories which can

Harding Lawson Associates

render qualitative and quantitative data.

 <u>Level III</u> - Laboratory analysis using methods other than Contract Laboratory Program (CLP) Routine Analytical Services (RAS). This level is used primarily to support engineering studies using standard USEPA-approved procedures. Some procedures may be equivalent to the USEPA RAS, without the CLP requirements for documentation.

- <u>Level IV</u> CLP RAS. This level is characterized by rigorous QA/QC protocols and documentation, providing qualitative and quantitative analytical data.
- <u>Level V</u> Non-standard methods. This level includes analyses that may require modification and/or development. CLP Special Analytical Services are considered Level V.

For AOC 69W RI efforts, field measurements such as pH, temperature, conductivity, and readings from a PID and O_2 /Explosimeter constituted Level I field analytical data. Field GC analysis constituted Level II field analytical data. Off-site analyses of soil and groundwater for organics, inorganics, TOC, TPHC, water quality parameters, and pesticides and PCBs were considered approximately equivalent to USEPA analytical support Level III. The sampling approaches and analytical procedures described in the Fort Devens POP (ABB-ES, 1995a) have been selected to meet Level III data quality.

DQOs were established to support the level of detail required for RI activities. Data generated during field and off-site laboratory tasks were used to characterize AOC 69W conditions and to perform baseline risk assessments.

DQOs and QC for field measurements and laboratory analyses conform to USAEC and USEPA requirements (as specified in the USACE Quality Assurance Manual, 1990, and Guidance for Conducting Remedial Investigations and Feasibility Studies Under CERCLA, 1988).

USAEC requirements and analytical processes are discussed in Section 3.0 of this report. They focus on the use of laboratory control spikes in associated data lots to measure the performance of the laboratory in the use of USAEC methods. Many

Harding Lawson Associates

of the USAEC methods are identical to standard USEPA methods. The performance demonstration process, required by laboratories performing USAEC work, is discussed in Subsection 3.2.1. The data review and evaluation process are described in Subsection 3.2.3.

Off-site laboratory data were evaluated for precision, accuracy, representativeness, completeness and comparability (PARCC) to meet USEPA Level III requirements. This was accomplished through the collection of field quality control blanks such as field blanks, trip blanks, and equipment rinsate blanks, and through the evaluation of laboratory blanks such as method blanks. The specific purpose of collecting each of these is discussed in Section 3.0 of this report. Laboratory control spikes are run in the certification process to generate control charts that help to establish control limits that are used to ensure accuracy of the results. This process is also described in the text of the report in Section 3.0. Matrix spike and duplicate samples were also analyzed to meet PARCC data quality objectives. These are presented in Appendix D-1.

The precision of the data is a measurement of the ability to reproduce a value under certain conditions. It is a quantitative measurement based on the differences of two values. Precision was evaluated using the RPD of MS/MSD sample pairs and field duplicate sample pairs. Evaluations of the precision of the data are found in Appendix D.

Accuracy measurements identify the performance of a measurement system based on tests with known values. The off-site laboratory, sampling, and media effects on accuracy were assessed by reviewing the percent recoveries of spiked analytes for MS/MSDs, off-site laboratory control samples, and surrogate compounds.

Representativeness refers to the extent to which a measurement accurately and precisely represents a given population within the accepted variation of laboratory and sampling measurements. Collection techniques that obtained samples characteristic of the matrix and location being evaluated were chosen. Historical information was used to identify sample locations. Representativeness was also evaluated using method blanks and field QC sample data. By evaluating method blank and field QC samples, false positive results should be identified. Representativeness was also measured by evaluating field duplicate pair precision.

Harding Lawson Associates

Completeness refers to the percentage of usable, valid values obtained through data evaluation. Completeness was determined by the success rate in meeting holding time criteria and acceptance of sample lots by USAEC.

Comparability is a qualitative assessment describing the confidence with which one data set may be compared with another. Comparability was assured using standard operating procedures for sampling, and by reporting analytical results in standard units.

5.4 SUMMARY OF REMEDIAL INVESTIGATION FIELD PROGRAMS

Several field efforts comprised the RI field program. The initial RI field effort was performed in the fall of 1995. As a result of the findings of this field effort it was determined that additional work was necessary to characterize contaminant distribution at AOC 69W. Additional phases of work were performed as Modification 001 conducted in the falls of 1996 and 1997.

The RI techniques were conducted in conformance with the Revised Final Task Order Work Plan for AOC 57, AOC 63AX, and AOC 69W (ABB-ES, 1996a), the Final RI/FS Task Order Work Plan Addendum for AOC 69W (ABB-ES, 1996b), the Final RI/FS Work Plan Addendum for Supplemental Air Sampling for AOC 69W (ABB-ES, 1997a), the Analytical Approach for AOC 69W Groundwater Sampling (ABB-ES, 1997), and the Fort Devens POP (ABB-ES, 1995a).

The RI field program for AOC 69W included:

- a geophysical survey consisting of EM-31, EM-61, magnetometer, and ground penetrating radar;
- twenty-nine TerraProbesM points (ZWR-95-26X through ZWR-95-28X, and ZWR-95-30X through ZWR-95-55X);
- three soil borings (ZWB-95-01X, ZWB-95-02X, and ZWB-96-03X);
- eight surface soil samples (ZWS-95-35X, ZWS-95-37X, ZWS-95-38X, ZWS-95-39X, ZWS-95-42X, ZWS-95-45X, ZWS-95-46X, and ZWS-95-

Harding Lawson Associates

47X);

- nine sediment and three toxicity test samples collected in Willow Brook (ZWD-95-01X through ZWD-95-06X);
- installation and development of seven monitoring wells (ZWM-95-15X through ZWM-95-18X and ZWM-96-19X through ZWM-96-21X) and two piezometers (ZWP-95-01X and ZWP-95-02X);
- two rounds of groundwater samples collected from the newly installed and the existing monitoring wells, and two supplemental rounds of groundwater sampling using low-flow techniques;
- one test pit (ZWE-95-01X);
- indoor air sampling at 12 locations and in one monitoring well headspace (ZWA-97-01X through ZWA-97-13X) inside the elementary school;
- field analysis of soil samples from soil borings, TerraProbesM points, and test pits by GC and IR, and groundwater by GC;
- aquifer testing on the new and existing monitoring wells; and
- horizontal and vertical survey of all RI explorations.

HLA established a project field office in Building 2012 on the former Main Post. The field office was used for equipment storage and maintenance, sample management, shipping and receiving, staff meetings, and communications. A telephone and facsimile machine were maintained in the field office, and each field crew was issued a hand-held cellular phone. A central equipment decontamination pad was constructed near Building 202. HLA and subcontractor staff were briefed about the nature of AOC 69W, health and safety information, Devens traffic regulations, and key technical requirements. HLA began implementation of the AOC 69W field program in September 1995 with equipment mobilization and a geophysical survey.

Harding Lawson Associates

W003971.080

The subcontractors used by HLA in conducting the RI program were as follows:

- D.L. Maher, North Reading, MA Drilling and monitoring well installation (1995);
- New Hampshire Boring, Londonderry, NH Drilling and monitoring well installation (1996);
- ENPRO Services, Inc., Newburyport, MA Test pit excavation;
- ESE, Gainesville, FL Chemical analysis of environmental samples;
- Tundra Air Consultants, Woburn, MA 1996 Indoor air quality sampling and assessment, Devens Elementary School;
- Cashins & Associates, Reading, MA 1997 Indoor air quality sampling and assessment, Devens Elementary School;
- ENSR Consulting and Engineering, Acton, MA Chemical analysis of air samples;
- Springborn Laboratories, Inc., Wareham, MA Sediment toxicity testing and evaluation; and
- Martinage Engineering Associates, Inc., Reading, MA Surveying of site explorations.

All field activities were conducted in accordance with the Fort Devens POP (ABB-ES, 1995a) and USAEC's Geotechnical Guidelines (USAEC, 1987). The following subsections describe the field activities for the RI at AOC 69W.

5.4.1 Surficial Geophysical Survey

EM-31 and EM-61 surveys were conducted on a 20 foot grid in the parking lot and grassy area to the north of the elementary school at AOC 69W in an attempt to locate the "skimmer system" and associated 250 gallon UST as well as to gather data on subsurface geology and identify subsurface utilities prior to intrusive explorations.

Harding Lawson Associates

Anomalies identified during the EM-31 and EM-61 surveys were further investigated with GPR. A GPR survey was also performed within the school courtyard to confirm that the 10,000 gallon fuel oil UST and the ancillary piping had actually been removed. Geophysical data and interpretation are provided in Appendix C. Surficial geophysical survey procedures are outlined in Subsection 4.4.3 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

5.4.2 TerraProbeSM Investigation and Surface Soil Sampling

HLA's van-mounted TerraProbeSM System was used to collect soil and groundwater samples for chemical field analysis and soil samples for off-site confirmatory analysis. The TerraProbeSM pushed (hydraulically) and hammered (with vibratory hammer) a 1-inch-diameter steel probe into the ground, which allowed collection of subsurface soil samples from discrete depths. Ninety-two soil samples were collected from the 29 TerraProbeSM points at depths ranging between 0 and 18 feet bgs and analyzed in the field for TPHC using IR (USEPA Method 418.1), and for BTEX, select chlorinated solvents, and TPHC by GC (USEPA Method 8015, GRO only) (Figure 5-3). Select samples were also analyzed for TPHC by GC (USEPA Method 8015, DRO).

Eight TerraProbeSM subsurface soil samples were selected for off-site analysis based upon field analytical results, sampling depth, and exploration location relative to future re-use designation (e.g., gateway or open space/recreation) in order to provide adequate data for the human health risk assessment and characterize contaminant distribution. Eight TerraProbeSM locations were also selected for additional surface soil sampling in order to further characterize the surface contaminant distribution, and to provide adequate data for the risk assessment. Surface and subsurface soil samples selected for off-site analysis were analyzed for PAL VOCs, SVOCs, inorganics, and TPHC.

Groundwater was sampled from each TerraProbe[™] point with a peristaltic pump from inside the probe rods. Prior to sampling, the boring was purged using a peristaltic pump to evacuate the hollow sampling rods. Upon recharge or the removal of two boring volumes a groundwater sample was collected. Groundwater samples were analyzed in the field for BTEX and select chlorinated solvents by GC and TPHC using IR (USEPA Method 8015, GRO only).

Harding Lawson Associates

TerraProbe[™] and surface soil sampling procedures are presented in Subsections 4.5.1.3 and 4.5.1.1, respectively, of Volume I of the Fort Devens POP (ABB-ES, 1995a). Sample results are discussed in detail in Section 7.0 of this report.

5.4.3 Test Pitting

One test pit (ZWE-95-01X) was excavated in the grassy area north of the elementary school in an attempt to identify a geophysical anomaly identified by the geophysical survey (see Figure 5-3). A track mounted backhoe was used to excavate the pit to a depth of 5.5 feet bgs where groundwater was encountered. Three soil samples were collected from the test pit for field screening. Test pitting procedures are outlined in Subsection 4.4.4 of Volume I of the Fort Devens POP (ABB-ES, 1995a). During the test pitting, an HLA geologist described activities and observations in test pit logs that are presented in Appendix A.

5.4.4 Soil Borings and Sampling

Based on the results of previous investigations and TerraProbeSM sampling, six soil borings, ZWB-95-01X, ZWB-95-02X, and ZXM-95-15X through ZWM-95-18X, were completed during the 1995 field investigation (see Figure 5-3 and Table 5-7). ZWB-95-01X and ZWB-95-02X were drilled using 4.25-inch inside diameter (ID) hollowstem augers (HSAs) for the collection of soil samples to augment the TerraProbeSM investigation, while ZWM-95-15X through ZWM-95-18X were drilled using 6.25-inch ID HSAs for the purpose of monitoring well installation. Four additional soil borings were installed as part of the 1996 supplemental field investigation (Figure 5-3). ZWB-96-03B was drilled using 4-inch inside diameter (ID) casing for the collection of soil samples within the courtyard. ZWM-96-19X through ZWM-96-21X were advanced with 4-inch casing for the collection of soil samples and monitoring well installation in each boiler room and within the courtyard. Boring logs are presented in Appendix A.

ZWM-95-15X, ZWB-96-03X, and ZWM-96-19X through ZWM-96-21X were sampled continuously with 2-inch and 3-inch outside diameter (OD) split spoons using the standard penetration test technique to characterize subsurface stratigraphy and vertical distribution of contamination near the suspected source areas. The remaining borings were sampled at approximately 5-foot intervals with an additional sample collected from the water table interval. Two soil samples, one at ground

Harding Lawson Associates

W003971.080

surface and one from the water table, were collected from soil borings ZWB-95-01X and ZWB-95-02X. Soil samples were analyzed for PAL VOCs, PAL SVOCs, PAL inorganics, and TPHC. Two soil samples were collected from each of the borings installed during the 1996 supplemental investigation, one from the water table and another based upon PID measurements and field observations. These samples were analyzed at an off-site laboratory for EPH/VPH and TPHC. A soil sample was collected from the screened interval of all monitoring well borings and analyzed for TOC and grain size distribution.

Soil sample analysis results are discussed in Section 7.0 of this report. Soil borings were completed in accordance with procedures presented in Subsection 4.4.6.1 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

5.4.5 Monitoring Well Installation

Based on the field analytical results of the TerraProbe[™] and soil boring sampling four new monitoring wells (ZXM-95-15X through ZWM-95-18X) were installed during the 1995 phase of the RI (see Figure 5-3 and Table 5-8). Three additional monitoring wells were installed in 1996 (ZWM-96-19X through ZWM-96-21X) to better characterize upgradient and potential source area conditions. All of the monitoring wells are water table wells. Monitoring wells installed in 1995 are constructed of 4-inch ID PVC while monitoring wells installed in 1996 are constructed of 2-inch ID PVC.

Monitoring well construction was completed in accordance with USAEC requirements and Subsection 4.4.6.4 of Volume I of the Fort Devens POP (ABB-ES, 1995a). Monitoring well construction diagrams are provided in Appendix B and a summary of each monitoring well installation is presented in Table 5-8.

5.4.6 Monitoring Well Development

Each of the newly installed RI monitoring wells was developed using the pump and surge method to remove any water added to the boring during drilling and/or well installation, and to remove sediment from the monitoring well screen prior to groundwater sampling. Dedicated equipment was used to minimize the possibility of cross-contamination occurring between wells. The procedures for well development are presented in Subsection 4.4.6.5 of Volume I of the Fort Devens

Harding Lawson Associates

POP (ABB-ES, 1995a). Well development records are provided in Appendix F.

5.4.7 Groundwater Sampling

Four rounds of groundwater samples were collected at AOC 69W. The first two rounds were collected as part of the 1995 field investigation from the four new (ZWM-95-15X through ZWM-95-18X) and six existing monitoring wells (69W--94-09X through 69W-94-14X) (Table 5-1 and Figure 5-3). Round 1 samples were collected in October and November of 1995, and Round 2 samples were collected in February of 1996. Two additional rounds of groundwater sampling were conducted to focus on EPH/VPH parameters, and included the three wells installed in the school (ZWM-96-19X, ZWM-96-20X, and ZWM-96-21X). Round 3 samples were collected in September and October of 1996, and Round 4 samples were collected in December of 1997.

Groundwater samples for the first and second rounds were submitted for laboratory analysis consisting of PAL VOCs, SVOCs, inorganics (both filtered and unfiltered), pesticides/PCBs, water quality parameters, TPHC, and TDS. Groundwater sampling procedures followed the methodology presented in Subsection 4.5.2.2 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

The third round of groundwater sampling was performed in conjunction with the 1996 supplemental investigation from monitoring wells 69W-94-10, 69W-94-11, ZWM-95-16X, and ZWM-96-19X through ZWM-96-21X. The fourth round of groundwater sampling was performed in December of 1997 prior to the soil removal action at AOC 69W (discussed in Subsection 5.5), consistent with the Analytical Approach for AOC 69W Groundwater Sampling (ABB-ES, 1997b). Both these Round 3 and 4 groundwater sampling events were conducted following USEPA Region I low-flow sampling protocols as described in "Low Flow (minimum stress) Purging and Sampling Procedure for the Collection of Ground Water Samples from Monitoring Wells: SOP # GW 0001" (USEPA, 1996). Groundwater samples from both rounds were analyzed for EPH/VPH aliphatic and aromatic ranges, and target PAHs and VOCs (ABB-ES, 1997b). In addition, Round 3 samples were analyzed for target PAHs by Method 8270 and target VOCs by Method 8260.

Harding Lawson Associates

Field data records are presented in Appendix G, and off-site laboratory analytical results are discussed in detail in Section 7.0 of this report.

5.4.8 In-Situ Hydraulic Conductivity Testing and Water Level Measurement

Hydraulic conductivity tests were performed in November of 1995 and January of 1997 on both the newly installed and existing monitoring wells to further define the hydraulic conductivity of the geologic units at AOC 69W. Appendix E presents data and analysis of the hydraulic conductivity testing. Rising head tests were performed on all monitoring wells. The rate of water level recovery to static conditions was monitored with a pressure transducer and electronic data logger. Groundwater head displacement was accomplished with a solid cylindrical PVC slug using the techniques described in Subsection 4.8.2 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

The data from all in-situ hydraulic conductivity tests were analyzed using the methods of Bouwer and Rice (1976) and Hvorslev (1951). Discussion of the in situ hydraulic conductivity testing is presented in Section 6.0 of this RI report. Test data and analyses are provided in Appendix E.

5.4.9 Sediment and Toxicity Test Sampling

Nine sediment samples were collected from six locations in Willow Brook to characterize the distribution of contamination in this downgradient area and for evaluating potential ecological risks (Figure 5-3). One sample was collected at the surface (i.e., from 0 to 6 inches bgs) at all six locations, whereas the other three samples were collected from 2 to 2.5 feet bgs at locations ZWD-95-01X, ZWD-95-02X, and ZWD-95-03X.

Three sediment locations (ZWD-95-04X, ZWD-95-05X, and ZWD-95-01X) were located upgradient of the discharge area for AOC 69W groundwater (closest to the intersection of Antietam Street and MacArthur Avenue) while the other three locations (ZWD-95-02X, ZWD-95-03X, and ZWD-95-06X) were located downgradient of the discharge area for AOC 69W groundwater (closest to Verbeck Gate). All sediment samples were submitted for laboratory analysis consisting of PAL VOCs, SVOCs, metals, pesticides/PCBs, DRO, GRO, TOC, and TPHC (by Method 418.1).

Harding Lawson Associates

W003971.080

In addition, three toxicity test samples were collected concurrently with surface sediment samples ZWD-95-02X, ZWD-95-03X, and ZWD-95-06X, and were sent to Springborn Laboratories, Inc. for subchronic toxicity testing using the midge (*Chironomus tentans*) and amphipod (*Hyalella azteca*). Sediment collected for toxicity testing and chemical analysis were homogenized in the field.

Sediment sampling procedures are presented in Subsection 4.5.3 of Volume I of the Fort Devens POP (ABB-ES, 1995a). The results of the chemical analyses are discussed in detail in Section 7.0, and the results of the toxicity testing are discussed in detail in Subsection 9.2.

5.4.10 Indoor Air Sampling

Indoor air sampling was completed in October of 1997 at the Devens Elementary School to determine if fuel-related contaminants in soil and groundwater beneath the school contribute to ambient contaminant levels that present an unacceptable health risk. Air sampling was conducted in general accordance with procedures set forth in the "Final Remedial Investigation/Feasibility Study Workplan Addendum for Supplemental Air Sampling, Devens Elementary School" (ABB-ES, 1997c).

Ambient conditions in the school were prepared to simulate worst-case exposures likely to be encountered during inactive periods (i.e., weekends). A total of 12 air samples were collected in and around the school in pre-cleaned and leak-tested canisters at the locations illustrated in Figure 5-4. The air flow into each canister was pre-set to allow each canister to fill over an eight hour time period. The weather conditions on the day of the survey were sunny with 5-10 mph winds from the north and temperatures between 50° - 60° F.

Air samples were analyzed following EPA TO-14 Gas Chromatography/Mass Spectrometry (GC/MS) guidelines. Complete laboratory analytical results are discussed in Section 7.0, and a detailed summary of the air sampling effort is provided in Appendix M.

5.4.11 Equipment Decontamination

Several different sampling and analytical procedures were used during the AOC 69W field program, which led to a variety of decontamination procedures. To document

Harding Lawson Associates

the effectiveness of decontamination procedures, periodic equipment rinsate blanks were collected and submitted for chemical analyses. Analytical results for the rinsate blanks are presented in Appendix D. Decontamination procedures followed during the RI are presented in Subsection 4.3 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

One central decontamination pad was constructed and maintained at Building 202 for the decontamination (via steam cleaning) of the drill rig and other equipment, including but not limited to drill rods, well materials, split spoons, augers, drill bits and vehicles. This decontamination pad was approximately 20 feet long and 20 feet wide and was built with three low sides and a sloped floor to collect liquid residuals. The bottom was lined with high density plastic sheeting and plywood sheeting.

5.4.12 Investigation-Derived Waste

During the field program at AOC 69W a variety of investigation-derived waste (IDW) was produced including: purge water, soil cuttings, well development water, decontamination fluids from the decontamination pad, grout, and personnel protective equipment. As the IDW was produced, it was screened in the field with a PID. As soil cuttings were generated from the drilling process they were segregated into piles (5 feet of drilling depth per pile) on plastic sheeting as the cuttings came off the auger fights. The IDW collection, handling, and disposal procedures followed during the RI are presented in Subsection 4.10 of Volume I of the Fort Devens POP (ABB-ES, 1995a).

5.4.13 Location and Elevation Survey

Upon completion of the field program at AOC 69W, a location and elevation survey was conducted to accurately locate the new explorations including: new and existing monitoring wells, soil borings, TerraProbeSM points, test pit, and geophysical grids, and sediment sample locations.

The survey was conducted by Martinage Engineering, Inc. of Reading, Massachusetts. Horizontal control was established with a Leitz Sokkia II Total Station Vernier reading to one second accuracy. Vertical control was established using a Topcon Auto Level accurate to 0.001 foot.

Harding Lawson Associates

Monitoring wells were surveyed for horizontal control, and vertical control of the ground surface, top of the protective casing, and the top of the PVC well riser. Test pits soil borings, TerraProbeSM points, and sediment sample locations were surveyed for horizontal control and vertical control. Survey procedures and accuracy followed during the RI are presented in Subsection 4.9 of Volume I of the Fort Devens POP (ABB-ES, 1995a). Appendix H presents a summary of the survey data for AOC 69W.

5.5 SUMMARY OF REMOVAL ACTION AT AOC 69W

In the winter of 1997 and 1998 (December through March), a removal action was undertaken by Roy F. Weston, Inc. (Weston) to remove approximately 3,500 cubic yards (yd³) of petroleum-contaminated subsurface soil associated with the 1972 fuel oil leak, and to remove the 10,000 gallon UST, the 250 gallon UST, and the associated piping that may serve as a potential conduit to downgradient areas. The removal action encompassed an area approximately 120 by 180 feet immediately north of the elementary school. A narrow "extension" of the removal action reached as far as Willow Brook to the northwest, past the 250 gallon UST. Visual inspection of the 10,000 gallon UST confirmed that it was intact (i.e., no holes or leaks were observed).

Reaching a maximum depth of 13 feet near the school and 8 feet near Willow Brook (in the vicinity of the 250 gallon UST), the excavation intercepted groundwater which ranged from approximately 6.5 to 8 feet bgs at the school, and 3.5 to 5 feet bgs at the 250 gallon UST. No free product was observed in soil or groundwater during the excavation; however, an oily sheen was observed on groundwater emanating from the south sidewall of the excavation (adjacent to the school).

The removal action at AOC 69W removed surface and subsurface soil from several areas that had been investigated as part of the RI. The footprint of the excavation and the soil locations that were removed as part of the soil removal are shown in Figure 5-3 and 5-5. RI soil sample locations that were removed include: ZWB-95-01X, ZWB-95-02X, ZWR-95-27X, ZWR-95-28X, ZWR-95-30X, ZWR-95-31X, ZWR-95-36X, ZWR/ZWS-95-37X, and ZWR/ZWS-95-38X. In addition, the following monitoring wells were also removed: 69W--94-10, 69W--94-11, and 69W--94-13. Physical and chemical data generated for the RI soil samples are discussed in Section

Harding Lawson Associates

7.0 (Nature and Distribution of Detected Site Contaminants); however, these samples were not included in the risk assessment as they no longer represent an exposure to human or ecological receptors.

Excavated soil was field screened using a target goal of 1,000 ppm TPHC by NDIR. All soil removed from AOC 69W was stockpiled at what is now referred to as "Stockpile D" at the Central Soil Storage Facility (Building 202). The stockpile was constructed in a manner so as to direct all groundwater that was removed with soil during the remedial action into an adjacent manhole.

Thirty-one confirmatory soil samples were collected along the walls and floor of the excavation to confirm the NDIR field screening results, and to ensure that remaining soil concentrations were below the MCP Method 1 S-1/GW-1 standards for EPH/VPH. Chemical data generated from these subsurface soil samples are discussed in Section 7.0 and are used in the human health risk assessment. All confirmatory soil samples collected as part of the removal action are shown in Figure 5-5. A complete report of the AOC 69W removal action is presented in Appendix N.

Harding Lawson Associates

6.0 SITE HYDROLOGY, GEOLOGY AND GROUNDWATER CHARACTERIZATION

6.1 SITE HYDROLOGY

AOC 69W is located in the northern portion of the Main Post near Verbeck Gate. The site is comprised of the former Fort Devens Elementary School (Building 215) a parking lot, and adjacent lawn located on the northern side of the school. The site extends from the school to Willow Brook which is located approximately 300 feet to the northwest (Figure 5-1 and 5-2).

Willow Brook originates from Robbins Pond located in the central portion of the Main Post. From Robbins Pond the brook flows north through what was formerly a primarily residential area of the Main Post. Willow Brook continues a northerly route along the eastern side of MacArthur Avenue, past the western side of AOC 69W eventually flowing off the Main Post east of Verbeck Gate. Willow Brook then joins Nonacoicus Brook on the North Post and ultimately discharges to the Nashua River. In addition to Robbins Pond, the brook is fed by surface water runoff and discharge from storm sewer systems 19, 20A, 21, 22, and 23 (ADL, 1994).

In the vicinity of AOC 69W Willow Brook consists of a two- to three-foot deep channel approximately five to six feet wide. The channel is lined with asphalt and concrete rubble. The portion of Willow Brook near AOC 69W was dry during both phases of the AOC 69W RI field efforts (August and September of 1995 and August 1996). Water levels in Willow Brook and the streamside piezometer WBP-93-01 were continuously monitored from November 2 to December 8 of 1993 and April 18 to July 8, 1994 by ETA in support of the North and Main Post detailed flow model. The piezometer and a weir for estimating stream flow were located approximately 1,000 feet downstream (north) of AOC 69W at the culvert crossing under Goodblood Drive adjacent to MacArthur Avenue near the Verbeck Gate. Precipitation, surface water, groundwater, and stream flow data collected by ETA are provided in Appendix E. The base flow for this portion of Willow Brook was estimated at 0.01 cubic feet per second (cfs) for November of 1993 and 0.09 cfs for the first week of December 1993. Estimates for April, May and June of 1994 were 0.33, 0.39, and 0.15 cfs, respectively. Data from the streamside piezometer and the surface water station indicates that groundwater discharges to Willow Brook. Following precipitation

Harding Lawson Associates

events the water level in Willow Brook would rise above surrounding groundwater levels and then equilibrate to levels below groundwater.

Other hydrologic features in the vicinity of AOC 69W include a wetland environment with standing water located on the eastern side of Willow Brook approximately 500 feet north of the school. The wetland continues north along the eastern side of Willow Brook for approximately 300 feet.

6.2 SITE GEOLOGY

This subsection presents descriptions of the geologic formations encountered at AOC 69W. Figure 6-1 shows the orientation of the geologic cross-section and Figure 6-2 presents geologic cross-section A-A'.

6.2.1 Overburden Soils

Surficial and subsurface soils at AOC 69W are classified by the SCS as the Hinckley-Merrimac (Freetown)-Windsor Association (Figure 2-3). The soil is described as being deep; excessively to moderately well drained; nearly level to very steep (see Subsection 2.2.5). Boring logs are presented in Appendix A and results of grain size analysis are provided in Appendix I.

Near surface soils beneath the school and parking lot consist of reworked native soils consisting of dark yellowish-brown fine to coarse sands, gravely sands and silty sands. The soil borings and test pit located in the grassy area to the north of the school show that the surficial soils in this area are comprised of reworked silty sand with gravel and cobbles intermixed with pockets of black organic material. The organic material was found at a maximum depth of 4 feet bgs (69W--94-10, 69W--94-12, and 69W--94-14). These deeper organic layers are believed to be undisturbed native deposits. Excavation of the test pit ZWE-95-01X, located approximately 175 feet north of the parking lot, revealed fill material and debris to a depth of 3 feet bgs.

The predominant soil type at AOC 69W consists of dark yellowish-brown fine to coarse sands, gravely sands and silty sands. Explorations in the vicinity of Willow Brook and its associated wetlands (ZWM-95-15X, ZWP-95-02X, and ZWM-95-18X) revealed a four-to five-foot layer of dark grayish-brown, sandy silt overlying the sands

Harding Lawson Associates

and silty sands. In addition, olive-brown to olive gray fine to coarse sands were observed starting at 9 to 10 feet bgs beneath the school and parking lot (explorations 69W--94-10, ZWB-95-01X, and ZWM-95-16X).

6.2.2 Bedrock Geology

Bedrock was not encountered in any of the borings at AOC 69W; however, bedrock mapping performed in the Devens area indicates that AOC 69W is located near an inferred contact between the Devens Long Pond facies of the Ayer granite and the Oakdale Formation.

The Devens-Long Pond facies is described as being a gneissic, equigranular to porphyroblastic biotite granite and granodiorite (Wones and Goldsmith, 1991). The most prominent example of the Devens-Long Pond facies is the outcrop which forms Shepley's Hill located to the west of AOC 69W.

The Oakdale Formation is classified as fine-grained metasiltstone and phyllite, comprised of quartz and minor feldspar and ankerite. The Oakdale Formation is commonly deformed by kink banding (Zen, 1983; Peck, 1975; Russell and Allmendinger, 1975).

Depth to bedrock at AOC 69W is unknown. Bedrock borings in the vicinity of Shepley's Hill Landfill, located approximately 2,500 feet to the northeast of AOC 69W, indicate that depth to bedrock in this area can be highly variant.

6.2.3 Site Geology Interpretation Summary

The geologic conceptual model developed for AOC 69W suggests that the existing wetlands to the east of Willow Brook were historically more widespread. Organic layers observed in soil borings indicates that the wetlands at one point extended to the southeast as far as the current UST location. Soil boring data further suggests that the presumed historic wetlands in the vicinity of the parking lot were filled in by natural processes and more recently the wetlands which existed in the grassy area to the north of the parking lot were filled during some phase of Fort Devens construction. Soils surrounding and underlying Willow Brook consist of dark grayishbrown sandy silt 4 to 5 feet in thickness. The most predominant soil type at AOC 69W consists of dark yellowish-brown fine to coarse sands, gravely sands and silty

Harding Lawson Associates

sands. These soils underlie the wetland deposits and appear to be reworked in the near surface beneath the parking lot and school. In addition, these native sands appear to have been the fill material used to create the grassy area north of the parking lot.

6.3 HYDROGEOLOGY

This subsection presents data and interpretations of hydrogeologic conditions at AOC 69W. Groundwater levels used in this subsection are provided in Table 6-1 and interpretive water table elevation contours are presented in Figure 6-3. Water level elevations at AOC 69W were measured on December 7, 1995, March 26, 1996, July 23, 1996, and January 15, 1997. In addition, water level information obtained during the low-flow groundwater sampling event in December of 1997 are also included. In-situ hydraulic conductivity results are provided in Table 6-2 and Appendix E.

As a result of the type of contaminants (primarily fuel related compounds) identified at AOC 69W all of the monitoring wells have been installed as water table wells (i.e., their screened interval and sandpack span the water table) (see Figure 6-2 and Table 5-8). A total of 13 monitoring wells, 69W--94-09, 69W--94-10, 69W--94-11, 69W--94-12, 69W--94-13, 69W--94-14, ZWM-95-15X, ZWM-95-16X, ZWM-95-17X, ZWM-95-18X, ZWM-96-19X, ZWM-96-20X, and ZWM-96-21X were installed in overburden soils. In addition, two piezometers ZWP-95-01X and ZWP-95-02X were installed adjacent to Willow Brook.

The water table occurs in the overburden across AOC 69W (Figure 6-2). Figure 6-3 presents interpreted water table elevation contours for AOC 69W based on the January 15, 1997 data set. Groundwater flow is predominately south-southeast to north-northwest toward Willow Brook. Depth to groundwater at AOC 69W ranges from approximately 4 to 5 feet bgs on the north side of the school building to approximately 1 foot bgs adjacent to Willow Brook. The January 15, 1997 data set indicates that groundwater was discharging to Willow Brook. It was not possible to verify this at the time as the brook was frozen and snow covered. Groundwater discharge to Willow Brook would be consistent with the fact that the January 1997 water levels were the highest recorded for AOC 69W.

Harding Lawson Associates

It was not possible to calculate vertical hydraulic gradients at AOC 69W as there are no deep overburden wells; however, data collected from the streamside piezometer and gaging station show groundwater discharges to Willow Brook indicating upward vertical gradients (see Subsection 6.1, Hydrology).

Horizontal hydraulic gradients were calculated from each set of water level measurements. Gradients were calculated using multiple wells that, as much as possible, share a common flow path. In general, horizontal hydraulic gradients are relatively consistent across AOC 69W. Calculations are provided in Appendix E. The geometric mean of horizontal hydraulic gradients calculated for all data sets range between 0.006 ft/ft (December 7, 1995) and 0.008 ft/ft (January 15, 1997).

In-situ Hydraulic Conductivity Results. In-situ hydraulic conductivity test results presented in Table 6-2 indicate that estimates of hydraulic conductivity as calculated by the Bouwer and Rice method range between 1.3×10^{-3} cm/sec (3.7 ft/day) at 69W--94-09 and 3×10^{-2} cm/sec (85 ft/day) at 69W--94-14 and ZWM-95-15. The geometric mean of the monitoring wells hydraulic conductivities was calculated as 1.6 $\times 10^{-2}$ cm/sec (45 ft/day). All of the hydraulic conductivity test data and results are presented in Appendix E.

Groundwater Velocity Analyses. Flow velocities were estimated for AOC 69W using maximum, minimum, and mean horizontal hydraulic gradients and hydraulic conductivities as determined by the Bouwer and Rice method (calculations are provided in Appendix E). An overburden effective porosity of 30 percent was assumed for the predominately sandy soils. The maximum groundwater flow velocity was estimated at 2 feet per day (ft/day). A minimum flow velocity of 0.07 ft/day was calculated for the water table. The geometric means of observed hydraulic conductivities and hydraulic gradients yielded a flow velocity of 0.7 ft/day.

6.3.1 Site Hydrogeology Interpretation Summary

Groundwater at AOC 69W occurs in the overburden and bedrock aquifers. Bedrock aquifer characteristics were not monitored at AOC 69W. Flow directions are predominately south-southeast to north-northwest. Groundwater discharges to Willow Brook in the vicinity of AOC 69W at times of high groundwater levels. Due to surface water runoff and the storm sewer system, water levels in Willow Brook may be higher than groundwater levels following precipitation events. AOC 69W is

Harding Lawson Associates

located within the draft delineated Zone 2 for MacPherson Production well located approximately 3,000 feet to the north (ETA, 1996).

1

Harding Lawson Associates

W003971.080

7.0 NATURE AND DISTRIBUTION OF DETECTED SITE CONTAMINANTS

The following subsections address the nature and distribution of analytes detected in soil, groundwater, sediment, and air collected from AOC 69W during the RI. Supplemental groundwater and air sample data collected in 1997 are also presented. Data obtained from the off-site laboratory and from the on-site field analytical laboratory are presented in this section.

During implementation of the RI field program, field analytical results were used to direct placement of soil borings and monitoring wells, and were used to define the vertical and/or horizontal distribution of contaminants. Field analytical results were also used to select samples for off-site laboratory analysis. Samples were collected from hot zones to gather information on the nature and concentrations of contaminants, and samples were collected from clean areas for off-site confirmation. Field analytical data were used to supplement the off-site laboratory analytical data in the assessment of the nature and distribution of detected analytes.

This assessment of site-related contaminants relies upon tables to present the field and off-site laboratory analytical data. The tables contain only detected analytes and concentrations for samples within a given media. The text within the assessment provides detail, interpretation, and analysis of the tabulated data. A complete summary of the field and off-site analytical data is presented in Appendix L.

7.1 APPROACH TO CONTAMINATION ASSESSMENT

Off-site laboratory analytical results and field analytical data are the primary data used to assess impacts at the site from suspected past disposal and storage practices. A complete data set of field and off-site analytical data, including non-detect results, is presented in Appendix L. Tentatively identified compounds (TICs) reported for off-site laboratory data are discussed in Subsection 7.1.1 and are presented in Table 7-1.

Analytes detected in QC blanks analyzed at the off-site laboratory are presented in Subsection 7.1.2. A blank contamination evaluation was performed with this data to identify probable sampling and off-site laboratory-related contaminants. The

Harding Lawson Associates

contamination assessment included determining uncertainty regarding potential false positive results due to sampling and off-site laboratory contaminants. With the exception of the 1997 Supplemental Air and EPH/VPH sample data, data presented in the tables were not qualified or corrected for blank contamination. However, based on the blank contamination assessment performed, a "*" flag has been added to the data when applicable, to indicate probable blank contamination. A detailed review of laboratory method blank and field quality control blank analyses from the RI program is presented in Section 3.0 and in the DQR for the 1995, 1996, and 1997 investigations are located in Appendix D. The DQR include evaluations for data collected from AOCs 57, 63AX and 69W because the investigations were conducted simultaneously.

An evaluation of analytical data precision and accuracy was conducted using results of field duplicate and MS/MSD analyses. Accuracy of VOC and SVOC results were also evaluated using surrogate recovery data from each sample analysis. The results for some analytes have been identified as estimated based on the field duplicate, surrogate, and/or MS/MSD data. In some cases, possible data biases have been identified. A summary of data usability interpretations is contained in Subsection 7.1.2. Detailed discussions of surrogate, field duplicate, and MS/MSD results are presented in the DQRs in Appendix D.

7.1.1 Tentatively Identified Compounds/Non-Project Analyte List Compounds

During off-site laboratory analysis, non-project analyte list compounds present in VOC and SVOC samples were tentatively identified by comparing the GC/MS spectra to those contained in the National Bureau of Standards mass spectral library. Once the tentative identification was made based on matching spectra, the appropriate USAEC code name was assigned for that compound.

Reported concentrations of TICs are considered estimated and are not based on calibration standards. If no compound identification was possible, the compound became listed as an unknown with an assigned number. The assigned number which accompanies the prefix "UNK" is determined by the relative retention time to the internal standard. For example, if the relative retention time of the compound compared to 1,4-difluorobenzene is 1.42, the compound would be assigned the number "UNK142" in IRDMIS.

Harding Lawson Associates

W003971.080

The requirements for making tentative identification of compounds are listed in the Fort Devens POP (ABB-ES, 1995a) as follows:

- 1. Relative intensities of major ions in the reference spectrum (ions > 10 percent of the most abundant ion) should be present in the sample spectrum.
- 2. The relative intensities of the major ions must agree within 20 percent.
- Molecular ions present in the reference spectrum should be present in the sample spectrum.
- 4. Ions present in the sample spectrum but not in the reference spectrum should be reviewed for possible background contamination or presence of co-eluting compounds.
- 5. Ions present in the reference spectrum but not in the sample spectrum should be reviewed for possible subtraction from the sample spectrum because of background contamination or co-eluting compounds. Data system library reduction programs can sometimes create these discrepancies.
 - 6. If in the technical judgment of the mass spectral interpretation specialist, no valid tentative identification can be made, the compound should be reported as unknown.

7.1.1.1 TICs Detected in Samples from AOC 69W. VOC and SVOC TICs and unknown compounds were detected in several soil and groundwater samples collected from AOC 69W. These compounds are differentiated from target analytes in the USAEC's IRDMIS with an "S" flag in the flagging code field. All TICs associated with samples from AOC 69W collected during the 1995 RI investigations and from the Round 2 Groundwater Sampling Event are summarized below and presented in Tables 7-1. TICs were not reported in any of the samples collected during the 1996 investigation.

It is important to note that in addition to the GC/MS method used to identify and report the alkanes and aromatics identified as TICs, USEPA Methods 418.1, 9071 and 8015 were used during the off-site analysis of soil and water samples to quantify

Harding Lawson Associates

W003971.080

and classify hydrocarbons within these chemical classes. Supplemental data was also collected using Massachusetts EPH and VPH methods to further characterize the non-target hydrocarbons. Field analysis was conducted on many samples during the RI using a modified version of USEPA Method 418.1. The field analysis method was designed to provide data on the distribution of these fuel hydrocarbons. Field analytical results were used to direct field exploration programs and provide supporting data for the off-site sample results. The off-site laboratory USEPA Method 418.1 results are the primary data used to make quantitative evaluations of these chemicals as TPHC.

<u>1995 RI TICs</u>. TICs detected in groundwater, soil and sediment samples collected during the 1996 RI consisted primarily of alkenes, alkanes, alkyl-substituted-alkanes, and alkyl-substituted benzenes, toluenes and naphthalenes.

The field samples with the highest concentrations and the most frequent detection of these TICs include soil samples BXZW0107, RXZW2607, RXZW3006, RXZW3607, RXZW3704, groundwater samples MXZW10X4, MXZW13X3, MXZW13X4, MXZW16X1. The concentration of TICs detected in these samples ranged from 4 μ g/L to 5,000 μ g/L in groundwater samples, and from approximately 0.0052 μ g/g to 300 μ g/g in soil samples. The presence of alkanes and alkylsubstituted compounds in these samples may be indicative of gasoline, fuel oil, or other fuel related contamination.

Other compounds detected in samples collected from AOC 69W include alpha-pinene, beta-pinene, and gamma-sitosterol. These are interpreted to be related to natural organics and not indicative of waste contamination.

Samples also contained unknown VOCs and SVOCs ranging from 0.03 μ g/g to 50 μ g/g in sediment samples, 0.006 μ g/g to and 10,000 μ g/L to 300,000 μ g/L in groundwater samples.

Harding Lawson Associates

7.1.2 Potential Laboratory and Sampling Contaminants

An evaluation of results from rinse, trip, and laboratory method blank analyses was conducted to determine possible contaminant contributions originating from non-siterelated sources. Potential sources of contamination include materials used during borehole advancement and monitoring well installation, field sampling procedures, field equipment decontamination, sample shipment, laboratory storage, and laboratory analysis.

Because the majority of off-site analytical data was generated using USAEC methods, USEPA data validation guidelines related to the evaluation of blank contamination were not implemented. The following blank contamination assessment approach for organics was used regarding laboratory method blank and field QC sample blank contamination:

- 1.) Non-target VOCs and SVOCs TICs that are common organic laboratory contaminants (USEPA, 1988) are not considered chemicals of concern. These common organic laboratory contaminants include:
 - Siloxanes; diethyl ether; 1,1,2-trichloro-1,2,2-trifluoroethane; fluorotrichloromethane; and phthalates at levels less than 100 μ g/L or 4 μ g/g in samples collected during the 1995 Field Investigation and the Round 2 Groundwater sampling event.
 - trifluorochloromethane at levels less than 0.1 μ g/g in samples collected during the 1996 Field Investigation.
 - Solvent preservatives such as cyclohexane, and related by-products including cyclohexene, cyclohexanone, cyclohexenone, cyclohexanol, cyclohexenol, chlorocyclohexene, and chlorohexanol.
 - Aldol condensation products of acetone including 4-hydroxy-4-methyl-2-pentanone, 4-methyl-2-penten-2-one, and 5-dimethyl-2(5H)-furanone.
- The additional TICs detected in SVOC blanks include heptacosane, nonacosane and dioctyladipate. The VOC TICS detected hexane and ethanol.

Harding Lawson Associates

7-5

W003971.080

The investigations during which these TICs were detected are shown in Table 7-1.

- 4) For organic target compounds, trends in method and field blanks were evaluated. Several target compounds routinely detected have been identified by USEPA as common laboratory contaminants including:
 - phthalates
 - methylene chloride, acetone, toluene, and methyl ethyl ketone (2butanone) at concentrations comparable to concentrations observed in blanks.
- 5) The pesticides malathion detected in method blanks at 0.188 μ g/L and alphaand gamma-chlordane in method blanks at concentrations up to 0.01 μ g/g.

Organic target analytes detected in method blanks and rinse blanks during the 1995 and 1996 Field investigations and the 1995 Round 2 Groundwater Sampling Event are presented in Table 7-2 and 7-3. VOCs detected in trip blank samples collected during the 1995 Field Investigations and the Round 2 Groundwater Sampling Event are presented in Table 7-4. Trip blanks analyzed during the 1996 investigation did not have any detections of VOCs reported. Organic compounds detected in samples at similar concentration ranges as those in blanks are discussed qualitatively in the contamination assessment, and carried through the risk assessment calculations. In addition, the spatial distribution and relative concentration of common organic laboratory contaminants, and rationale for elimination of data from consideration as site-related contaminants, are presented in the contamination assessment in Subsection 7.2.

Inorganic elements were not reported in rinse blanks and water method blanks. Inorganic detections in the soil method blanks are not presented because the source of elements is believed to be the blank soil matrix rather than laboratory contamination (see Appendix D-1 and D-3, Section D-2.0). Inorganic sample data presented in the data tables and risk assessment tables were not revised based on blank contamination results. All inorganic detections were used for risk assessment calculations.

Harding Lawson Associates

During the RI, samples were analyzed for a variety of water quality parameters to generate data to support the development of alternatives during the FS process. No rinse blank or method blank contamination was reported for the water quality parameters analyzed.

A detailed discussion of laboratory QC sample results is presented in Appendix D for each field program.

1997 Supplemental Groundwater and Air Sampling

Groundwater and air sample data were collected on fuel hydrocarbons (EPH/VPH), volatiles, and semivolatiles. Results were validated for blank contamination as outlined in USEPA guidelines (USEPA, 1988; USEPA, 1996), and results were qualified prior to use in the RI. Validation reports are presented in Appendix D. Blank results indicate no contamination was observed for groundwater sample analyses. Acetone, toluene, and xylene contamination was observed at low concentrations in blanks associated with the air analyses. The majority of the reported acetone results were qualified as non-detect U during validation, and it is possible that positive detections of acetone are related to sampling or laboratory contamination.

7.1.3 Analytical Data Accuracy and Precision

Analytical data accuracy and precision was evaluated using MS and field duplicate analyses for the majority of off-site analytical methods. Surrogate recoveries were also reviewed to evaluate the accuracy of VOC and SVOC measurements. This evaluation was conducted to support the AOC 69W 1995 and 1996 RI field programs and the AOC 69W 1996 Round 2 Groundwater Sampling Event. Detailed discussions and presentation of these results are included for the 1995 and 1996 investigations in Appendix D.

Matrix spike, field duplicate, and surrogate results for the majority of the and target analytes evaluated during the RI indicate the accuracy and precision of results were within project goals outlined in the Fort Devens POP (ABB-ES, 1995a) and USEPA control limits (USEPA, 1988; USEPA, 1989). Trends were reviewed for each set of QC sample data from each field event to determine if qualification of the accuracy of results was needed. The results for some analytes in AOC 69W samples have

Harding Lawson Associates

been identified as invalid or as estimated values with potential biases noted. The following items summarize the qualification of results:

AOC 69W 1995 RI

- 1. Based on spike recovery data discussed in Appendix D-1, Subsection D.3.1.1, positive detections of selenium in soil are considered estimated with no particular low or high bias.
- 2. Results for MS and MSDs discussed in Appendix D-1, Subsection D.3.1.1 indicate lead results for soil analyzed by GFAA are estimated, and results may be biased low.
- 3. High frequency of MS/MSD recoveries above the upper control limits indicate that there may be some matrix interference (Appendix D-1, Subsection D.3.1.1). Positive results reported for arsenic in soil samples should be considered estimated and potentially biased high.
- 4. Due to the low MS recovery in sample DXZW0200 (Appendix D-1, Subsection D.3.1.1), positive results for arsenic in sediment samples from AOC 69W should be considered biased low, and non-detect results are unusable.
- 5. Due to the low MS recovery in sample DXZW0200 (Appendix D-1, Subsection D.3.1.1), positive results for manganese in sediment samples from AOC 69W should be considered biased low, and non-detect results should be considered unusable.
- 6. Based on spike recoveries for hardness (Appendix D-1, Subsection 3.3.3), all hardness results for groundwater samples should be considered invalid.
- 7. VOC surrogate recovery evaluations are presented in Appendix D-1, Subsection D.3.2.1:
 - Groundwater sample MXZW12X3 from AOC 67W had high surrogate recoveries for 1,2-dichlorobenzene-D4. Positive

Harding Lawson Associates

W003971.080

results for ethylbenzene, tetrachloroethene, toluene, xylenes, and chloromethane results are considered estimated and potentially biased high.

- The recoveries of surrogate standards toluene-D8 and 4bromofluorobenzene in soil sample BXZW0107 from AOC 69W were high. Positive concentrations of ethylbenzene and xylenes in this sample are estimated and possibly biased high.
- The surrogate recovery for toluene-D8 in one sediment sample from AOC 69W (RXZW3006) also exceeded the upper control limit. Positive results for toluene, xylene, and styrene should be considered estimated, and potentially biased high.
- 8. TPH results reported for all AOC 69W sediment samples should be considered estimated values based on duplicate RPD results presented in Subsection D.4.1.4 in Appendix D-1.
- 9. As discussed in Subsection D.4.2.1 in Appendix D-1, iron results reported in groundwater samples from AOC 69W are considered estimated based on outlier RPDs between spike sample results.
- 10. Positive detections of heptachlor epoxide and gamma-chlordane in AOC 69W groundwater sample MXZW10X4 are considered estimated based on RPD exceedances between spiked sample results (Subsection D.4.2.2 in Appendix D-1).

Groundwater, Round 2

- 1. Phosphate results from AOC 69W groundwater samples are considered estimated values based on outlier RPDs between field duplicate results.
- 2. VOC surrogate recovery evaluations are presented in Appendix D-1, Subsection D.3.2.1:

Harding Lawson Associates

- Based on high surrogate recoveries of 1,2-dichlorobenzene-D4 for AOC 69W groundwater samples MXZW14X4, MXZW15X2, and MXZW18X2 positive results for VOCs in these samples should be considered estimated, and biased high. However, only low concentrations of toluene (< $1.2 \mu g/L$) were reported in all samples and high surrogate recoveries are not interpreted to impact data usability.
- Ethyl benzene results reported in AOC 69W groundwater sample MXZW10X4 is estimated and potentially biased high based on high surrogate recoveries of 1,2-dichlorobenzene-D4.

AOC 69W Fall 1996 RI

- 1. Based on MS recoveries discussed in Appendix D-3, Subsection D.3.1.1, positive detections, and CRLs for non-detect results for mercury, arsenic, and manganese in soil are considered estimated values and potentially biased low.
- 2. Based on MS recoveries discussed in Appendix D-3, Subsection D.3.1.2, lindane CRL results in AOC 69W groundwater samples may be biased low and should be considered estimated. Lindane was not detected in groundwater samples.

1997 Supplemental Groundwater and Air Samples

A summary of data validation and data qualification actions is contained in Appendix D-4. Based on data validation reviews all groundwater results are usable as reported. The following should be considered when using the air sample results:

• Results for sample ZWA-97-04X have been rejected and are considered unusable due to loss of canister negative pressure during sample collection.

Harding Lawson Associates

7.2 INVESTIGATIONS AT AOC 69W

7.2.1 1994 ADL AREE 69W Investigation

The following subsection details the analytical findings of the AREE 69W investigation conducted in 1994 by ADL.

In July of 1993 ADL investigated the Fort Devens Elementary School, designated AREE 69W, as part of the basewide AREE 69 (Past Spill Sites) investigation. The investigation focused on the 1978 fuel oil release and was comprised of a document review and site visit. The study concluded that there was a potential for fuel oil contamination in the soil and groundwater (ADL, 1995).

Further investigation was performed at AREE 69W from March through June of 1994. The investigation involved sampling, field screening, and laboratory analysis of surface soil, subsurface soil, groundwater, surface water, sediment, and a geophysical survey to locate subsurface utilities.

7.2.1.1 Surface Soil. Six surface soil samples were collected from a depth of 0 to 1 foot bgs, from the grassy area north and northwest of the parking lot (Figure 5-2). The samples were analyzed by portable field instruments for BTEX and TPHC, using GC and NDIR techniques, respectively. The sample with the highest observed TPHC concentration was submitted for laboratory analysis of PAL VOCs, PAL SVOCs, TPHC, PAL inorganics, and TOC. Field and off-site soil analytical data are provided in Tables 7-5 and 7-6, respectively.

TPHC field screening concentrations ranged from 9.5 ppm to a high of 131 ppm (observed at location HA-5, located just off the northwest corner of the paved area). No BTEX were detected in the screening results. Laboratory results from the single surface soil sample submitted for laboratory analysis revealed no compounds at concentrations exceeding MCP Method 1, S-1/GW-1 Standards. cPAHs detected in the surface soils at the site consisted of benzo(a)anthracene and chrysene at a combined concentration of 0.29 μ g/g.

7.2.1.2 Subsurface Soil. Subsurface soil samples were collected during the installation of groundwater monitoring wells and during the Geoprobe investigations. During the first round of Geoprobe sampling, subsurface samples were collected

Harding Lawson Associates

from 3 to 5 feet bgs at 16 locations (Figure 5-2) for field analysis of TPHC and BTEX. Of the 32 samples analyzed in the field, three samples exhibiting the highest TPHC concentrations and one sample with the lowest TPHC concentration were submitted for laboratory analysis of PAL VOC, PAL SVOCs, TPHC, PAL inorganics and TOC analysis. During the second Geoprobe sampling round, nine additional locations were investigated (Figure 5-2). Subsurface soil samples were collected from a depth of 3 to 5 feet bgs and field screened for TPHC. Field and off-site analytical soils data are provided in Tables 7-5 and 7-6, respectively.

Soil samples were collected at depth intervals of 2 to 4 feet, and 11 to 13 feet bgs during the monitoring well installation effort. These samples were screened in the field for TPHC and BTEX. The samples from the 2 to 4 and 4 to 6 foot depth intervals were submitted for laboratory analysis of TPHC, PAL VOCs, PAL SVOCs, PAL inorganics, and TOC analysis.

TPHC concentrations in soils collected with the Geoprobe and from monitoring well soil borings ranged from 7.5 ppm to 15,500 ppm (GP-2 at 3 to 5 feet bgs). Concentrations of benzo(b)fluoranthene, benzo(a)anthracene, and chrysene detected in Geoprobe soils were above MCP Method 1, S-1/GW-1 standards.

Based on the field screening and laboratory analysis results, TPHC and cPAH soil contamination appears to have been concentrated in the area of the existing UST (the presumed source area), and may have migrated downgradient towards Willow Brook.

7.2.1.3 Groundwater. Groundwater samples were collected from each Geoprobe location and from the six newly installed groundwater monitoring wells. Sixteen groundwater samples were collected during the first Geoprobe sampling round and field screened for TPHC and BTEX (GP-01 through GP-16). Filtered and non-filtered groundwater samples collected during the second Geoprobe sampling round were field screened for TPHC (GP-17 through GP-25). Sample locations are shown in Figure 5-2. Field and off-site analytical results from groundwater samples are provided in Tables 7-5 and 7-7, respectively.

Field screening results from the 25 Geoprobe groundwater samples indicated that TPHC was present in groundwater. BTEX was not detected. Five sample locations from the first Geoprobe sampling round exhibiting the highest field screening TPHC

Harding Lawson Associates

concentrations were resampled and submitted to the laboratory for analysis of PAL VOCs, PAL SVOCs, TPHC and water quality parameters. No samples from the second geoprobe sampling round were sent for laboratory analysis. Results indicated that TPHC, inorganic analytes (arsenic, lead, antimony, beryllium, chromium, and nickel), and organic compounds (1,1-dichloroethene, benzene, carbon tetrachloride, chloroform, tetrachloroethene, trichloroethene, 2-methyl naphthalene, and naphthalene) were detected at concentrations exceeding MCP Method 1, GW-1 Standards. Most of these exceedances occurred at locations GP-1, GP-2, GP-6 and GP-15, in the area of the UST and downgradient of this location. No cPAHs were detected in the Geoprobe groundwater samples.

Six monitoring wells installed at the site confirmed the results of the Geoprobe investigation. Groundwater samples were submitted for analysis of TPHC, PAL VOCs, PAL SVOCs, unfiltered inorganics and water quality parameters. Results indicated that TPHC, arsenic, beryllium, cadmium, chromium, lead, nickel, 2-methyl naphthalene, acenaphthene, and naphthalene were detected at concentrations exceeding MCP Method 1, GW-1 Standards. These exceedances occurred at monitoring wells 69W-94-10, 69W-94-11, 69W-94-13 and 69W-94-14. No cPAHs were detected in the groundwater samples.

Groundwater sample results indicate that the area around the UST has the greatest number of compounds exceeding MCP Standards. Groundwater northwest of the UST was also found to have elevated concentrations of inorganics and TPHC, suggesting that contaminants have potentially migrated downgradient of the UST location.

7.2.1.4 Surface Water and Sediment. Surface water and sediment samples were collected from two locations in Willow Brook (Figure 5-2). One sample location (69W-94-16) was placed in line with the inferred plume migration pathway indicated by the Geoprobe survey, and the other (69W-94-15) was placed upstream of this area. Samples were analyzed for TPHC, PAL VOCs, PAL SVOCs, unfiltered inorganics, and water quality parameters. Surface water and sediment off-site analytical results are provided in Tables 7-8 and 7-9.

The results indicated the presence of cPAHs in both sediment samples, and TPHC in sample 69W-94-16. Specifically, the cPAHs benzo(a)anthracene, chrysene, benzo(b)fluoranthene, and benzo(k)fluoranthene were detected in the 69W-94-15

Harding Lawson Associates

(upstream) sediment sample. In sediment sample 69W-94-16, TPHC, benzo(a) anthracene and chrysene were detected. Total cPAHs in the upstream sample barely exceeded 7.0 ppm. Total cPAHs in the downstream sample were an order of magnitude less than the clean-up values. Other PAHs and metals were detected in both samples.

TPHC and cPAHs were not detected in surface water samples.

7.2.2 AOC 69W RI and Removal Action Soils Results

The following subsections present field and off-site laboratory analytical soil results for samples collected at AOC 69W during the RI and during the soil removal action. Field analytical soil data from the RI are presented in Table 7-10. Off-site laboratory analytical soil data from the RI are presented in a hits-only format in Tables 7-11 and 7-12. Off-site laboratory analytical soil data from the remedial action are presented in Table 7-13. Complete field analytical and off-site laboratory analytical soil data including non-detects are presented in Appendix L.

7.2.2.1 **RI Field Analytical Soil Results**. Soil samples were collected for field analysis from TerraProbeSM points, soil borings, and test pits. The field analytical samples were collected in an attempt to define the nature and distribution of the site-related contaminants as well as to delineate potential contaminant source areas. A discussion of the results is presented below.

1995 RI TerraProbeSM and Test Pit Soil Sampling Field Analytical Results. In September of 1995, a total of 92 soil samples were collected from 29 TerraProbeSM points (ZWR-95-26X through ZWR-95-28X, and ZWR-95-30X through ZWR-95-55X) for on-site analysis of BTEX, select VOCs, GRO, and TPHC. Select samples were also analyzed for DRO. Soil samples were generally collected from near ground surface, a midpoint, and the water table as exploration conditions allowed. On-site analytical results for the TerraProbeSM soils are provided in Table 7-10.

VOCs were detected in twelve of the TerraProbesM soil samples. The majority of the detections were the petroleum related compounds toluene, chlorobenzene, ethylbenzene and xylenes. These detections can be divided into two groups, samples from 6 to 10 feet from within the parking lot adjacent to the school (ZWR-95-26X)

Harding Lawson Associates

at 7 feet, ZWR-95-28X at 10 feet, ZWR-95-30X at 6 feet, and ZWR-95-31X at 7 feet) and surficial samples located in the grassy area north of the parking lot (ZWR-95-44X at 0 feet, ZWR-95-45X at 1 foot bgs, and ZWR-95-46X at 0 feet). The highest detected total and individual concentrations of these analytes were found immediately adjacent to the school in the samples from ZWR-95-28X (6,600 μ g/kg at 10 feet bgs) and ZWR-95-30X (6,817 μ g/kg at 6 feet bgs).

Figures 7-1 and 7-2 provide field analytical results for soils from 0 to 5 feet bgs and 6 to 10 feet bgs, respectively. Other detected VOCs include chloroform at 3.8 μ g/kg and 2.5 μ g/kg in the 4 feet bgs samples from ZWR-95-33X and ZWR-95-47X, respectively. Carbon tetrachloride was found at 2.4 μ g/kg in the 10 feet bgs sample from ZWR-95-30X and TCE was detected at 2.5 μ g/kg in the 10 feet bgs sample from ZWR-95-38X.

TPHCs were detected in 42 of the TerraProbeSM soil samples with a maximum detected concentration of 7,700 μ g/g at 6 feet bgs in ZWR-95-30X. The most significant detections (i.e., in excess of 500 $\mu g/g$) were located in the area of the underground "skimmer system" which leads from the school to the 250 gallon underground holding tank located approximately 300 feet to the northwest. The "skimmer system" that was installed in 1972 appears to have acted as a conduit for migration of TPHC soil contamination. Figures 7-1 and 7-2 provide field analytical TPHC data. As is shown in Figure 7-3 (including 1996 field analytical data), there appear to be two distinct areas of TPHC contamination in AOC 69W subsurface soils. The highest concentrations were observed in the area stretching from the new boiler room located inside the school building to the 250 gallon UST in the wooded area to the northwest. TPHCs were also detected in subsurface soils at 4 to 7 feet bgs in the TerraProbeSM borings ZWR-95-26X and ZWR-95-33X located adjacent to the school building outside of the old boiler room. TPHCs in this area are attributed to the 1978 fuel oil release and appear to be localized to this area and distinct from the TPHC concentrations to the west.

1996 Monitoring Well Boring and Soil Boring Sample Field Analytical Results. In 1996, three monitoring wells (ZWM-96-19X, ZWM-96-20X, and ZWM-96-21X) and one soil boring (ZWB-96-03X) were installed within the Elementary School boiler rooms and courtyard to define the upgradient extent of the petroleum related contamination observed during the 1995 RI investigation. All of the borings were sampled continuously to characterize the vertical distribution of contaminants near

Harding Lawson Associates

W003971.080

SECTION 7

the suspected source area. A total of twenty-eight soil samples, including duplicates, were collected from the four explorations and analyzed in the field for select VOCs, SVOCs, and TPHC (Table 7-10). Figure 7-4 provides results of the field analysis in a hits only format. The majority of detected contaminants were observed in the 9 to 13 feet bgs samples from the monitoring well boring ZWM-96-19X. Only one VOC was detected, $580 \ \mu g/kg$ of m/p xylene at 9 feet bgs in ZWM-96-19X. Naphthalene, the only SVOC compound detected, was found in four of the samples from ZWM-96-19X: $680 \ \mu g/kg$ at 3 feet bgs, $3,800 \ \mu g/kg$ at 9 feet bgs, $4,500 \ \mu g/kg$ at 11 feet bgs, and $1,400 \ \mu g/kg$ at 13 feet bgs. TPHC was the only compound detected in the other explorations. The maximum observed concentrations were found in the 9 and 11 feet bgs samples from ZWM-96-19X at 840 \ \mu g/g and 790 \ \mu g/g, respectively. The only TPHC detections in the other explorations were 62 \ \mu g/g at 2 feet bgs in ZWM-96-20X and 57 \ \mu g/g at 8 feet bgs in ZWM-96-21X.

7.2.2.2 RI Off-Site Analytical Soils Results. Subsequent to reviewing field analytical results for soil, samples were submitted for laboratory analysis from TerraProbeSM points, soil borings, and surface soil locations. Samples were collected to further define the lateral and vertical extent of contamination and to provide off-site confirmation of the on-site analysis. A discussion of all the results is presented below. As discussed in Subsection 5.5, a removal action was implemented to eliminate the source area and the associated piping which may have acted as a potential conduit to downgradient areas. Therefore, some of the chemical concentrations discussed in the following paragraphs are no longer present at the site (they are noted accordingly).

1995 Soil Boring, TerraProbeSM, and Surface Soil Sample Analytical Results. In the fall of 1995 a total of 28 soil samples were collected from the following locations:

- 2 soil and 3 monitoring well borings (ZWB-95-01X, ZWB-95-02X, ZWM-95-15X, ZWM-95-16X, and ZWM-95-18X),
- 8 TerraProbesM points (ZWR-95-26X, ZWR-95-30X, ZWR-95-35X, ZWR-95-36X, ZWR-95-37X, ZWR-95-38X, ZWR-95-39X, and ZWR-95-45X,
- 8 surface soil locations (ZWS-95-35X, ZWS-95-37X, ZWS-95-38X, ZWS-95-39X, ZWS-95-42X, ZWS-95-45X, ZWS-95-46X, and ZWS-95-

Harding Lawson Associates

W003971.080

47X (coincident with TerraProbeSM locations).

Soil analyses included PAL VOCs, SVOCs, inorganics, TPHC, and TOC. RI exploration locations are shown in Figure 5-3. Off-site analytical results are provided in Table 7-11.

Arsenic, beryllium, calcium, cobalt, copper, iron, lead, magnesium, manganese, mercury, nickel, selenium, sodium, and zinc were detected at levels in excess of established Devens background concentrations. The greatest number of exceedances were observed in the 7 feet bgs sample from ZWB-95-01X, the 4 feet bgs sample from ZWR-95-37X, the 4 feet bgs sample from ZWR-95-39X, and the surface soil sample ZWS-95-42X. Two of these locations (ZWB-95-01X and ZWR-95-37X) were removed as part of the remedial action in 1997 and 1998. ZWS-95-42X was found to contain 0.85 μ g/g of beryllium and 238 μ g/g of lead.

Detected VOCs were comprised primarily of the fuel related compounds toluene, ethylbenzene, and xylene (TEX). The maximum observed concentration of total TEX was 0.48 $\mu g/g$ (ethylbenzene and xylenes only) in the 7 feet bgs sample from ZWB-95-01X. Other detected analytes include 1,1,2,2-tetrachloroethane in the 5 feet bgs sample from ZWB-95-02X and 0.0034 $\mu g/g$ of styrene in the 6 feet bgs sample from ZWR-95-30X. All of these locations were removed as part of the soil removal action in 1997 and 1998. The VOCs acetone, dichloromethane (methylene chloride), and trichlorofluoromethane (freon) were identified in a number of samples at concentrations consistent with sampling or analysis contamination. Figures 7-5 and 7-6 illustrate VOC, SVOC, and TPHC concentrations in surface and subsurface soils at AOC 69W.

A number of SVOCs were identified in both surface and subsurface soils at AOC 69W. The highest observed concentrations of the PAHs were observed in subsurface soils immediately adjacent to the school building near the new boiler room and in surface soils in the grassy area north of the school. The 7 feet bgs sample from the soil boring ZWB-95-01X contained 149 μ g/g of total SVOCs including 80 μ g/g of 2-methylnaphthalene and 40 μ g/g of naphthalene; this location was removed as part of the remedial action. The highest observed total SVOC concentration in surface soil, 31 μ g/g, was found in ZWS-95-39X. SVOC concentrations appear to be coincident with the pipe leading from the school to the 250 gallon UST.

Harding Lawson Associates

W003971.080

SECTION 7

TPHC were detected in seventeen of the samples collected for off-site analysis. As with the other compounds the highest observed concentrations were found adjacent to the school. The soil boring ZWB-95-01X was shown to contain 14,400 μ g/g of TPHC at 7 feet bgs and the TerraProbeSM point ZWR-95-30X contained 3,240 μ g/g at 6 feet bgs. Other notable concentrations (e.g., in excess of 500 μ g/g) include 661 μ g/g in the surface sample from ZWB-95-01X, 1,390 μ g/g in the 5 feet bgs sample from ZWB-95-02X, 902 μ g/g in the 7 feet bgs sample from ZWR-95-26X, 566 μ g/g in the 7 feet bgs sample from ZWR-95-36X, 1,400 μ g/g in the 4 feet bgs sample from ZWR-95-37X, 936 μ g/g in surface soil sample ZWS-95-35X, and 652 μ g/g in surface soil sample ZWS-95-46X. Most of these locations (excluding ZWR-95-26X, ZWS-95-35X, and ZWS-95-46X) were removed during the soil removal action in 1998. In general, the highest observed TPHC concentrations in soil were located at the water table.

1996 Monitoring Well Boring and Soil Boring Soil Sample Analytical Results. Data from the 1995 RI field investigation suggested that contamination existed underneath the school building. As a result, additional investigation was performed in August of 1996. One soil boring (ZWB-96-03X) and one monitoring well (ZWM-96-21X) were installed in the school courtyard, and a monitoring well was installed in each of the school's two boiler rooms (ZWM-96-19X and ZWM-96-20X). A total of nine soil samples were collected from these explorations and analyzed at off-site laboratories for EPH/VPH parameters, TPHC, and TOC. Analytical data for the 1996 samples are provided in Tables 7-11 and 7-12.

EPH/VPH analysis of the samples yielded one detection. The 9 feet bgs sample from ZWM-96-19X contained an equivalent concentration of 150 μ g/g of total EPH compounds. Identification of the aliphatic concentrations showed 560 μ g/g of the n-C 9 to n-C 18 range and 110 μ g/g of the n-C 19 to n-C 36 range. Aromatics in the n-C 10 to n-C 22 range were identified at 120 μ g/g. None of the targeted PAHs were detected above the reporting limits. VPH analysis showed the same sample to contain 4,100 μ g/kg of VPH compounds, including 270 μ g/kg of the n-C 5 to n-C 8 aliphatic range, 8,300 μ g/kg of the n-C 9 to n-C 12 aliphatics, and 3,500 μ g/kg of the n-C 9 to n-C 10 aromatics. Likewise, none of the targeted VOCs were detected above RLs.

Two of the four soil samples contained detectable levels of TPHC. The 9 feet bgs sample from the new boiler room exploration ZWM-96-19X contained 1,740 μ g/g of

Harding Lawson Associates

TPHC and the 6 feet bgs sample from ZWB-96-03X, located in the courtyard, contained 57.5 μ g/g.

7.2.2.3 Remedial Action Off-Site Soil Analytical Results. Based on a review of the analytical data from the RI, a remedial action was implemented at AOC 69W in the winter of 1997 and 1998 to remove the contaminated soil associated with the 1972 fuel leak and downgradient piping. A total of 31 soil samples were collected from the walls and floor of the excavation, and were analyzed at an off-site laboratory for EPH/VPH parameters to confirm that the remaining soil concentrations were generally consistent with or lower than the MCP S-1/GW-1 soil standards. Analytical data for the remedial action subsurface soil samples are provided in Table 7-13, and are shown in Figures 7-7 (east and west) and 7-8.

EPH/VPH analysis of the samples yielded the highest contaminant concentrations in the southernmost portion of the excavation, adjacent to the elementary school. Concentrations of n-C 9 to n-C 18 aliphatic EPH range, n-C 11 to n-C 22 aromatic EPH range, n-C 9 to n-C 10 aromatic VPH range, and of the individual PAHs 2methylnaphthalene and napthalene exceed their MCP S-1/GW-1 standards in the sample locations along the southern wall of the excavation (69W-HS-SSW-1, 69W-HS-SSW-2, 69W-HS-OB-15, and 69W-HS-FL-2). MCP S-1/GW-1 standards were also slightly exceeded for the n-C 11 to n-C 22 aromatic EPH range at 69W-HS-WSW-3 (in the northern portion of the excavation), and for the n-C 9 to n-C 18 aliphatic VPH range at 69W-V-WSW-1 and 69W-V-ESW-1 (in the northwestern portion of the excavation, near the 250 gallon UST).

The 11 feet bgs sample from 69W-HS-FL-2 contained the highest EPH/VPH results, with the n-C 9 to n-C 18 aliphatic EPH range at 10,000 μ g/g, the n-C 19 to n-C 36 aliphatic EPH range at 1,200 μ g/g, the n-C 10 to n-C 22 aliphatic EPH range at 2,300 μ g/g, the n-C 9 to n-C 12 aliphatic VPH range at 1,300 μ g/kg, and the n-C 9 to n-C 10 aromatic VPH range at 960 μ g/kg. Target PAHs at 69W-HS-FL-2 were detected at 177 μ g/g, and target VOCs were not detected at all.

Target PAHs at 69W-HS-SSW-2, 69W-HS-OB-15, and 69W-HS-SSW-1 were detected at 93, 50, and 98 μ g/g, respectively. Target PAHs were also detected between 1 and 10 μ g/g at 69W-HS-FL-1, 69W-HS-WSW-3, 69W-PL-FL-6, and 69W-HS-NWS-1; however, none of the individual PAH concentrations exceeded MCP S-1/GW-1 standards at these locations. Target PAHs were less than detection limits at all other

Harding Lawson Associates

W003971.080

locations, and target VOCs were less than detection limits at all sample locations.

EPH/VPH concentrations in the 9 foot bgs sample for ZWM-96-19X, located in the new boiler room upgradient of the excavation, were lower than the EPH/VPH concentrations in the samples from the south wall of the excavation.

7.2.2.4 Summary of Soil Impacts. A review of the field and off-site analytical data from the 1995 and 1996 RI field investigations indicated that there were two areas of fuel related soil contamination at AOC 69W. The larger area extended from the new boiler room to the 250 gallon UST in the wooded area approximately 300 feet northwest of the school. The contamination was attributed to the 1972 release of fuel oil from piping between the 10,000 gallon UST and the new boiler room. Analytical data and visual evidence suggested that the release may have been inside or near the new boiler room. As a result of the release a "skimmer system" was installed in 1972 to remove oil from the source area and presumably from near surface soils in the grassy area north of the school. Contaminant distributions indicated that the underground piping associated with this system may have acted as a conduit for contaminant migration. Detected contaminants were primarily TPHC, PAHs, and EPH/VPH at approximately 6 to 10 feet bgs adjacent to the school and 0 to 4 feet bgs downgradient in the grassy area and vicinity of the 250 gallon UST. Observed subsurface contaminants were identified primarily at or near the water table. Surficial contamination downgradient of the school (in the vicinity of Willow Brook) is attributed to sorption during times of high water levels.

Based on the nature and distribution of contaminants, a remedial action was undertaken in the winter of 1997 and 1998 to remove contaminated soil associated with the 1972 release; soil was excavated and removed to a maximum depth of 13 feet bgs near the school, and 8 feet bgs near the 250 gallon UST. Confirmatory subsurface soil sample results from the remedial action suggest that concentrations of fuel-related contaminants still exceed MCP S-1/GW-1 standards immediately adjacent to the elementary school, but are generally low in downgradient areas (only a few concentrations in soil slightly exceeded MCP S-1/GW-1 standards).

The other identified area of soil contamination is located adjacent to the school building outside of the old boiler room. This contamination is attributed to the 1978 release of fuel oil due to ruptured piping. An excavation at the time of the release showed visible fuel oil contamination emanating from underneath the school.

Harding Lawson Associates

Analytical data indicate that the contaminants are primarily TPHC at depths of 4 to 7 feet bgs beneath the paved parking lot. Contaminants appear to be localized in the area immediately adjacent to the school. Based upon the absence of site-related contaminants in downgradient soils (e.g., ZWR-95-27X, ZWR-95-54X, and ZWR-95-55X), future migration is not likely as the area is paved, thereby, inhibiting leaching of soils via precipitation.

7.2.3 AOC 69W RI Groundwater Results

The following discussion of groundwater sampling includes field analytical results of water samples collected from TerraProbeSM borings in 1995 as well as the off-site laboratory analytical results for the four rounds of RI groundwater sampling (two rounds in conjunction with the 1995 field effort, and two rounds of low-flow sampling as part of the 1996 and 1997 field efforts).

7.2.3.1 RI Field Analytical Groundwater Results. A total of 29 groundwater samples were collected from the TerraProbeSM points and analyzed in the field for BTEX, select VOCs, and GRO for select samples. Data were used to delineate horizontal contaminant distribution and confirm placement of monitoring well locations. Field analytical results are provided in Table 7-14 and Figure 7-9.

Seven samples contained one or more of the fuel-related contaminants chlorobenzene, ethylbenzene, and xylenes. The majority of detections were from TerraProbeSM points adjacent to the north side of the school building, including the highest observed concentrations of ethylbenzene (73 μ g/L) and xylene (120 μ g/L) in ZWR-95-28X. Xylenes were also found in two samples from TerraProbeSM points ZWR-95-37X and ZWR-95-45X at concentrations of 7 μ g/L and 7.3 μ g/L, respectively. These explorations were located in the vicinity of the 250 gallon underground storage tank. The only other VOCs detected in groundwater samples were 1,1,1-TCA at 7 μ g/L in ZWR-95-27X, and 5.1 μ g/L of 1,1,1-TCA and 12 μ g/L of TCE in ZWR-95-33X. Both of these explorations are located adjacent to the school building near the TPHC contamination attributed to the 1978 fuel oil release from the old boiler room.

Harding Lawson Associates

W003971.080

7.2.3.2 RI Groundwater Off-Site Laboratory Analytical Sample Results.

Round 1 (November 1995) and Round 2 (February 1996) Groundwater Sampling. As part of the RI field investigation HLA installed four monitoring wells at Area 2 in 1995 (ZWM-95-15X, ZWM-95-16X, ZWM-95-17X, and ZWM-95-18X) to supplement the six existing AREE 69W monitoring wells (69W--94-09, 69W--94-10, 69W--94-11, 69W--94-12, 69W--94-13, and 69W--94-14) (Figure 5-3). Two rounds of groundwater sampling were conducted on all of the monitoring wells (except for 69W--94-09) between Rounds 1 and 2. Groundwater samples were analyzed for PAL VOCs, SVOCs, total and filtered PAL inorganics, pesticides/PCBs, TPHC, TDS, and water quality parameters. Analytical results for the Round 1 and Round 2 sampling events are provided in Table 7-15 and Figure 7-10 (TPHC results only).

Several inorganic analytes were detected above the calculated Devens background concentrations in groundwater. Arsenic, calcium, iron, manganese, potassium, and sodium were detected above background in the filtered samples. All of the above inorganic analytes as well as copper were detected above background in one or more of the unfiltered samples. The greatest numbers of background exceedances in both Rounds 1 and 2 were observed in samples from monitoring wells 69W--94-10 and 69W--94-13; soil from around these monitoring wells were removed as part of the soil removal action. These were also the only wells to have inorganics concentrations in excess of MCLs. The arsenic concentrations in 69W--94-10 in both the filtered and unfiltered samples ranged between 150 μ g/L and 180 μ g/L while the arsenic concentrations in the Round 1 and Round 2 samples from 69W--94-13X ranged between 260 μ g/L (Round 1 unfiltered) and 110 μ g/L (Round 2 filtered). The arsenic is believed to be due to reducing conditions in the aquifer and the reducing conditions are attributed to the aerobic degradation of the fuel oil contamination.

The most commonly detected VOCs were the fuel related compounds toluene, ethylbenzene, and xylenes. One or more of these compounds were found in eight of the monitoring wells (69W--94-10, 69W--94-12, 69W--94-13, 69W--94-14, ZWM-95-15X, ZWM-95-16X, ZWM-95-17X, and ZWM-95-18X). Monitoring wells 69W--94-10, 69W--94-12, and 69W--94-14 were the only wells to have detections of the above compounds in both Rounds 1 and 2. The maximum concentrations of these compounds were observed in 69W--94-10, 18.4 μ g/L in the Round 1 sample comprised of toluene, ethylbenzene, and xylenes, and 20 μ g/L in the Round 2 sample comprised entirely of ethylbenzene. As previously mentioned, the soils from around

Harding Lawson Associates

69W--94-10 were removed during the removal action.

The chlorinated VOCs 1,1,1-TCA and TCE were detected in the Round 1 sample from 69W--94-11 at 1.5 μ g/L and 3.3 μ g/L respectively. TCE was also observed at 1.2 μ g/L in the Round 1 sample from ZWM-95-16X. TCE and 1,1,1-TCA were not detected in any of the Round 2 samples. Chloroform was detected in the Round 1 samples from 69W--94-12 and 69W--94-13 at 0.55 μ g/L and 1.1 μ g/L, respectively. Acetone, the only other VOC detected, was present in three of the samples at concentrations consistent with laboratory or sampling contamination.

The majority of Round 1 and Round 2 SVOC compounds were detected in the monitoring wells 69W--94-10 and 69W--94-13, which were both removed as part of the soil remedial action. 69W--94-10 was shown to contain 1,380 μ g/L of SVOCs in Round 1 and 1,500 μ g/L of SVOCs in Round 2. Comprising the total concentrations were 2-methylnaphthalene at 500 μ g/L (Round 1) and 600 μ g/L (Round 2), fluorene at 80 μ g/L (Round 1 only), naphthalene at 200 μ g/L (Rounds 1 and 2), phenanthrene at 100 μ g/L (Round 1) and 200 μ g/L (Round 2), and bis(2-ethylhexyl) phthalate at 500 μ g/l (Rounds 1 and 2). 69W--94-13 contained considerably less total SVOCs at 32 μ g/L and 29 μ g/L for Rounds 1 and 2.

The Round 2 sample from 69W--94-10 was shown to contain 0.115 μ g/L of gamma chlordane and 0.059 μ g/L of heptachlor epoxide. This was the only sample shown to contain pesticides. None of the Rounds 1 or 2 groundwater samples contained detectable levels of PCBs.

69W--94-10 was the only monitoring well to contain TPHCs in both the Round 1 and Round 2 samples, 159,000 μ g/L and 228,000 μ g/L respectively. Analysis of the Round 1 sample from 69W--94-11 indicated that TPHC concentrations were below the detection limit. A duplicate sample was collected from this well during the Round 2 sampling event. The original sample was shown to contain 2,420 μ g/L of TPHC, but the TPHC concentration in the duplicate was below the detection limit. Similar results were seen in samples from other wells. The Round 1 samples from 69W--94-13 (523 μ g/L), 69W--94-14 (1,960 μ g/L), ZWM-95-15X (281 μ g/L), and ZWM-95-16X (1,340 μ g/L) all contained TPHC while all of the Round 2 samples were below detection levels.

Harding Lawson Associates

W003971.080

It should be noted that for all the samples the total suspended solids decreased dramatically, over an order of magnitude for most samples, from Round 1 to Round 2.

Round 3 (September and October 1996) Low-Flow Groundwater Sampling. Groundwater samples were collected from six monitoring wells as part of the 1996 field effort to delineate potential source areas. The monitoring wells were sampled following USEPA low-flow (minimum stress) purging and sampling protocols (USEPA, 1996). Sampled wells included the three newly installed courtyard and boiler room monitoring wells (ZWM-96-19X, ZWM-96-20X, and ZWM-96-21X) as well as the existing monitoring wells 69W--94-10, 69W--94-11, and ZWM-95-16X. Existing wells were chosen based upon their proximity to suspected source areas and results from Rounds 1 and 2 sampling. Groundwater samples were analyzed for EPH/VPH, TPHC (Method 418.1), water quality parameters, TDS, and TOC. Total concentrations for EPH/VPH compounds are reported as equivalent concentrations while aliphatic ranges, aromatic ranges, and targeted analytes are reported as actual measured concentrations. Analytical results for the Round 3 groundwater sampling are provided in Tables 7-15 and 7-16. Figure 7-10 provides a comparison of all TPHC (Rounds 1, 2, and 3) and EPH/VPH data.

Three of the monitoring wells contained measurable levels of VPH. 69W--94-10, which was removed as part of the soil remedial action, exhibited the highest concentration of total VPH at 830 μ g/L. The total concentration consisted of 17 μ g/L of the n-C 5 to n-C 8 aliphatics, 550 μ g/L of the n-C 9 to n-C 12 aliphatics, and 790 μ g/L of the n-C 9 to n-C 10 aromatic range. This sample also contained the only detections of targeted VOCs: 35 μ g/L of ethylbenzene and 94 μ g/L of naphthalene. EPH compounds were only detected in 69W--94-10. Total EPH compound concentrations were 740 μ g/L comprised of 590 μ g/L of the n-C 9 to n-C 18 range aliphatics and 710 μ g/L of the n-C 10 to n-C 22 range aromatics. Targeted PAH (SVOC) analytes consisted of 89 μ g/L of 2-methylnaphthalene, 45 μ g/L of naphthalene, and 15 μ g/L of acenaphthene. These concentrations in groundwater, which likely reflect soil contamination, have been mitigated by the soil removal around 69W--94-10.

The monitoring well installed in the new boiler room, ZWM-96-19X contained 47 μ g/L of VPH compounds. Components of the total VPH concentration consisted of 34 μ g/L of the n-C 9 to n-C12 aliphatic range and 45 μ g/L of the n-C 9 to n-C10

Harding Lawson Associates

aromatic range. A duplicate sample from ZWM-96-19X substantiated these results. VPH compounds were also detected in ZWM-95-16X located adjacent to the school outside of the old boiler room. Total VPH compounds were 7 μ g/L consisting entirely of the n-C 9 to n-C 10 aromatic range.

TPHCs were below detection limits in all of the Round 3 samples. Total suspended solids in the Round 3 samples were comparable to the Round 2 levels with a maximum concentration of 37,000 μ g/L in 69W--94-10.

Round 4 (December 1997) Low-Flow Groundwater Sampling. Groundwater samples were collected from thirteen monitoring wells as part of the 1997 field effort to further characterize potential contamination in groundwater. The monitoring wells were sampled following USEPA low-flow (minimum stress) purging and sampling protocols (USEPA, 1996). Sampled wells included the three courtyard and boiler room monitoring wells (ZWM-96-19X, ZWM-96-20X, and ZWM-96-21X) and the monitoring wells located upgradient and downgradient of the elementary school (69W--94-09 through 69W--94-14, and ZWM-95-15X through ZWM-95-18X). Subsequent to the fourth round of groundwater sampling, soil from around four monitoring wells (ZWM-95-15X, 69W--94-10, 69W--94-11, and 69W--94-10) was removed as part of the soil removal action. Two monitoring wells (69W--94-10) were also removed.

Groundwater samples were analyzed for EPH/VPH (with target PAHs and VOCs), target VOCs by Method 8260, and target PAHs by Method 8270. Aliphatic ranges, aromatic ranges, and targeted analytes measured by EPH/VPH are reported as actual measured concentrations. Analytical results for the Round 4 groundwater sampling are provided in Tables 7-15 and 7-16. Figure 7-11 presents the results of the 1997 EPH/VPH, VOC, and PAH groundwater results.

Three of the monitoring wells contained measurable levels of VPH. 69W--94-10 and 69W--94-13 exhibited the highest n-C 9 to n-C 12 aliphatic ranges (120 and 140 μ g/L, respectively), and the highest n-C 9 to n-C 10 aromatic ranges (430 and 330 μ g/L, respectively). These samples also contained the only targeted VOCs by VPH (ethylbenzene and naphthalene), and VOCs by Method 8260 (ethylbenzene). The concentrations of n-C 9 to n-C 10 aromatic VPHs and naphthalene (targeted VOC) exceed the MCP GW-1 standard in both wells. As previously mentioned, soil from around these two monitoring wells was removed as part of the soil removal action

Harding Lawson Associates

W003971.080

in 1997 and 1998. ZWM-95-16X, located adjacent to the school outside of the old boiler room, contained 75 μ g/L of the n-C 9 to n-C 10 aromatic VPH range; however, no target VOCs by VPH, or VOCs by Method 8260 were detected.

EPH compounds were detected in several of the monitoring wells, including 69W--94-10, 69W--94-11, 69W--94-13, 69W--94-14, and ZWM-95-15X (all located within or near the area of the soil remedial action). The n-C 9 to n-C 18 aliphatic EPHs ranged from 32 to 84 μ g/L, and the n-C 11 to n-C 22 aromatic EPHs ranged from 84 to 480 μ g/L in these well. The MCP GW-1 standard for n-C 11 to n-C 22 aromatic EPHs was exceeded by concentrations detected in 69W--94-10 and 69W--94-Targeted PAH (SVOC) analytes detected in monitoring wells include 2-13. methylnaphthalene, acenaphthene, acenaphthylene, fluoranthene, and naphthalene. SVOCs (by Method 8270) were also detected in two of the wells (69W--94-10 and 69W--94-13), including 2-methylnaphthalene, naphthalene, fluorene. and phenanthrene. The concentrations of 2-methylnaphthylene in both wells, and the concentration of naphthalene in 69W--94-10 exceed the MCP GW-1 standards for target PAHs by EPH and for PAHs by 8270.

7.2.3.3 Summary of Groundwater Impacts. Fuel-related VOCs, SVOCs, TPHC, and inorganics comprise the observed groundwater contaminants at AOC 69W. Varying degrees of groundwater contamination, as identified by field and off-site analysis, were observed to extend from the new boiler room towards the 250 gallon UST located approximately 300 feet to the northwest. The area of groundwater contamination was coincident with the underground pipe associated with the "skimmer system" installed in response to the 1972 fuel oil release. Contaminant concentrations were highest between the new boiler room and monitoring well 69W--94-13, which was also the area of highest observed soil concentrations. The soil around several monitoring wells that exhibited the highest chemical concentrations was removed during the soil removal action (including 69W--94-10 and 69W--94-13).

Arsenic, calcium, iron, manganese, potassium, and sodium were detected in filtered samples at levels in excess of calculated Devens background levels. The greatest number of background exceedances in Rounds 1 and 2 were observed in monitoring wells 69W--94-10 and 69W--94-13. Contaminated soils surrounding these wells were removed during the soil removal effort.

Harding Lawson Associates

There does not appear to be any significant groundwater contamination associated with the 1978 fuel oil release in the vicinity of the old boiler room. Low levels of chlorinated VOCs were detected during the 1995 field analysis and Round 1 groundwater sampling; however, there were no chlorinated VOCs detected during the Round 2 groundwater sampling effort.

TPHC data (by Method 418.1) from groundwater samples is highly variant from round to round and even between samples and duplicates. The most reasonable explanation is a methodology problem during IR analysis, as the EPH/VPH analyses do not show the same variations. A data quality review could not substantiate this hypothesis.

7.2.4 AOC 69W RI Sediment Results

Sediment samples were collected from six locations within Willow Brook in the vicinity of AOC 69W (Figure 5-3): ZWD-95-01X, ZWD-95-02X, ZWD-95-03X, ZWD-95-04X, ZWD-95-05X, and ZWD-95-06X. Surface sediment samples (0 to 6 inches bgs) were collected at every location, and deep sediment samples were collected from 2 to 2.5 feet bgs at ZWD-95-01X, ZWD-95-02X, and ZWD-95-03X. All of the sediment samples were analyzed for PAL VOCs, SVOCs, inorganics, pesticides, PCBs, TPHC (Method 418.1), petroleum fingerprinting, and TOC. Analytical data are provided in Table 7-17 and Figure 7-12. Willow Brook was dry at the time of RI sediment sampling.

There is no established set of background concentrations for sediments at Devens; therefore, inorganics concentrations in sediments were compared to background concentrations in soils. The inorganic analytes arsenic, calcium, chromium, cobalt, and manganese were all detected at levels in excess of established Devens background concentrations for soil. The greatest number of background exceedances were observed in the surface and 2 feet bgs samples from the upgradient sampling point ZWD-95-01X.

The only observed VOCs were the common laboratory and sampling contaminants dichloromethane (methylene chloride) and trichlorofluoromethane (freon). Observed concentrations of these analytes in the sediment samples were consistent with laboratory and/or sampling contamination.

Harding Lawson Associates

SVOC compounds were observed in all but one of the RI sediment samples. The maximum observed concentration of total SVOCs $(27.7 \ \mu g/g)$ was observed in the 2 feet bgs sample from ZWD-95-02X. This sample also contained the maximum observed concentrations of all but one (benzo[k]fluoranthene) of the individual SVOCs. This sample is located in Willow Brook adjacent to the 250 gallon UST. The surface sample from this location was the only sample without detectable levels of SVOCs. The second highest concentration of total SVOCs (11 $\mu g/g$) was observed in the 2 feet bgs sample from the downgradient sampling point ZWD-95-03X. Both upgradient and downgradient surface sediment samples were shown to contain similar levels of SVOCs. Of the eight SVOCs observed in AOC 69W groundwater, only one (phenanthrene) was detected in sediment samples.

Pesticides were also detected in all but one of the RI sediment samples. The highest levels of total pesticides, as well as the highest individual concentrations, were found in the upgradient surface sediment sample ZWD-95-05X. This sample was found to contain 2.1 μ g/g of 4,4'-DDD, 0.081 μ g/g of 4,4'-DDE, 0.4 μ g/g of 4,4'-DDT, 0.013 μ g/g of alpha chlordane, 0.024 μ g/g of gamma chlordane, 0.06 μ g/g of Dieldrin, and 0.05 μ g/g of Endosulfan II. This was also the only sample to contain these last four pesticides. The only sample in which pesticides were not detected was the 2 feet bgs sample from ZWD-95-02X. The lack of detections at depth, and the preponderance of high pesticides concentrations bordering the maintained lawn suggests that the pesticides are not site related.

The surface sample from ZWD-95-05X was the only sample to contain PCBs. Aroclor-1260 was detected at 0.51 μ g/g.

The three highest concentrations of TPHC were observed in the upgradient surface samples: ZWD-95-05X (1,230 μ g/g), ZWD-95-01X (896 μ g/g), and ZWD-95-04X (386 μ g/g). The highest concentration of TPHC in the downgradient samples, 287 μ g/g, was observed in ZWD-95-06X. Petroleum fingerprinting indicated that the TPHC detections were comprised primarily of gasoline and diesel patterns. For much of its course, Willow brook is bordered by a maintained lawn adjacent to MacArthur Avenue. The elevated TPHC concentrations at the upgradient locations are likely attributable to refueling and operation of lawn care equipment. In addition, the stream bed is lined with chunks of asphalt.

Harding Lawson Associates

W003971.080

YG.

7.2.5 AOC 69W RI Air Sampling Results

As discussed in Subsection 5.4.10, indoor air sampling was completed in October of 1997 at the Devens Elementary School to determine if fuel-related contaminants in soil and groundwater beneath the school contribute to ambient contaminant levels in the school. The off-site analytical results are summarized in Table 7-18, and a complete report of activities and sample data are included in Appendix M. Air sample locations and results are shown in Figure 7-13.

Acetone, toluene, and xylene were detected in one or more of the quality control blank samples associated with the data set indicating potential for false positive contamination. For some samples, results were qualified non-detect "U" in accordance with USEPA Region I guidelines. The analytical results for ZWA-92-04X have been rejected because the sample vacuum was lost during the sampling period.

A well headspace sample (ZWA-97-09X) was collected from ZWM-95-19X on October 14, 1997, to confirm the fingerprint of soil vapor contamination, and to refine the analyte list for the air sampling program. Chemical results for this sample indicate the presence of tetrachloroethylene (360 micrograms per cubic meter $[\mu g/m^3]$) in soil gas from beneath the new boiler room. Tetrachloroethylene was included in the air sampling program, but was not detected in any other air samples, including samples collected from the dirt-floored crawlspace beneath the kitchen and cafeteria (i.e., the areas most likely to be affected by subsurface contamination).

Samples collected from the crawlspace beneath the kitchen (ZWA-97-06X) and the cafeteria (ZWA-97-05X) contained toluene (150 and 13 μ g/m³, respectively). The sample from beneath the kitchen also contained xylene (18.3 μ g/m³), octane 5.5 (μ g/m³), and ethylbenzene (5.2 μ g/m³).

The three outdoor background samples, ZWA-97-11X, ZWA-97-12X, and ZWA-97-13X, also contained toluene; the results for ZWA-97-12X and SWA-97-13X are "U" qualified because they were also detected in the blank. The indoor background sample, ZWA-97-07X, had detectable levels of six compounds, including xylene (24.8 $\mu g/m^3$), 2-methylheptane (7.2 $\mu g/m^3$), 3-methylheptane (8.9 $\mu g/m^3$), ethylbenzene (9.9 $\mu g/m^3$), nonane (5 $\mu g/m^3$), and octane (9.1 $\mu g/m^3$). Toluene was qualified "U" nondetected during validation. This indoor background sample contained all compounds

Harding Lawson Associates

W003971.080

detected in air samples from the site.

Samples from within the school contained anywhere from three compounds (ZWA-97-08X, from the room adjacent to the new boiler room) to eight compounds (ZWA-97-03X, from a room near the old boiler room in the northeast corner of the school). In general, compound concentrations were consistent throughout the school, and were of the same magnitude as those in the indoor background sample. The only notable detection was toluene $(1,000 \ \mu g/m^3)$ in sample ZWA-97-03X, which exceeds available inhalation toxicity values. All other detections are less than the inhalation toxicity values.

The results of the air sampling program suggest that there are low levels of alkanes and aromatic hydrocarbons in the elementary school air samples that are generally consistent with background levels of these chemicals in air.

Harding Lawson Associates

8.0 CONTAMINANT FATE AND TRANSPORT

This subsection discusses the migration potential and probable environmental fate of general contaminant groups identified at AOC 69W. Compounds and analytes detected include VOCs, SVOCs, inorganics, and TPHC. The observed distribution of these contaminants in different environmental media (soil, groundwater, and sediment) is the result both of the release pattern and of their physical and chemical properties. For organic chemicals, these properties include specific gravity, solubility, volatility, and organic carbon partition coefficient (K_{∞}). For inorganic constituents, the physical and chemical properties include oxidation state of the analyte, pH, and specific solute species. Site-specific conditions governing fate and transport (e.g., persistence and migration) of analytes include contaminant concentration, topography, meteorological conditions, and in the case of groundwater, hydrogeology.

8.1 COMPOUND PROPERTIES AND TRANSPORT PROCESSES

The primary contaminants detected in soil at AOC 69W are fuel-related VOCs, SVOCs, and TPHC. In addition, some VOCs and SVOCs may have been introduced in samples as laboratory contamination.

The persistence of compounds in soil is determined by chemical properties, source configurations and releases, geochemical and biochemical reactions, and soil and meteorological conditions. Factors and processes that control the persistence of chemicals in water-bearing units, in addition to the aforementioned factors, are water-bearing unit characteristics, advection, and hydrodynamic dispersion. Compounds may exist in the surface and subsurface in gaseous, aqueous, or solid phases. The fate of these compounds is controlled by a combination of all of these factors.

The following subsections discuss general physical and chemical properties, and how these properties affect transport and general attenuation processes.

Harding Lawson Associates

W003971.080

8.1.1 Physical and Chemical Properties Significant to Fate and Transport

This subsection discusses the physical and chemical properties that affect the fate and transport of contaminants in the environment. Physical and chemical properties of organic contaminants of concern detected at AOC 69W are presented in Table 8-1. Table 8-2 summarizes the relative mobilities of selected inorganic elements in different chemical environments.

Most physical and chemical properties of PAL analytes, including specific gravities, K_{∞} , relative solubility, and relative volatility, are described in "Basics of Pump-and-Treat Groundwater Remediation Technology" (USEPA, 1990b). This reference document does not include inorganics, because analyses conducted measure the total amount of a particular constituent in the sample rather than the actual chemical form or metal oxidation state. The distribution of specific solute species, pH, and oxidation are important factors in establishing the total solubility or mobility of a given inorganic element.

Specific gravity is the ratio of the mass of a given volume of a liquid substance to the mass of an equal volume of water. Liquids with specific gravities greater than 1 are termed "heavier" than water.

Solubility measures the partitioning between the aqueous phase and solid form of a chemical, and the tendency of a material to dissolve in water. Substances with lower solubilities are more likely to remain in a separate phase when in contact with water; substances with higher solubilities will dissolve into, and move with, water.

Volatility measures the tendency of a chemical to partition into the gaseous phase. Volatility can be predicted by an analyte's vapor pressure and Henry's Law Constant value (H_c). Volatility of a compound increases with increasing vapor pressure. Compounds with H_c values less than 1.0×10^{-5} (e.g., pyrene) have a low degree of volatility, and those with H_c values below 3.0×10^{-7} are considered non-volatile. H_c values between 1.0×10^{-5} and 1.0×10^{-3} (e.g., naphthalene and phenanthrene) are moderately volatile, while those with values exceeding 1.0×10^{-3} (e.g., VOCs) are considered highly volatile.

 K_{∞} measures the extent that an organic chemical partitions between a solid phase and a liquid phase, and is used to predict to what extent a chemical could be

Harding Lawson Associates

W003971.080

adsorbed to soil organic carbon. Chemicals with a K_{∞} greater than 10,000 will adsorb strongly to soil organic carbon (e.g., fluoranthene, phenanthrene, and pyrene) Chemicals with a K_{∞} ranging from 1,000 to 10,000 will moderately adsorb, and move slowly in the soil profile (e.g., naphthalene). Chemicals with a K_{∞} of less than 1,000. weakly adsorb to soil organic carbon and tend to be more mobile. Examples of weakly adsorbed compounds include many VOCs such as benzene and xylene.

8.1.2 General Transport and Attenuation Processes

Migration and persistence are controlled by various transport and attenuation processes. Processes that tend to disperse contaminants include surface water and groundwater movement (which includes the movement of dissolved and suspended contaminants), facilitated transport, leaching by dissolution or desorption, and surface erosion.

The solubility of a compound in water is considered to be the most important transport factor, because it determines the maximum concentration dissolved in water. Knowledge of the solubility of a chemical provides considerable insight into the fate and transport of that chemical. In general, highly soluble compounds are less likely to partition into soil or sediment, or to volatilize from water, and are more likely to biodegrade (Montgomery, 1991).

Dissolved phase transport can occur via two processes: advection or dispersion. Advection involves transport with flowing groundwater and migrating with the mean velocity of the solvent (groundwater plus dissolved compounds). When compounds move through the ground by advection, they are subject to spreading within the ground, which allows compounds with little or no affinity for soils to migrate faster than the mean groundwater velocity. This spreading is the result of a process known as dispersion. Hydrodynamic dispersion has two components: molecular diffusion and mechanical dispersion (USEPA, 1989a). Diffusion is the process by which ionic or molecular constituents move under the influence of concentration gradients. Mechanical dispersion occurs as the groundwater flows through the media, and compounds spread out through the tortuous pathways of the soil matrix, and mix with clean water. The result is a dilution of the compound by a process known as dispersion (Fetter, 1988). At very low groundwater velocities, diffusion is the dominant process; at higher velocities, mechanical dispersion is the dominant process. Dispersivity is dependent on vertical and horizontal permeability variations,

Harding Lawson Associates

increasing with the degree of heterogeneity and anisotropy, and is dependent on whether flow is principally through porous media or nonporous media (e.g., fractured bedrock) (Walton, 1988).

The rate a compound migrates can be influenced by facilitated transport, which is the combined effects of physical, chemical, and/or biological phenomena that act to increase mobility. Examples of facilitated transport include particle transport, cosolvation, and phase shifting (Keely, 1989).

Particle transport involves the movement of small, solid-phase particles (such as inorganic and organic colloids), macromolecules, or emulsions to which compounds have adhered by sorption, ion exchange, or other means. High molecular weight organic compounds such as PAHs, pesticides, and heavy metals, have a high affinity for mobile subsurface particles, and this affinity increases their mobility (Huling, 1989). Small particles, especially mobile organic carbon phase particles such as biocolloids and macromolecules (e.g., humic substances) are transported in the aqueous phase and may act as mobile sorbents. PAHs and pesticides were detected at AOC 69W.

Cosolvation is the process by which the solubility and mobility of one compound is increased by the presence of another (Keely, 1989). Naturally occurring organic compounds (e.g., humic acids) can undergo complexation reactions with metals and pesticides. Complexation reactions can increase the solubility of metals (including iron, aluminum, copper, nickel, and lead) and pesticides (e.g., DDT). In a cosolvent system, as the fraction of a water-miscible cosolvent increases, the solubilities of the metals or pesticides increase. However, the cosolvent concentration normally needs to be high to ensure a substantial increase in solute velocity. Therefore, cosolvation is important primarily near sources of groundwater impact (USEPA, 1989a). High concentrations of water-miscible phases (e.g., ketones) were not detected at AOC 69W.

Chemical phase shifts involve changes in pH and/or the redox potential of the groundwater. These shifts can increase solubilities and mobilities by ionizing neutral organics, solubilizing precipitated metals, forming complexes, or limiting biological activity (Keely, 1989). These processes are particularly important in determining the mobility of heavy metals.

Harding Lawson Associates

Processes that tend to attenuate migration of impacted groundwater include retardation resulting from sorption, volatilization, degradation, and precipitation. The sorption properties of individual solutes are dependent on soil and groundwater characteristics. In general, the relative amount of sorption by soil or sediment materials that do not contain organic matter is as follows: clay > silt > sand > gravel (Walton, 1988). The soil beneath AOC 69W is a silty sand to sand. Sorption would be expected to exert a moderate to minimal influence in retarding the migration of fuel-related VOCs and SVOCs in the soil.

The tendency of organic chemicals to be sorbed is also dependent on the organic content of the soil and the degree of hydrophobicity (lack of affinity for water) of the solute (contaminant). The rate of travel for each chemical depends on the groundwater seepage velocity and the degree of sorption. If an organic chemical is extensively adsorbed by particles, it will be rendered relatively immobile. The rates and degree of volatilization, photolysis, hydrolysis, and biodegradation are directly dependent on the extent of adsorption (Montgomery, 1991). The vadose zone typically contains greater amounts of organic material and metal oxides (which may also act as sorbents) than the saturated zone, which may make the rate of movement in the vadose zone substantially less than that in the saturated zone (USEPA, 1989a).

The soil partition or sorption coefficient (K_{∞}) is defined as the ratio of adsorbed chemical per unit weight of organic carbon to the aqueous solute concentration. The coefficient indicates the tendency of a compound to adsorb to organic carbon (degree of retardation) and, therefore, provides a means for estimation of the relative mobility of solutes (Montgomery, 1991). Mobility is a function of the relative rate of transport of a chemical versus the rate of groundwater flow. Chemicals that have relatively low mobilities (i.e., high retardation or sorption) move slowly compared to the velocity of the groundwater. Chemicals that have relatively high mobilities (i.e., low retardation or sorption) move at a rate closer to groundwater velocity. VOCs detected at AOC 69W have relatively high mobility potential, while SVOCs have moderate to high mobility potential (Table 8-1).

Volatilization is the transport of a compound from the liquid to the vapor phase and, ultimately, into the atmosphere. Volatilization rates are affected by soil properties, vapor pressure, temperature, and sorption. VOCs partition between the aqueous and gaseous phase in unsaturated soils. This process will occur most readily for compounds with a high vapor pressure and a high H_e. These compounds tend to

Harding Lawson Associates

W003971.080

partition off into the gas phase and occupy the available soil pore space. In addition, VOCs in the saturated zone or in surface water will partition to the gaseous phase, particularly those with lower solubility (e.g., xylenes). VOCs with greater aqueous solubility (e.g., benzene) tend to remain in solution.

Volatilization is an important process in shallow soils and surface water. In recharge areas composed of sandy or gravelly soil, volatilization may be an important process, especially for compounds with moderate to high volatility (Montgomery, 1991). The effectiveness of volatilization normally decreases with depth in the soil column.

Chemicals released to the environment are susceptible to several degradation pathways, including chemical degradation (e.g., oxidation and reduction); photolysis or photochemical degradation; and biodegradation. Compounds formed by these processes may be more or less toxic and/or more or less mobile than the parent compound.

Oxidation typically involves the loss of electrons during a chemical reaction. In general, substituted aromatic compounds such as ethylbenzene and naphthalene can be oxidized. Oxidation rates for aromatic compounds are typically an order of magnitude faster than for chlorinated aliphatic compounds (e.g., 1,2-dichloroethane [1,2-DCA]). Overall, abiotic (without biological life) oxidation of organic compounds in groundwater systems is limited.

Photochemical breakdown processes involve structural changes in a molecule induced by radiation in the ultraviolet-visible light range. This process may occur in surficial soils at AOC 69W but would not affect contamination in the subsurface soils.

Biodegradation may be defined as the breakdown of organic compounds by microorganisms through metabolic processes. Variables affecting the rate of biodegradation include:

- number of microorganisms
- chemical properties, concentrations, and distribution
- presence of food and nutrients
- temperature
- pH
- moisture and oxygen content

Harding Lawson Associates

W003971.080

The rate of biodegradation tends to be higher for low molecular weight compounds. Naturally occurring soil and aquatic microorganisms capable of degrading aromatic hydrocarbons (e.g., BTEX) have been studied, and a relationship between dissolved oxygen and biodegradation has been documented (Jamison, et al., 1975; and Bailey, et al., 1973). As the aromatic hydrocarbons are mobilized by dissolution from soil or sediment, they are likely to be rapidly degraded as long as microorganisms and dissolved oxygen are available. Degradation rates for fuel-related contaminants are much slower under anaerobic conditions.

8.2 FATE AND TRANSPORT OF CONTAMINANTS DETECTED AT AOC 69W

This subsection discusses the potential fate and transport of contaminants, by chemical class, detected at AOC 69W.

VOCs. Soil samples collected at or below the water table at AOC 69W contained the fuel-related VOCs TEX as well as the chlorinated aliphatics 1,1,2,2-PCA, acetone, dichloromethane, styrene, and Freon (see Table 7-12). TEX are the primary VOCs detected in groundwater samples from AOC 69W (see Table 7-15 and 7-16). Limited chlorinated aliphatics were detected in one of the Round 1 groundwater samples.

The majority of site-related VOCs detected at AOC 69W can be classified as aromatic hydrocarbons (e.g., TEX). Processes and forces that will control the fate of these VOCs include volatilization, advection/dispersion, and biodegradation.

Factors affecting VOC percolation to groundwater are density and volatility. Compounds with higher density and low volatility are most likely to be transported to groundwater.

Dissolution of VOCs from unsaturated zone soil via infiltrating precipitation may be a transport mechanism for unpaved areas at AOC 69W due to the sandy nature of the soils and the relatively shallow water table.

Volatilization is believed to be the most significant transport mechanism for VOCs in the unsaturated soils at AOC 69W. The fuel-related VOCs at AOC 69W are likely partitioning between the aqueous and gaseous phases in the source area

Harding Lawson Associates

unsaturated soils. This process occurs most readily for compounds with a high vapor pressure and a high H_c (e.g., toluene). In addition, VOCs in the saturated zone will partition to the gaseous phase, particularly those with lower solubility (e.g., TEX). As groundwater transports the fuel-related VOCs away from the source areas, the VOCs with lower solubility will partition to some extent into the gas phase and occupy the available soil pore space above the water table in the unsaturated zone. VOCs with greater aqueous solubility (e.g., benzene which was not detected at AOC 69W) tend to partition more strongly to the aqueous phase.

Dissolved phase transport of VOCs in groundwater is a significant transport mechanism at AOC 69W. Factors affecting partitioning of VOCs from soil to groundwater include solubility and K_{∞} . VOCs with high solubilities and low $K_{\infty}s$, such as benzene, will partition to groundwater from the saturated zone soils. TEX were detected in saturated zone soils and groundwater, which is probably a result of the moderate $K_{\infty}s$ and solubilities (see Table 8-1). Processes that tend to attenuate migration of impacted groundwater at AOC 69W include retardation resulting from sorption, volatilization, and degradation.

Biodegradation reactions act to reduce the total mass of VOCs. Naturally occurring soil microorganisms capable of degrading aromatic hydrocarbons have been studied, and a relationship between dissolved oxygen and biodegradation has been documented (Jamison, et al., 1975; and Bailey, et al., 1973). As the aromatic hydrocarbons are mobilized by dissolution from the soil or sediment, they are likely to be rapidly degraded as long as dissolved oxygen and sufficient microorganisms are available.

Fuel-related VOC contaminants at AOC 69W are expected to be reduced through volatilization, biodegradation, and/or dilution and dispersion.

SVOCs. Soil samples collected at or below the water table at AOC 69W contained fuel-related SVOCs (see Table 7-12). Acenaphthene, fluorene, naphthalene, 2-methylnaphthalene, phenanthrene, and phthalates are the primary SVOCs detected in groundwater samples from AOC 69W (Table 7-15 and 7-16).

Dissolution of SVOCs from unsaturated zone soil via infiltrating precipitation may be a probable transport mechanism for unpaved areas at AOC 69W due to the sandy nature of the overburden soils.

Harding Lawson Associates

W003971.080

Volatilization is a minor transport mechanism for SVOCs in the soils and groundwater at AOC 69W. The fuel-related SVOCs at AOC 69W, such as naphthalene and phenanthrene, are considered moderately volatile, and therefore volatilization is not as significant a transport mechanism as it is for VOCs.

Dissolved phase transport of SVOCs in groundwater is considered a significant transport mechanism at AOC 69W. Factors affecting partitioning of SVOCs from soil to groundwater include solubility and K_{∞} . SVOCs are generally regarded as immobile because of strong adsorption to the organic carbon fraction of soil predicted through higher $K_{\infty}s$ and low solubilities (Tinsley, 1979; Kenaga and Goring, 1978). SVOCs with moderate solubilities and moderate to high $K_{\infty}s$, such as pyrene and phenanthrene, will partition slightly to groundwater from the saturated zone soils (see Table 8-1). Results of saturated zone soil samples and groundwater samples indicate that pyrene was not detected in groundwater; however, phenanthrene with a higher solubility than pyrene was detected in groundwater. Processes that tend to attenuate migration of impacted groundwater at AOC 69W include retardation resulting from sorption, volatilization, and degradation.

Biodegradation reactions act to reduce the total mass of lower molecular weight PAHs (e.g., naphthalene). Naturally occurring soil microorganisms capable of degrading aromatic hydrocarbons have been studied, and a relationship between dissolved oxygen and biodegradation has been documented (Jamison, et al., 1975; and Bailey, et al., 1973). As the aromatic hydrocarbons are mobilized from the soil by groundwater movement, they are likely to be degraded as long as dissolved oxygen and sufficient microorganisms are available.

The fate of fuel-related SVOC contaminants at AOC 69W is expected to be reduced through volatilization, biodegradation, and/or dilution and dispersion. The slow rate of migration (due to partitioning to soil) for the PAHs allows for significant degradation, even if degradation rates are small, before they can travel significant distances. The fuel-related PAHs also tend to be more persistent with increasing molecular weight.

Chlorinated pesticides (e.g., DDD, DDE, DDT, and dieldrin) form a small and diverse group of synthetic chemicals characterized by cyclic structures with variable numbers of chlorine atoms. Pesticides generally have low solubilities and high partition coefficients and, therefore, are usually immobile within the soil or sediment

Harding Lawson Associates

(See Table 7-1). Most pesticides are resistant to degradation in the environment (Moore and Ramamoorthy, 1984); however, the breakdown process is relatively rapid after the halogen bond is broken. DDT converts rapidly to DDD under anaerobic conditions. DDE is the major degradation product of DDT under aerobic conditions and is believed to be produced primarily by chemical processes (Moore and Ramamoorthy, 1984). The distribution of pesticides within Willow Brook sediments suggests that they are not site related contaminants.

Inorganics. Inorganics detected at AOC 69W include metals (aluminum and lead), transition metals (iron, manganese, vanadium, chromium, cobalt, nickel, zinc, and copper), alkaline earth metals (calcium, magnesium, and barium), alkali metals (sodium and potassium), and nonmetallic elements (arsenic).

The mobility of inorganics in soil-water systems is strongly affected by compound solubility, pH, soil cation exchange capacity, soil type, oxidation-reduction potential, adsorption processes, major ion concentrations, and salinity. The distribution of inorganics would most likely be controlled by adsorption processes. Once adsorbed to soil, the inorganics may migrate with the soil by mechanical transport of particles. The migration of dissolved inorganics is dependent upon their individual adsorption characteristics (Oak Ridge National Laboratories, 1989). Mobilities of inorganic elements relative to the redox state of the environment are presented in Table 8-2. Elevated levels of iron and manganese in 69W-94-10 and 69W-94-13 is attributed to the reductive dissolution of dissolved oxygen by the aerobic declaration of fuel related contaminants. This is supported by observed in-situ measurements of the oxidation reduction potential (Appendix G).

8.3 SITE CONCEPTUAL MODEL

Figure 8-1 presents a simplified site conceptual model flow chart encompassing the essential features of AOC 69W showing the potential source and transport mechanisms for the contaminants detected at AOC 69W. The model reflects the current understanding of the site with respect to sources of contamination, the distribution of contamination, and the potential migration pathways.

Harding Lawson Associates

W003971.080

Based on the results of the RI, the primary site-related contaminants at AOC 69W are fuel-related contaminants in soil and groundwater. VOCs, SVOCs, pesticides, PCBs, and TPHC were detected during the investigation.

Based on the results of the field investigation, it appears that the primary contaminant source was contaminated surface and near surface soils located in the vicinity of the new boiler room. The soil contamination is due to the 1972 fuel oil release from ruptured piping. The 1978 fuel oil release from piping in the vicinity of the old boiler room does not appear to have been a significant source of soil or groundwater contamination.

The primary release mechanism appears to be infiltration into groundwater from source area contaminants above the water table. Potential secondary release mechanism is the contaminated soil downgradient of the source areas. The contaminated soil downgradient of the source areas is believed to be due to sorption of dissolved phase contaminants as well as transport via the underground pipe associated with the "skimmer system".

Harding Lawson Associates

9.0 BASELINE HUMAN HEALTH RISK ASSESSMENT

9.1 OVERVIEW

Human health and ecological risk assessments have been conducted to evaluate potential risks to humans and the environment under current or foreseeable future site conditions at AOC 69W. The methods used to perform the risk assessments are consistent with relevant national and regional USEPA risk assessment guidance (e.g., Risk Assessment Guidance for Superfund (USEPA, 1989a); EPA New England Risk Updates (USEPA, 1992a; 1994a; 1995; 1996) and incorporate data from the various remedial investigation and removal action sampling activities at AOC 69W. The results of the baseline human health risk assessment are presented in Subsection 9.1, and the results of the ecological risk assessment are presented in Subsection 9.2.

Summary of Site History

AOC 69W is composed of the Former Fort Devens Elementary School (Building 215), which is not currently in use, and a parking lot and adjacent lawn located on the northern side of the school. The site is less than 2 acres, and extends from the school to Willow Brook, which is located approximately 250 feet to the northwest (Figure 9-1).

The Fort Devens Elementary School was originally built in 1951, and comprised the east/southeast half of the existing school. The original school was heated by an oil-fired boiler, and the heating oil was stored in a 10,000-gallon UST located in what is currently the courtyard. In 1972 an addition to the school was built, which approximately doubled the school's size and created the existing courtyard.

During construction of the addition, the original 10,000-gallon UST was removed, and a new 10,000-gallon UST was installed north of the school in the middle of the current parking lot. Underground piping was installed to connect the new tank to a new boiler room in the addition. At the same time, oil lines were installed from the new boiler room to the original boiler room, so that the original boiler room could stay in use. An oil return line was also installed between the old boiler room and the new tank.

Harding Lawson Associates

W003971.080

SECTION 9

According to a former school employee, the company that installed the new underground storage tank crimped the pipe leading to the new boiler room, causing the pipe to split and leak a reported 7,000 to 8,000 gallons of fuel to the ground over a period of a few weeks in Spring 1972. When the problem was discovered, the pipe was unearthed and repaired. The exact location of the pipe split has not been determined. A subsurface oil recovery system was installed shortly after the piping was repaired. It is believed to have consisted of a pipe buried about three feet below grade, running from near the existing UST to a 250-gallon concrete vault located approximately 300 feet northwest of the school (Figure 9-1). Oil apparently flowed through the pipe to the vault, where it was periodically removed. Details of the collection system have not been found, although the presence of the pipe and vault have been confirmed by geophysical and visual means.

In 1978, another leak reportedly occurred from a failed subsurface piping joint near the old boiler room. While investigating the leak, a large hole was dug adjacent to the school between the old boiler room and the UST (near monitoring well ZWM-95-16X, Figure 9-1). Fuel oil was reportedly periodically removed from this hole for a period of a month while the leak was located and repaired.

Based upon documentation of the fuel oil releases, the site was designated AREE 69W and investigations were performed in 1994. The results of the AREE 69W investigation (A.D. Little, 1994) indicated the presence of fuel-related contaminants in both soil and groundwater between the school and the existing UST, and in the area extending from the existing UST northwest to near Willow Brook. Based on the results from the AREE 69W investigation, the Site was redesignated AOC 69W and was recommended for the RI/FS Process as per the IAG.

Site Conceptual Model

The RI identified fuel oil-related contamination adjacent to the school foundation near the loading dock and New Boiler Room, extending to the vault located approximately 300 feet northwest of the school near Willow Brook (Figure 9-1). Figure 9-2 graphically presents the site conceptual model, which relates sources of fuel-related contamination to migration pathways and the environmental media which human and ecological receptors may potentially be exposed to. The model is based on the findings of the RI.

Harding Lawson Associates

As indicated in Figure 9-2, fuel oil was released to the surface and subsurface soils. The fuel oil contamination, which was quantified in this RI by measuring EPH and VPH fractions along with their associated target analytes (i.e., BTEX and PAHs), contained EPH and VPH as well as low levels of VOCs and PAHs, a finding that is consistent with weathering fuel oil. Inorganic analytes and several other semivolatile organic analytes were detected in site media. These analytes are evaluated in the risk assessments, but do not appear to be associated with any releases from the Site. The lighter EPH and VPH hydrocarbon chains, as well as VOCs and soluble inorganics, can leach from the soil and migrate to groundwater.

The RI identified petroleum-related contamination in groundwater at AOC 69W. Groundwater at AOC 69W is approximately five to nine feet bgs, and flows toward Willow Brook. When the groundwater level is high, such as in the spring, the groundwater is believed to discharge to Willow Brook. This suggests that there is the potential for petroleum-related contamination in the groundwater to discharge to the Once in the brook, the petroleum-related compounds may adsorb to brook. sediment. Petroleum-related compounds were detected in Willow Brook sediment. However, there are several possible upgradient sources of these compounds that are unrelated to AOC 69W, so the presence of these compounds in the stretch of Willow Brook near AOC 69W is not necessarily related to conditions at AOC 69W. Although the groundwater beneath AOC 69W is not used as a source of drinking water, and is not expected to be in the future, it flows in the general direction of the MacPherson groundwater supply well, which is located approximately 3,000 feet downgradient of AOC 69W. However, the contamination in the groundwater at AOC 69W does not effect the same aquifer that the MacPherson well is screened in. and with the completion of the recent soil removal action, is not expected to migrate substantially further (see Section 8.0).

The lighter EPH and VPH fractions, as well as VOCs, can volatilize from the soil or groundwater to ambient air. In addition, these compounds, if present beneath the Former Devens Elementary School, can theoretically volatilize and migrate to the indoor air at the school. Indoor air sampling performed at the school building identified target compounds in the indoor air. However, with the possible exception of ethylbenzene, xylene, and 2-methylheptane in samples collected in the northwestern portion of the building, there was no evidence to suggest that constituents detected in indoor air were associated with the possible presence of fuel oil-contaminated soil and/or groundwater beneath the building (Appendix O-1). In

Harding Lawson Associates

general, indoor air sample concentrations were within the ranges of aliphatic and aromatic hydrocarbon background concentrations published in the literature. This information, combined with the presence of several possible sources of these compounds inside the building, suggested that detections in indoor air samples are merely representative of the ambient conditions in public buildings such as the Former Elementary School. The presence of ethylbenzene, xylene, and 2methylheptane in air samples collected from rooms in the northwestern portion of the school is likely also related to sources within the building (e.g., carpeting, paint, adhesives, and fuel oil leaks in the new boiler room), as detected concentrations are consistent with the ranges of typical background concentrations. However, because these constituents were detected in indoor air, crawl space air, and subsurface soil or groundwater beneath the that portion of the building, the possible presence of a complete migration pathway could not be ruled out (Appendix O-1).

According to the Devens Reuse Plan (Vanasse Hangen Brustlin, Inc., 1994), the site is designated for reuse as "Open Space" in the vicinity of Willow Brook and the delineated wetlands and "Gateway" for the remainder of AOC 69W. The Reuse Plan also includes reopening the school. Under the present and anticipated future land use conditions, people (e.g., trespassers, maintenance workers, pupils, and faculty) could be exposed to fuel-related contamination in soil, groundwater discharge to surface water in Willow Brook, sediment in Willow Brook, and indoor air. Ecological receptors could be exposed to fuel-related contamination in groundwater discharge to surface water Willow Brook and sediment in Willow Brook.

To mitigate possible exposures and possible continued migration of fuel-related compounds from the soil source areas, the Army conducted a soil removal action in winter 1997-1998. The removal action involved excavation and off-site disposal of cubic yards of contaminated soil between the school foundation and the vault near Willow Brook were removed in the winter of 1997 and 1998. The human health and ecological risk assessments are based on post-removal action soil conditions.

9.2 HUMAN HEALTH RISK ASSESSMENT

The human health risk assessment for AOC 69W was performed using methods consistent with USEPA Region I risk assessment guidance, including USEPA New England Risk Updates (USEPA, 1992a; 1994a; 1995; 1996), Risk Assessment

Harding Lawson Associates

Guidance for Superfund (USEPA, 1989a), Standard Default Exposure Factors (USEPA, 1991), and Dermal Exposure Assessment: Principles and Applications (USEPA, 1992d). In addition, although USEPA Region I does not require quantitative evaluation petroleum-related constituents (i.e., as measured by the total petroleum hydrocarbon parameter), this risk assessment provides a quantitative evaluation of petroleum-related constituents using the methods described in the MADEP Petroleum Policy (MADEP, 1997). Quantitative evaluation of petroleum-related constituents using the methods described in the Source of contamination at the site is associated with historic fuel oil releases.

The human health risk assessment for AOC 69W consists of the following components:

- Selection of Chemicals of Potential Concern (CPCs)
- Exposure Assessment
- Toxicity Assessment
- Risk Characterization
- Uncertainty Evaluation
- Summary and Conclusions

Each of these components is discussed in the following subsections.

9.2.1 Selection of Chemicals of Potential Concern

The first step in the risk assessment involves compiling and evaluating the analytical site data to identify those chemicals present in environmental media as a result of potential sources at AOC 69W. Site-related chemicals that were selected for risk evaluation are referred to as CPCs.

9.2.1.1 Identification and Selection of Analytical Data. Samples were collected at AOC 69W from surface and subsurface soil, sediment, groundwater, and indoor air. The sampling and analytical programs are discussed in Section 5.0

Soil

Surface and subsurface soil samples were collected throughout the source area during the RI (Figure 9-1). For the human health risk assessment, surface soil was defined

Harding Lawson Associates

W003971.080

SECTION 9

as extending from 0 to 1 foot bgs, and subsurface soil was defined as extending from 1 to 10 feet bgs (USEPA, 1995). Deeper soil is considered isolated and not accessible. Soil samples from many of the Terraprobe locations were submitted for screening-level analysis of VOCs and TPH using on-site field analysis. Samples from select locations and depths were submitted for off-site laboratory analysis of VOCs, SVOCs, pesticides/PCBs, inorganics, and TPH. Only the off-site analytical data were used in this risk assessment because the off-site analyses were performed in accordance with more rigorous QA/QC protocols, and were associated with lower analytical detection limits.

However, the soil associated with off-site analytical samples from several locations in the source area were removed during the soil excavation performed in the fall of 1997. Those soil locations include: ZWR-95-30X (0 ft bgs), ZWR-95-38X (3 ft bgs), ZWB-95-02X (0 and 5 ft bgs), ZWR-95-36X (0 ft bgs), and ZWB-95-01X (7 ft bgs and shallower). These samples, therefore, were not included in the risk assessment data sets. However, 31 confirmatory samples were collected during the removal action from the floor and walls of the soil excavation and analyzed for EPH/VPH. These samples were incorporated in the risk assessment data set. Although confirmatory samples collected from an excavation could be combined and reported as a single analytical result, the excavation area represents a large portion of the site. Therefore, each confirmatory sample collected from the excavation was evaluated as a unique and separate sample.

Table 9-1 lists the sample locations included in the surface soil and subsurface soil data sets. The data sets include RI and removal action confirmatory samples collected in and around the source areas, and from the boiler room and courtyard borings at the school. Sample ZWR-95-26X (0 ft bgs), which was collected from beneath the pavement adjacent to the school, was included in the subsurface soil data set because the soil associated with this sample would have accessibility similar to that assumed for subsurface soil.

Groundwater

At AOC 69W, several rounds of groundwater data have been collected:

ADL AREE 69W Rounds 1 and 2, April 1994 and June 1994

Harding Lawson Associates

- HLA AOC 69W RI Round 1, October 1995
- HLA AOC 69W RI Round 2, February 1996
- HLA AOC 69W RI Round 3, October 1996
- HLA AOC 69W RI Round 4, December 1997

The groundwater monitoring well locations are shown in Figure 9-1. Since all historical AREE groundwater wells at and downgradient of the Site were re-sampled by HLA at least once, the AREE groundwater data were not used in this risk assessment. Groundwater samples from HLA Rounds 1 and 2 were analyzed for VOCs, SVOCs, pesticides/PCBs, unfiltered and filtered inorganics, and TPH. In addition, groundwater samples from Round 2 were analyzed for EPH/VPH and associated target analytes. Groundwater samples from Rounds 3 and 4 were analyzed for EPH/VPH and associated target analytes. With the exception of well 69W-95-17X, which is an upgradient location, all groundwater locations at AOC 69W were included in the risk assessment data set. This includes groundwater monitoring wells 69W-94-10, 69W-94-11, and 69W-94-13, which were removed during the soil remediation. Table 9-1 lists the samples included in the groundwater data set.

Although there are slight temporal variations in the magnitude of groundwater concentrations associated with data for HLA rounds 1 through 4, there are no substantial differences in the identity of detected analytes or trends in the direction of groundwater concentrations. To provide a conservative assessment that considers the slight temporal variations in groundwater concentrations, unfiltered groundwater data from all four HLA sampling rounds were used for the risk assessment. Because the Round 4 groundwater sampling was performed just prior to the soil removal action, the groundwater data do not reflect the beneficial effects that the soil source removal action will have on the groundwater concentrations. Groundwater concentrations in the future are expected to show a downward trend as a result of the soil source area removal action. Therefore, the groundwater data used in this risk assessment reflect the highest groundwater concentrations that are anticipated under current and future land use conditions.

Since EPH/VPH was analyzed in at least one sample from each groundwater monitoring well location, EPH/VPH data were used in the risk assessment in lieu

Harding Lawson Associates

SECTION 9

of the TPH data collected during Rounds 1 and 2. This is a preferable approach because it allows for a site-specific assessment of the petroleum-related constituents, and is consistent with the MADEP petroleum policy (MADEP, 1997). In addition, because select VOCs and PAHs were analyzed and reported under both SVOC and EPH/VPH methods, multiple results were available for these compounds. To maintain a conservative assessment, the data associated with the analytical method that reported the highest detected concentration of each target PAH or VOC analyte were used in the risk assessment. The analytical methods upon which the data for VOCs and SVOCs are based are indicated in Table 9-2.

Surface Water

No surface water data were collected in Willow Brook. The portion of the brook that could potentially be affected by releases from AOC 69W is normally dry, and appears to only contain surface water during storm events or when the groundwater level is at its most shallow depth during late winter and early spring. During this time, groundwater located upgradient of Willow Brook, including the groundwater from beneath AOC 69W, discharges to Willow Brook. Therefore, to provide a conservative assessment of surface water in Willow Brook, data for the groundwater beneath AOC 69W was used to represent theoretical surface water concentrations. To provide an added degree of conservatism, no modeling or dilution of the groundwater discharged was assumed, as described in detail in Subsection 9.1.2.2.

Sediment

Sediment samples were collected at six locations in Willow Brook (Figure 9-1). The samples were analyzed for VOCs, SVOCs, pesticides/PCBs, inorganics, and TPH. The sediment samples ZWD-95-01X, ZWD-95-04X, and ZWD-95-05X were collected upgradient of AOC 69W. These samples represent media that have not been affected by potential releases from AOC 69W, and were used in the RI to evaluate the significance of constituents detected in downgradient samples. The upgradient samples are therefore not included in the risk assessment data sets. The samples included in the risk assessment data set are provided in Table 9-1.

Indoor Air

Air sampling data were collected from sampling locations inside the school, and

Harding Lawson Associates

background locations outside the school in October, 1997. The indoor air samples were analyzed for target aliphatic and aromatic hydrocarbon VOCs (Appendix M). As described in Appendix O-1, because 2-methylheptane, ethylbenzene, and xylenes were detected in air and soil samples collected from the northwestern portion of the building, a potentially complete migration pathway could not be ruled out. The procedures used to evaluate and summarize data and to screen data for the selection of CPCs are discussed below.

9.2.1.2 Data Summary Procedures. Prior to selecting CPCs, the analytical data were grouped into the data sets listed on Table 9-1. The following steps, which are in accordance with USEPA (1989a; 1992c) guidance, were used to summarize the analytical data for this risk assessment:

Data quality was evaluated by validating the data in accordance with USEPA data quality assessment procedures (USEPA, 1992c). Data suitable for use in risk assessment (i.e., those not rejected) were used in the risk assessment. Although data assessment procedures specified in RAGs (USEPA, 1989) were followed, data qualified as estimated and blank-contaminated were used in the risk assessment; uncertainties that may affect the risk assessment results are discussed in Subsection 9.1.5. Several TICs were detected in surface and subsurface soil, groundwater, and sediment, but they are not included in this evaluation because they were few in number and low in concentration compared to identified analytes. A summary of TICs is provided in Table 7-1. The data quality assessment is provided in Appendix D.

Data were summarized by environmental medium (for example, surface soil, groundwater). All chemicals detected in at least one sample in each data set were identified.

For groundwater, analytical results for the multiple HLA RI sampling rounds at a single location were averaged together and evaluated as a single point.

The arithmetic mean concentration was calculated for each chemical using the detected concentration(s), and one-half the sample quantitation limit (SQL) for nondetect(s). Duplicate samples for a given sampling point were also averaged in this manner if a chemical was detected in only one sample of a duplicate pair. Groundwater data were also treated in this manner if a chemical was not detected

Harding Lawson Associates

in all sampling rounds at a given well.

Frequency of detection was calculated as the number of samples in which the chemical was detected over the total number of samples analyzed. Duplicate pairs and data for multiple rounds at a single well were counted as a single result for calculation of the frequency of detection.

The minimum and maximum sample quantitation limits were identified for each analyte in each data set.

95 percent upper confidence levels on the arithmetic mean concentration were calculated in accordance with "Supplemental Guidance to RAGS: Calculating the Concentration Term" (USEPA, 1992b), assuming a log-normal distribution. This guidance states that data sets with fewer than ten samples provide poor estimates of the true mean, with the upper confidence limit (UCL) frequently being greater than the highest measured concentration. Therefore, the 95 percent UCLs on the arithmetic means were not calculated for data sets with fewer than 10 samples; 95 percent UCLs were calculated only for SVOCs, EPH, and VPH in subsurface soil.

Summary sampling data for the surface soil, subsurface soil, groundwater, sediment, and indoor air data sets for AOC 69W are presented in Table 9-2. The table lists frequency of detection, range of SQLs, minimum and maximum detected concentrations, arithmetic mean concentration, and 95 percent UCL for each chemical detected (when applicable).

9.2.1.3 Data Screening Procedures. The procedures used for selection of CPCs, based on USEPA (1989a; 1995) guidance, are described below. The results, including reasons for selection or exclusion of CPCs, are presented in Table 9-2.

The summary data for soil and groundwater were compared to concentrations of site-specific naturally occurring inorganic analytes. The development of the background data sets for soil and groundwater are described in Appendix K; no background data were available for sediment. In accordance with USEPA Region I guidance, the comparison was not used to eliminate any analytes; however, maximum concentrations of an analyte that were below background levels was noted on Table 9-2. Risks attributable to background concentrations

Harding Lawson Associates

will be discussed in the uncertainty section if inorganic analytes are determined to be risk drivers.

A screening process was conducted as described by USEPA Region I (USEPA, 1995) to focus the quantitative assessment on the most likely risk drivers. As specified by the guidance, maximum concentrations of chemicals detected in a medium were compared to RBCs derived from USEPA Region III's RBC Table (USEPA, 1997b). Soil concentrations were compared to residential soil RBCs, groundwater concentrations were compared to tap water RBCs, and indoor air concentrations were compared to tap water RBCs. The RBCs published by USEPA Region III are derived for a 1x10⁻⁶ cancer risk level or a non-cancer hazard quotient (HQ) of 1. Per USEPA Region I guidance (USEPA, 1995), the RBCs based on noncarcinogenic effects have been adjusted for a HQ of 0.1 for the purposes of CPC selection. If the maximum concentration of an analyte exceeded the appropriate RBC, the analyte was retained as a CPC.

Analytes were also compared to ARARs. No contaminant was eliminated as a CPC if it exceeded an ARAR. In this case, the ARARs used included the USEPA screening level for lead in soil (USEPA, 1994b), and USEPA MCLs, secondary MCLs, and action limits for groundwater (USEPA, 1997a).

Essential nutrients were eliminated from the risk assessment because they are unlikely to result in adverse effects at low concentrations. Chemicals considered to be essential human nutrients are calcium, magnesium, potassium, and sodium. Iron is also considered a human essential nutrient, but a USEPA Region III screening value is available for it.

RBCs and ARARs were not available for TPHC, EPH, or VPH. These contaminants were therefore considered CPCs.

CPCs were selected in surface soil, subsurface soil, groundwater, sediment, and indoor air (Table 9-2). The CPCs selected for each medium are discussed below.

Harding Lawson Associates

W003971.080

<u>Surface Soil</u>. CPCs selected in surface soil included arsenic, beryllium, iron, manganese, and total petroleum hydrocarbons. Among these CPCs, the maximum detected concentrations of arsenic, iron, and manganese did not exceed the background concentrations, suggesting that their presence in AOC 69W surface soil is not attributable to releases at the Site.

<u>Subsurface Soil</u>. CPCs selected in subsurface soil included arsenic, iron, total petroleum hydrocarbons, EPH fractions C11-C22 aromatics, C19-C36 aliphatics, C9-C18 aliphatics, and VPH fractions C9-C12 aliphatics, and C9-C10 aromatics. The maximum detected concentrations of arsenic and iron did not exceed the background concentrations, suggesting that their presence in AOC 69W subsurface soil is not attributable to releases at the Site.

<u>Groundwater</u>. CPCs selected in unfiltered groundwater included aluminum, arsenic, iron, manganese 2-methylnaphthalene, naphthalene, bis(2-ethylhexyl)phthalate, chloroform, trichloroethylene, EPH fractions C9-C18 aliphatics, C11-C22 aromatics, and VPH fractions C9-C12 aliphatics, and C9-C10 aromatics, and C5-C8 aliphatics. Among these CPCs, only petroleum-related constituents (2-methylnaphthalene, naphthalene, EPH fractions, and VPH fractions) are interpreted to be related to the releases of fuel oil at AOC 69W.

<u>Sediment</u>. CPCs selected in downgradient sediment include arsenic, iron, manganese, and total petroleum hydrocarbons.

<u>Indoor Air</u>. No CPCs were selected in indoor air. As indicated in Table 9-2, the maximum detected concentrations of xylene, ethylbenzene, and 2-methylheptane were below the RBCs.

9.2.2 Exposure Assessment

The purpose of the exposure assessment is to identify the potential pathways by which human populations may be exposed to CPCs at AOC 69W, and to make quantitative estimations of those exposures. The following subsections describe the exposure assessment for the human health risk assessment at AOC 69W.

9.2.2.1 Exposure Pathways. An exposure pathway generally consists of four elements:

Harding Lawson Associates

W003971.080

- (1) A source and mechanism of chemical release to the environment;
- (2) A retention or transport medium for the released chemical;
- (3) A point of potential human contact with the impacted medium (i.e., the exposure point); and
- (4) A route of exposure (ingestion or dermal contact, for example) for a potential receptor.

When all four of these elements are present, an exposure pathway is considered "complete." In the risk assessment, only exposure pathways that are complete under present land use, or potentially complete under the anticipated future land use are evaluated. The exposure pathway assessment for AOC 69W draws on information regarding the source of CPCs, fate and transport of CPCs, and information on human populations potentially exposed to CPCs in environmental media. The conceptual site model presented in Section 9.0 provides an overview of the potentially complete migration pathways at AOC 69W. These migration pathways are evaluated in the context of current and anticipated future land use to identify potentially exposed populations, exposure media, and exposure routes to those exposure media.

Current and Anticipated Future Site Use

In evaluating potential human exposure pathways, exposures under both current and potential future site uses and surrounding land use conditions were evaluated. Current land use conditions were evaluated to consider actual or possible exposures. Future land use conditions were considered to address exposures that may occur as a result of possible reuse of the site.

The site consists of the Former Fort Devens Elementary School (Building 215), a paved parking lot, and an adjacent lawn located on the northern side of the school; Willow Brook is located in a wooded wetland area approximately 200 feet northwest of the school (Figure 9-1). Since the school is closed, the site is not used. The only activity at the site is associated with maintaining the grounds (e.g., lawn mowing) during the summer months. Groundwater beneath AOC 69W is not used as a source of potable or industrial water; the school obtains its water supply from the Devens municipal groundwater supply. Willow Brook does not contain standing water most

Harding Lawson Associates

W003971.080

SECTION 9

of the year. This feature, accompanied by the wooded nature of the area, indicate that it does not offer a good place for swimming or wading. Although Willow Brook is not ideal for swimming or wading, evaluation of pathways involving Willow Brook are considered in the exposure pathway and scenarios discussions.

The future site and surrounding land use conditions at AOC 69W are anticipated to change. The school will likely be re-opened in the near future as a Charter school which will include elementary, junior high, and high school grade levels. Under these land use conditions, pupils, faculty, and school maintenance workers will occupy the school. In addition, excavation workers associated with construction and/or utility installation and repair may access the Site. Groundwater beneath AOC 69W is not anticipated to be used as a source of potable or industrial water. However, the MacPherson groundwater supply well located approximately 3,000 feet downgradient of AOC 69W is used as a source of potable water.

Exposure Pathways and Scenarios

Possible exposure pathways encompassing both current and future conditions are summarized in Table 9-3 and discussed below.

<u>Soil Pathway</u>. Under both current and future land use, contact with surface soil could occur if a site maintenance worker mows the grassy area, or a child trespasser (under current use) or pupil (under future use) plays in the grassy, unpaved area. Contact with subsurface soil could occur if excavation activities such as utility line repair or building construction were to take place. If pavement currently in place was removed in the future, what is now subsurface soil would be exposed to potential receptors. However, it is likely that any paving removed would be replaced by buildings, new paving, or topsoil for landscaping. Therefore, child trespassers, pupils, and site maintenance workers were selected as the most reasonable receptors to surface soil. A excavation worker performing utility or construction was selected as the most reasonable receptor to subsurface soil.

The two principal routes by which the site maintenance worker, child trespasser, pupil, and utility/construction worker could be exposed to CPCs in surface soil are ingestion and dermal contact. An additional exposure route, the inhalation of compounds volatilizing from the soil or of soil particles entrained in air (dust), is discussed under the Air Pathway below.

Harding Lawson Associates

<u>Groundwater Pathway</u>. There is no current use of groundwater at AOC 69W. There are existing supplies of water in this area of Devens and it is unlikely that the aquifer at the site would be developed for future use. Groundwater flow is towards Willow Brook and the wetland area. Willow Brook is shallow in the area of AOC 69W, and a child (trespasser or pupil) who plays in the stream could be exposed to CPCs in surface water. The two principal routes by which the child could be exposed to the groundwater discharge to surface water are incidental ingestion and dermal contact. Inhalation of VOCs that may volatilize from surface water is considered insignificant since VOCs were detected at very low frequency and concentration in groundwater.

Although excavation workers could theoretically contact groundwater if they dig excavations to the depth of the groundwater table (5 to 9 ft bgs), excavation workers would wear protective clothing (e.g., boots and gloves) in anticipation of this. Therefore, direct contact with groundwater is not evaluated for construction workers.

Residential use of the shallow groundwater beneath AOC 69W is highly unlikely. However, because this groundwater flows in the general direction of the MacPherson water supply well that is located 3,000 feet downgradient and the site is located within the Defined Zone II, residential use of the groundwater as potable water is evaluated. Groundwater under the site is within a Zone II groundwater protection area and a potentially productive aquifer, as defined by the MADEP.

<u>Sediment Pathway</u>. Under both current and future land use, contact with sediment could occur if a child was to wade in Willow Brook or the wetland area. Therefore, a child trespasser and pupil were selected as the most reasonable receptors to CPCs in sediment. The two principal routes by which these receptors could be exposed to CPCs in sediment are incidental ingestion and dermal contact. An additional potential exposure route, the inhalation of compounds volatilizing or particulates eroding from the sediment, is not considered to be a significant route of exposure because VOCs were detected at very low frequency and concentration, and dust is unlikely to be generated from sediment.

<u>Air Pathway</u>. Air could be a contact medium if VOCs present in the soil or groundwater volatilize into indoor or ambient air, or if airborne particulates containing CPCs are generated from the soil. With the exception of VPH in subsurface soil and groundwater, VOCs were detected at low frequency and concentration in soil and groundwater. Volatile inhalation exposures to a site worker

Harding Lawson Associates

W003971.080

10.1

mowing the grass are unlikely to be substantial given the considerable dilution of VOCs that would occur in the ambient air. However, excavation workers who are in closer contact with subsurface soil and groundwater and potentially working in a semi-closed "trench-type" excavation, may incur more substantial exposures. Therefore, volatile inhalation exposures are evaluated for the excavation worker.

Inhalation of soil contaminants entrained on dust generated during site maintenance activities such as mowing the grassy area or a large-scale construction project is of concern for the site maintenance worker and the excavation worker. It is unlikely that a child trespasser or pupil who plays in the grassy area would generate substantial amounts of dust because the grassy area near the school is small and there are recreational ball fields located on the other side of the school for active recreational activities. However, to provide a conservation evaluation, dust inhalation exposures are evaluated for the trespasser and pupil.

There is a potential for the release of VOCs from soil or groundwater to indoor air. Adults (teachers or staff members) and pupils who may occupy the building in the future could be exposed via inhalation of the indoor air. This exposure pathway was evaluated in Appendix O-1 to identify areas of the building where the presence of constituents detected in indoor air could be attributable to migration from subsurface soil and/or groundwater beneath the school. In summary, the analysis presented in Appendix O-1 provided evidence that the presence of the constituents detected in air throughout the building is likely attributable to ambient and anthropogenic sources inside and outside the building, and is not attributable to subsurface fuel oil contamination beneath the building. Although detections of 2-methylheptane, xylenes, and ethylbenzene in air samples collected from the library and an adjacent classroom are likely attributable to sources within the building (e.g., carpeting, paint, adhesives, and fuel oil leaks in the new boiler room), a possible association between the presence of these constituents in air samples from these rooms and the soil in the vicinity of that area of the building could not be ruled out. Although these three analytes may be the only indoor air contaminants potentially associated with subsurface fuel oil contamination, all analytes detected in indoor air samples were evaluated in the risk assessment in order to provide a conservative assessment of potential exposures.

9.2.2.2 Estimation of Exposure. To quantitatively estimate the magnitude of exposures and the risks that may be experienced by an individual, the representative

Harding Lawson Associates

W003971.080

concentration of the CPC in the contact medium at each exposure point must be known or estimated. This concentration is referred to as an exposure point concentration (EPC). The EPC is then combined with receptor-specific variables which describe the magnitude with which the receptor comes into contact with the exposure medium to develop quantitative estimates of exposure.

Exposure Points

There is one exposure point for surface soil, subsurface soil, sediment, surface water (represented by groundwater discharge), and groundwater. The boundaries of each exposure point are defined by the locations of the samples included in each of the media, as listed on Table 9-1 and shown in Figure 9-2. For indoor air, each room sampled represents a seperate exposure point. The crawl space beneath the cafeteria (kitchen and the boiler room wall head) were not considered exposure points, since people would not be expected to occur in these areas.

Exposure Point Concentrations

The EPCs for surface soil, subsurface soil, and sediment are the lesser of the 95 percent UCL on the arithmetic mean concentration or the maximum detected concentration (USEPA, 1992b; 1994a). Because there were less than ten samples included in the data sets for the surface soil and sediment exposure points, 95 percent UCLs were not calculated; the EPCs for these two media are the maximum detected concentrations. The same EPC was used for the reasonable maximum exposure (RME) and CT (CT) exposure scenarios (USEPA, 1995). The 95 percent UCL concentrations and maximum detected concentrations are provided on Table 9-2.

For groundwater potable use, USEPA Region I guidance states that the groundwater EPC for RME conditions is the maximum concentration, whereas the EPC for CT exposures is the arithmetic mean concentration (USEPA, 1995).

As discussed previously, no surface water was present during the RI sampling program and, therefore, no surface water samples were collected from Willow Brook. The arithmatic mean groundwater concentrations are used as EPCs to provide conservative estimates of possible surface water concentrations. Using the arithmetic mean groundwater concentrations to estimate exposure to surface water (as

Harding Lawson Associates

groundwater discharge) overestimates the potential exposure for two reasons: 1) the groundwater contamination is primarily associated with areas near the school, and it is unlikely that these concentrations will ever reach the brook; and 2) groundwater becomes substantially diluted when it enters the brook.

For indoor air, each room sampled represents an individual exposure point. However, to streamline the risk assessment and provide a more conservative estimate of risks, the maximum concentration among all samples collected (i.e., all potential exposure points) was used as the EPC.

Exposure Estimates

Quantitative exposure estimates were derived by combining the EPCs with information describing the extent, frequency, and duration of exposure for the specific receptor (Table 9-4). An overview of the approaches used to quantify exposures is given below, along with specific details for potential exposure pathways. The approaches described in the following paragraphs to quantify exposures are consistent with guidance provided by USEPA (1989a; 1989b; 1991; 1992d; 1994a; 1995).

The term RME is defined as the maximum exposure that is reasonably expected to occur at a site (USEPA, 1989). The RME is intended to place a conservative upperbound on the potential risks, meaning that the risk estimate is unlikely to be underestimated but it may very well be overestimated. The likelihood that this RME scenario may actually occur is small, due to the combination of conservative assumptions incorporated into the scenario. The RME estimate for a given pathway is derived by combining the EPC of each chemical with reasonable maximum values describing the extent, frequency, and duration of exposure (USEPA, 1994a). The CT (CT) estimate combines the EPC with CT exposure parameters. Many of the exposure parameter values used in this assessment have been defined by USEPA (1989a; 1989b; 1991; 1994a). Both CT and RME exposures were estimated for each receptor exposure scenario evaluated.

The general equation for calculating chemical intake is as follows:

Intake = $\frac{C \times CR \times RAF \times EF \times ED}{BW \times AT \times CF}$

Harding Lawson Associates

W003971.080

9-18

where:

Intake	=	daily intake averaged over the exposure period
С	÷	(mg/k/day) concentration of the chemical in the exposure medium (mass/volume)
CR	=	contact rate for the medium of concern (mass/day)
RAF	=	relative absorption factor (unitless)
EF	=	exposure frequency (days/year)
ED	=	exposure duration (years)
BW	=	body weight of the hypothetically exposed individual (kg)
AT	÷	averaging time (for carcinogens, AT = 70 years; for noncarcinogens, AT = ED)
CF	÷	units conversion factor (365 days/yr)

Specific equations for each exposure scenario are provided in the risk calculation spreadsheets in Appendix O-2. For dermal uptake of soil, sediment, and surface water CPCs, USEPA Region I has developed the following guidance (USEPA, 1995):

<u>Soil/Sediment</u>. USEPA Region I uses the USEPA dermal exposure guidance (USEPA, 1992d), which provides quantitative dermal absorption factors for PCBs, dioxins, and cadmium. None of these chemicals are CPCs at AOC 69W. MADEP risk assessment guidance (MADEP, 1995) requires quantitative evaluation of all CPCs using dermal relative absorption factors published by the MADEP. To satisfy MADEP concerns, dermal exposures to soil/sediment CPCs were evaluated using MADEP (1995) guidance.

<u>Surface Water</u>. Using the approach identified in the USEPA dermal exposure guidance (USEPA, 1992d), the dermal route was evaluated quantitatively using dermal permeability coefficients. For inorganics, a steady-state approach was used, whereas for organic CPCs, a nonsteady-state approach was used. The nonsteady-state approach is applicable to organics that exhibit octanol/water partitioning and is not appropriate for inorganics. The equations and factors used to develop permeability coefficients for each identified CPC in groundwater are listed in Table 9-5.

Standard exposure parameters from USEPA guidance were used, where appropriate, to estimate CPC intake. Table 9-4 presents the parameters used in each exposure

Harding Lawson Associates

scenario. The parameters are discussed briefly below.

<u>Contact Rate</u>. The contact rate reflects the amount of the medium contacted per unit of time or event. For incidental ingestion of soil, the high-end contact rate is 100 mg soil per day for the site maintenance worker and child trespasser and pupil (USEPA, 1994a). The CT value for these receptors is 50 mg/day (USEPA, 1994a). For incidental ingestion of soil, the high-end contact rate is 480 mg soil per day for the excavation worker (USEPA, 1994a). A fraction-ingested variable can be used to account for the amount of soil ingested daily that is assumed to come from the area of concern. It is conservatively assumed that all soil ingested daily would originate at AOC 69W. The contact rate variables used for soil were also used for sediment.

For incidental ingestion of surface water, it was conservatively assumed that the highend contact rate for the child would be the same contact rate presented for residential exposure via ingestion of chemicals in surface water while swimming, 0.05 liter per hour (USEPA, 1989b). A contact rate of half the high-end contact rate, 0.025 liter per hour, was assumed for the CT exposure.

For ingestion of groundwater used as residential drinking water, the ingestion rate for the RME is 2 liters per day, and 1.4 liters per day for the CT (USEPA, 1994a).

The contact rate for inhalation exposure is 2.5 cubic meters (m^3) of air per hour for the site maintenance worker (USEPA, 1991, USEPA, 1989b). The inhalation rate is 4.8 m³ of air per hour or 38.4 m³ per workday for the utility/construction worker, based on a heavy activity rate (USEPA, 1989b).

<u>Relative Absorption Factor</u>. The relative absorption factor represents the ratio of a chemical's bioavailability (i.e., ability to be absorbed and potentially exert an effect) in an environmental matrix to its bioavailability when administered in the experimental dose-response study from which the toxicity criterion for that chemical was derived. The relative bioavailability factor is applied to account for the potentially reduced bioavailability of a chemical when ingested or dermally absorbed in a soil matrix, compared to when experimentally administered in a food mash, water, or a solvent medium. In keeping with the conservative nature of this assessment, a relative oral bioavailability of 100 percent (or 1.0) is assumed for all CPCs. Dermal relative absorption factors were obtained from MADEP sources (MADEP, 1994; 1995), and are presented in the risk calculation tables in Appendix

Harding Lawson Associates

0-2.

Exposure Frequency and Duration. An exposure rate of two times per week for 32 weeks (April-November), equal to 64 days per year for 25 years, is assumed for the site maintenance worker. It is assumed that the maintenance worker spends 8 hours per day on the days working at the Site. Contact with the soil is negligible when the ground is frozen or snow-covered. An exposure rate of less than five days per week accounts for work performed in other areas of the facility; it is unlikely that the small grassy area adjacent to the building would require activities even two-days per week.

An exposure rate of three days per week for 32 weeks (April-November), equal to 96 days per year, is assumed for the child trespasser exposure to surface soil. Pupil exposure to surface soil is estimated to be five days per week from September to November, and April to mid-June, and three days per week, mid-June to late August (when school is not in session), to result in 140 days/year total exposure. Contact with the soil is negligible during winter months when the ground is frozen or snow-covered. For the trespasser and pupil, exposure to surface water and sediment is assumed to occur two-days per week during the summer (June-August, or 24 days/year). This represents a conservative estimate of exposure since surface water is generally only present in the stretch of Willow Brook near AOC 69W after storm events or when the groundwater level is high during the early spring months. For both the pupil and the trespasser, it is assumed that exposure to surface water and sediment occurs 2.6 hours per day on the days-exposed. This value is based on the average time for swimming (USEPA, 1989a).

A trespasser ages 6 through 18 is assumed to occur at the site for 13 years (equivalent to the duration of age 6 through 18) for the RME scenario, and 9 years for the CT scenario. The CT exposure duration is based on the average length of residence at a single location. Since the Charter school will include grades kindergarten through high school, the ages of the pupils will span 6 through 18. Therefore, the RME scenario is based on a 13 year exposure duration. The central tendency exposure duration is 9 years, and is based on the average length of residence at a single location. The pupil is assumed to occur in the school building six hours per day, for 180 school days per year.

The excavation worker is assumed to work five days a week for 18 weeks, or 90 days over the course of a year.

Harding Lawson Associates

SECTION 9

<u>Body Weight</u>. The body weight used for the excavation worker and the site maintenance worker is 70 kilograms (kg), which represents the standard default value for adult body weight (USEPA, 1994a). The age-adjusted body weight for the child trespasser and pupil, each assumed to be 6 through 18 years of age, is 45 kg (USEPA, 1989b).

<u>Surface Area</u>. The body surface area potentially exposed to surface water while wading in Willow Brook is 5,053 cm³, based on the age-adjusted surface area of arms, hands, feet, and lower legs of children ages 6 through 18 (USEPA, 1989b). This value is used as a conservative estimate for exposures to surface soil. The body surface area of maintenance and excavation works is 5,200 cm³, based on the surface area of arms, hands, head, and neck of adult males.

<u>Averaging Time</u>. The averaging time for lifetime exposure, used for developing intake to evaluate carcinogenic risk, is 70 years. Averaging time for noncarcinogenic risk is equal to the exposure duration (USEPA, 1991).

9.2.3 Toxicity Assessment

The objective of the toxicity assessment is to define the relationship between the dose of a substance and the likelihood that a toxic effect, either carcinogenic or noncarcinogenic, will result from exposure to that substance. This is performed by identifying the potential adverse health effects associated with exposure to a substance, and then quantifying those dose-response relationships. Potential adverse health effects associated with exposure to a substance in short toxicity profiles provided in this risk assessment are summarized in short toxicity profiles provided in Appendix O-3. Dose-response information for the CPCs, which is used in the risk assessment to estimate the likelihood of adverse effects as a function of human exposure to the CPCs, is presented in Tables 9-6 through 9-10.

There are two types of dose-response values used in this risk assessment: cancer slope factors (CSFs) and reference doses (RfDs). USEPA has derived CSFs and RfDs to evaluate carcinogenic risks and noncarcinogenic (systemic) effects, respectively. The definitions of CSFs and RfDs, as stated in USEPA guidance are:

 Cancer Slope Factor - a plausible upper bound estimate of the probability of a response per unit intake of a chemical over a lifetime.

Harding Lawson Associates

W003971.080

The CSF is used to estimate an upper-bound probability of an individual developing cancer as a result of a lifetime exposure to a particular concentration of a potential carcinogen (USEPA Class A or B carcinogens) (USEPA, 1989a).

Chronic Reference Dose - an estimate of a daily exposure level for the human population, including sensitive subpopulations, that is likely to be without an appreciable risk of deleterious effects during a lifetime (e.g., as a Superfund program guideline, seven years to a lifetime) (USEPA, 1989a). The chronic RfD is used to estimate toxicity to all receptors occurring at a site for more than seven years.

 Subchronic Reference Dose - an estimate of a daily exposure level for the human population, including sensitive subpopulations, that is likely to be without an appreciable risk of deleterious effects during a portion of a lifetime (e.g., as a Superfund program guideline, two weeks to seven years) (USEPA, 1989a). The subchronic RfD is used to estimate toxicity to all receptors occurring at a site for between twoweeks and seven years (e.g., excavation worker).

In addition, because the toxicity and/or carcinogenicity of a compound can depend on the route of exposure (e.g., oral or inhalation), unique dose-response values (e.g., CSFs and RfDs) have been developed for the oral and inhalation exposure routes.

The primary source for dose-response values is the USEPA Integrated Risk Information System (IRIS) (USEPA, 1998). If no information is found in IRIS, the USEPA Health Effects Assessment Summary Tables (HEAST) (USEPA, 1997c) are used. If appropriate dose-response values are not available from either of these two sources, other USEPA sources are consulted (e.g., the USEPA National Center for Environmental Assessment).

Inhalation RfDs are not available for CPCs. HEAST and IRIS do, however, list reference concentrations for a number of the CPCs. Risks for these CPCs were evaluated by developing an average daily air concentration and calculating a hazard index (HI) instead of using an inhalation dose to calculate a dose-based HI. The equation for the average daily air concentration, shown on the spreadsheets in Appendix O, is the air concentration multiplied by the exposure time and the

Harding Lawson Associates

duration, divided by the averaging time.

The methodology used to develop dermal toxicity values is obtained from Risk Assessment Guidance for Superfund, Appendix A (USEPA, 1989a). The oral toxicity value is adjusted from administered dose to absorbed dose, if necessary. The gastric absorption efficiency of a particular compound is used to calculate an RfD based on absorbed dose. For example, if the RfD based on administered dose was 20 mg/kg/day, and the gastric absorption efficiency in the study that is the basis of the RfD was 10 percent, then: 20 mg/kg/day x 0.10 = 2 mg/kg/day. Therefore, the adjusted RfD is 2 mg/kg/day. This adjusted value is the dermal reference dose (RfDderm). Similarly, the dermal cancer slope factor (CSFderm) is obtained by adjusting the oral CSF. For example: if the CSF based on administered dose was 1.6 (mg/kg/day)-1, and the gastric absorption efficiency in the study that is the basis of the CSF is 20 percent, then: 1.6 (mg/kg/day)-1/0.20 = 8 (mg/kg/day)-1. This adjusted value is the CSFderm.

The oral (or in some cases, inhalation) absorption efficiency for individual compounds is obtained from IRIS, HEAST, Agency for Toxic Substances and Disease Registry (ATSDR) toxicity profiles, or independent studies. If the gastric absorption efficiency is not available from these sources, the efficiency is assumed to be similar to structurally similar compounds. If a similar compound is not identified, an assumed value for the class of compound (VOCs, SVOCs or inorganic analytes) is used. Based on best professional judgment, the default gastric absorption efficiency values for VOCs, are assumed to be 80 percent; for SVOCs, 20 percent; and for inorganics, 5 percent. If the RfD or the CSF is based on absorbed dose rather than administered dose, no adjustment for gastric absorption efficiency is necessary. Dermal dose-response data are presented in Table 9-10.

No USEPA-approved toxicity values are available for TPHC, EPH, or VPH. However, the "Characterizing Risks posed by Petroleum Contaminated Sites: Implementation of MADEP VPH/EPH Approach" (MADEP, 1997) recommends the application of Reference Doses for indicator compounds to various fractions of petroleum products analyzed for by the EPH and VPH methods as a means of assessing risks associated with media that contain petroleum. These RfDs were used to estimate the risks associated with each EPH/VPH fraction. When there is no information available concerning the concentrations of the various fractions comprising the petroleum, but there are data for TPHC, a quantitative evaluation

Harding Lawson Associates

was completed using the RfD for pyrene as a surrogate.

9.2.4 Risk Characterization

In this final step of the risk assessment process, the exposure and toxicity information are integrated to develop both quantitative and qualitative evaluations of risk. To quantitatively assess risks associated with CPCs in an environmental medium, the average daily intakes calculated in the Exposure Assessment are combined with the dose-response criteria presented in the Toxicity Assessment. The methodology used to quantitatively assess risks is described below.

9.2.4.1 Risk Characterization Methods. USEPA (1989a) has developed guidance for assessing the potential risks to individuals from exposure to carcinogenic and noncarcinogenic chemicals.

Cancer Risks

For exposures to a chemical exhibiting carcinogenic effects, an individual upper bound excess lifetime cancer risk (ELCR) is calculated by multiplying the estimated daily CPC intake by the relevant CSF:

ELCR = Intake $(mg/kg/day) \times CSF (mg/kg/day)$ -1

The resulting risk estimate is an upper-bound estimate of the probability of developing cancer as a result of exposure to the potential carcinogen over a 70-year lifetime under the specified exposure conditions. A risk level of 1x10-6, for example, represents an upper bound probability of one in one million that an individual will develop cancer. The upper bound cancer risk estimates provide estimates of the upper limits of risk, and the risk estimates produced are likely to be greater than the 99th percentile of risks faced by actual receptors (USEPA 1989a). This incremental lifetime risk is over and above what is considered an individual's background chances of developing cancer. In the U.S., approximately one in three people develop cancer during their lifetime (American Cancer Society, 1997). To assess the upper bound individual ELCRs associated with simultaneous exposure to all carcinogenic chemicals of concern, the risks derived from the individual chemicals are summed within each exposure pathway. This approach is consistent with the USEPA's guidelines for evaluating the toxic effects of chemical mixtures (USEPA 1989b). It

Harding Lawson Associates

SECTION 9

will overestimate risk if maximum concentrations occur in different locations and they are used as EPCs.

The relative significance of carcinogenic risk estimates is evaluated by comparison to a target risk range of 1x10-4 to 1x10-6 established in the National Contingency Plan (USEPA, 1989). USEPA's guidelines state that when the total incremental carcinogenic risk for an individual resulting from exposure at a hazardous waste site is within the range of 1x10-4 to 1x10-6, the decision about whether to take action or not is based on site-specific factors.

Non-Cancer Risks

Unlike carcinogenic effects, noncarcinogenic effects are not expressed as incidence probabilities. Rather, potential noncarcinogenic impacts were evaluated by means of calculating HQs and HIs. To assess effects associated with noncarcinogenic exposures, the ratio of the daily intake to the RfD is calculated by dividing the intake for each noncarcinogenic CPC by the RfD for that CPC to derive an HQ:

HQ = Intake (mg/kg/day) / RfD (mg/kg/day)

In general, HQs that are less than 1 indicate that the associated exposure is not likely to result in any adverse health effects, whereas HQs greater than 1 indicate a greater likelihood of adverse health effects. The effects from simultaneous exposures to all CPCs were computed by summing the individual HQs within each exposure pathway. This sum, the HI, serves the same function for exposures to multiple CPCs as the HQ does for exposure to an individual compound.

HIs for both RME and CT exposures are compared to a target level of 1 established by USEPA. HIs greater than 1 indicate the potential for the occurrence of adverse health effects. However, a conclusion should not be categorically drawn that all HIs greater than 1 are unacceptable. If the individual CPCs effect different target organs or work through different mechanisms of action, then an HI of greater than 1 does not indicate that threshold effect levels have been reached. In cases where an HI is greater than 1, the CPCs may be segregated by target organ/critical effect (e.g., liver, skin, etc.) and the HQs for each group of CPCs may be summed to determine if the target organ-specific HI is greater than 1 (USEPA, 1989a).

Harding Lawson Associates

W003971.080

9.2.4.2 Risk Characterization Results. The risk calculations for soil and groundwater are presented in Appendix O-2. The risk estimates are summarized in Table 9-11, and are discussed below.

Current Land Use

<u>Surface Soil</u>. Health risks associated with possible current exposures to surface soil were quantified for the site maintenance worker and the child trespasser (ages 6 through 18). Cancer risks for CT and RME scenarios were below or within the USEPA acceptable cancer risk range, with ELCRs ranging between $1x10^{-6}$ (maintenance worker CT) to $6x10^{-6}$ (trespasser RME). Non-cancer risks were below the USEPA threshold HI of 1.

Because the RME and CT cancer and noncancer risks are within or below the USEPA acceptable limits, surface soil does not pose an unacceptable health risk under current land use conditions.

<u>Sediment</u>. Health risks associated with possible exposure to sediment were quantified for the child trespasser (ages 6 through 18). Cancer risks for CT and RME scenarios were each at or below 1×10^{-6} , indicating that they are below the USEPA cancer risk range and are therefore acceptable. The HIs for the CT and RME scenarios were each 0.1, which are below the USEPA threshold HI of 1. This indicates that non-cancer risks are acceptable for current land use conditions.

<u>Groundwater Discharge</u>. Health risks associated with possible exposure to groundwater discharge to surface water were quantified for the child trespasser (ages 6 through 18). Cancer risks for CT and RME scenarios were 1×10^{-6} and 2×10^{-6} , respectively. These values are within the USEPA acceptable cancer risk range. The HIs for the CT and RME scenarios were each 0.2. These values do not exceed the USEPA threshold HI of 1. This evaluation, therefore, indicates that cancer and non-cancer risks for possible exposures to groundwater discharge to surface water are acceptable for current land use conditions.

<u>Cumulative Risks.</u> The cumulative risks to the trespasser for combined exposures to surface soil, sediment, and groundwater discharge to surface water are within the USEPA acceptable cancer risk range, and do not exceed a HI of 1 (Table 9-2). This indicates that multi-media exposures to CPCs at AOC 69W do not pose unacceptable

Harding Lawson Associates

risk for current land use.

Future Land Use

<u>Surface Soil</u>. Health risks associated with possible future exposures to surface soil were quantified for the pupil (ages 6 through 18) and the excavation worker. Cancer risks for the pupil CT and RME scenarios were within the USEPA acceptable cancer risk range, with ELCRs ranging between $5x10^{-6}$ (CT) and $9x10^{-6}$ (RME). Non-cancer risks were below the USEPA threshold HI of 1.

The cancer and non-cancer risks for the excavation worker were below the USEPA acceptable cancer risk range and below an HI of 1.

Because the RME and CT cancer and noncancer risks are within or below the USEPA acceptable limits, surface soil does not pose an unacceptable health risk under the anticipated future land use conditions.

<u>Subsurface Soil</u>. The total cancer risk to the excavation worker, assuming exposure to RME conditions for subsurface soil (one to ten feet bgs) is $1x10^{-7}$, which is below the USEPA acceptable cancer risk range. The hazard index is 0.9, which is below the USEPA threshold HI of 1. This indicates that subsurface soil does not pose an unacceptable cancer or non-cancer risk under possible future land use conditions.

Because an excavation worker would be potentially exposed to both surface soil and subsurface soil when digging, the cancer and non-cancer risks for surface soil and subsurface soil were added together. The total cancer risk is 4x10-7, and the total non-cancer risk is an HI of 1. These values do not exceed USEPA threshold risk levels, and therefore indicate no unacceptable risk.

<u>Sediment</u>. Health risks associated with possible exposure to sediment were quantified for the pupil (ages 6 through 18). Cancer risks for CT and RME scenarios were each at or below $1x10^{-6}$, indicating that they are below the USEPA cancer risk range, and are therefore acceptable. The HIs for the CT and RME scenarios were below 0.1, which are below the USEPA threshold HI of 1. This indicates that non-cancer risks are acceptable for future land use conditions.

Harding Lawson Associates

W003971.080

<u>Groundwater Discharge</u>. Health risks associated with possible exposure to groundwater discharge to surface water were quantified for the pupil (ages 6 through 18). Cancer risks for CT and RME scenarios were 1×10^{-6} and 2×10^{-6} , respectively. These values are within the USEPA acceptable cancer risk range. The HIs for the CT and RME scenarios were 0.2. These values are below the USEPA threshold HI of 1. This evaluation, therefore, indicates that cancer and non-cancer risks for possible exposures to groundwater discharge to surface water are acceptable for future land use conditions.

<u>Groundwater</u>. Health risks associated with possible future use of on-site groundwater as residential drinking water were evaluated. This evaluation differs from the evaluation of the future use of the school because the school re-opening is an anticipated future land use, whereas the use of groundwater as drinking water is not an anticipated, but rather a theoretical future land use. Since it is unlikely that the groundwater at AOC 69W will be used as a source of drinking water in the future, these risks represent a hypothetical scenario.

For unfiltered groundwater, the estimated RME cancer risk is $5x10^3$, which is above the USEPA acceptable cancer risk range. The estimated CT cancer risk is 1x10-4, which is at the upper bound of the USEPA acceptable cancer risk range. The RME non-cancer risk for the child and adults are an HI of 25 and 57, respectively. The CT non-cancer risks are an HI of 4 and 8, respectively. These HIs exceed the USEPA threshold HI of 1. This indicates that theoretical future of groundwater as drinking water at AOC 69W is associated with unacceptable risks.

However, greater than 99 percent of the cancer risk, and approximately 70 percent of non-cancer risk, is associated with possible exposures to arsenic in groundwater (Appendix O-2, Tables 27 through 30). The child resident RME HI for petroleum CPCs interpreted to be potentially related to the fuel oil release (i.e., EPH, VPH, and non-carcinogenic PAHs) is 5. The child CT HI for these CPCs is less than 1, which does not exceed the USEPA threshold HI of 1.

Based on the results of this risk assessment, it appears that theoretical use of groundwater as potable water is associated with unacceptable cancer and non-cancer risks. However, the unacceptable risks are primarily due to CPCs that are not directly related to the fuel oil release at AOC 69W.

Harding Lawson Associates

W003971.080

<u>Cumulative Risks.</u> The cumulative risks to the pupil for combined exposures to surface soil, sediment, groundwater discharge to surface water, and indoor air are within the USEPA acceptable cancer risk range, and do not exceed a HI of 1 (Table 9-2). This indicates that multi-media exposures to CPCs at AOC 69W do not pose unacceptable risk for future pupils who may attend the Charter School. Cumulative risks for the excavation worker (exposure to surface soil and subsurface soil) are also within acceptable limits, as discussed previously.

9.2.5 Evaluation of Uncertainty

The interpretation of risk estimates is subject to a number of uncertainties as a result of conservative assumptions inherent in risk assessment. All quantitative estimates of risk are based on numerous assumptions, most intended to be protective of human health (i.e., conservative). As such, risk estimates are not truly probabilistic estimates of risk, but rather conditional estimates given a series of conservative assumptions about exposure and toxicity.

In general, sources of uncertainty are categorized into site-specific factors (e.g., variability in analytical data, modeling results, and exposure parameter assumptions) and general factors that affect most risk assessments equally, such as toxicity information. Toxicity information for many chemicals is very limited, leading to varying degrees of uncertainty associated with calculated toxicity values. Sources of uncertainty for calculating toxicity factors include extrapolation from short-term to long-term exposures, amount of data (e.g., number of studies) supporting the toxicity factors, consistency of different studies for the same chemical, and responses of various species to equivalent doses. The general uncertainties that affect most risk assessments, and the direction of their potential effects on the risk assessment results (e.g., to over- or under-estimate risks) for AOC 69W are summarized in Table 9-12. Site-specific uncertainties that have the greatest potential effect on the results of this risk assessment are discussed below.

Data Evaluation and Selection of CPCs

Several analytes, most notably arsenic, iron, and manganese, in all media except groundwater were selected as CPCs even though their maximum concentrations did not exceed background concentrations. This suggests that the presence of these analytes are not directly associated with fuel oil releases at AOC 69W; including

Harding Lawson Associates

W003971.080

these analytes as CPCs results in an overestimation of risk for potential exposures to CPCs that are directly related to the fuel oil release at AOC 69W. In addition, the background concentration of arsenic in groundwater is associated with a potable use RME risk $(2x10^4)$ that exceeds the USEPA acceptable cancer risk range, and is greater than the AOC 69W central tendency risks for groundwater potable use $(1x10^4)$.

Use of the groundwater data for the analytical methods that reported the highest maximum detected concentrations represents a conservative approach that is likely to overestimate risk. Since SVOC data were collected in historical, and not recent sampling rounds, the groundwater concentrations may have decreased, and would be better represented by the more current EPH/VPH data. In addition, removal of the soil source area between the school and the vault is expected to result in a decrease in groundwater concentrations; the groundwater data used in the risk represent pre-remediation groundwater data used in the risk represent pre-remediation groundwater concentrations. Therefore, the groundwater data used in the risk assessment represent pre-remediation groundwater concentrations; the groundwater concentrations; the groundwater concentrations for potential current and future exposures to groundwater.

Use of the Region III RBCs in the CPC selection for indoor air represents a conservative approach. As described in Subsection 9.1.1.3, Region III RBCs for ambient air, adjusted for an HQ of 0.1, were used for CPC selection of constituents detected in indoor air (because RBCs were not available for 2-methylheptane, 3-methylheptane, nonane, and octane, a value equal to 10 percent of the RfC published by MADEP was used in place of the RBCs). The RBCs were developed using inhalation RfCs and exposure assumptions that are appropriate for chronic inhalation exposures to an adult (e.g., 24 hours per day, 365 days per year). However, staff members would only occupy the school for eight hours per day, and 250 days per year at the most. Therefore, the air concentrations which the staff members would potentially be exposed to would be: 250 days/365 days x 8 hours/24 hours, or 23 percent of the RBCs. For a child who would potentially occupy either the library for no more than two or three hours per day, and the classroom for no more than six hours per day 180 days per year, the actual exposure concentrations would be 6 percent to 12 percent of the RBCs.

Exposure Assessment

Harding Lawson Associates

W003971.080

In general, the exposure scenarios selected for evaluation in this risk assessment overestimate the frequency and magnitude of potential exposures. For example, it is unlikely that a child trespasser will wade in the wetland area two days per week June through August, particularly since surface water is at it's lowest depth during those months. Similarly, it is unlikely that a pupil will be exposed to the grassy area in front of the school five times per week, particularly since the school play grounds and athletic fields are elsewhere at the school. Therefore, the exposure scenarios chosen to represent possible exposures to Site media are unlikely to result in underestimation of risks.

There is some uncertainty associated with the EPCs for inorganics and non-fuel related VOCs in subsurface soil because only two subsurface soil samples were analyzed for these parameters. Additional samples might have shown higher (or lower) concentrations. However, non-fuel related VOCs and inorganics were not the primary chemicals of concern at AOC 69W (i.e., there presence in subsurface soil was not directly related to the fuel oil release at AOC 69W), and were not substantial risk drivers in subsurface soil.

The intended future use of the school is a Charter School which will include grades kindergarten through high school. The risk assessment is protective for pupils in these grades (i.e., ages 6 through 18). However, if the school was used only for younger pupils - grade-school aged children for example - risks would be underestimated because the body weight for younger children would be lower (e.g., 35 kg for children 6 through 13) than the body weight for children ages 6 through 18 (45 kg). Risks for grade-school aged children exposed to surface soil, groundwater discharge, and sediment would be higher than those estimated in this risk assessment by a factor of 1.3 (45 kg / 35 kg = 1.3). However, the combined risks for exposure to surface soil, groundwater discharge, sediment, and indoor air would still not exceed a HI of 1 or an ELCR of 1E-05. Therefore, although this risk assessment evaluated risks for children over a wide age range, risks for younger children would still not exceed threshold risk levels.

As discussed previously, groundwater beneath AOC 69W is not used as a source of drinking water. Although the groundwater flows toward a the MacPherson water supply well located 3,000 feet downgradient of the site, the groundwater contamination at AOC 69W does not affect that aquifer, and it does not appear that site-related fuel oil contamination will ever reach the well. Groundwater monitoring

Harding Lawson Associates

W003971.080

data have indicated that fuel-related compounds are contained in groundwater associated with soil source area, which was recently removed. However, monitoring data from downgradient wells at the site (e.g., ZWM-95-18) indicate that no fuel oil contamination has reached the wells since their installation in 1995. Since the soil source area has recently been removed, groundwater concentrations of fuel related constituents in the former source area are expected to decrease. This, in conjunction with the limited groundwater migration identified to date, suggest that it is extremely unlikely that groundwater containing fuel-related compounds from AOC 69W would migrate to the downgradient water supply well. Moreover, any compounds that did migrate downgradient would be substantially diluted, and would not likely pose a risk of concern.

All analytes detected in indoor air samples collected from rooms within the school building were evaluated in this risk assessment. This represents a conservative assessment of potential exposures to constituents interpreted to be site-related. As discussed in Appendix O-1, based on the exposure pathway analysis, it appears that the presence of only three analytes in indoor air (ethylbenzene, 2-methylheptane, and xylene) is potentially attributable to subsurface contamination beneath the school. Of these, none were detected in the vicinity of the northwestern portion of the school at concentrations that would include them as a CPC in the risk assessment, and only ethylbenzene was detected at a concentration within the school building at a concentration that included it as a CPC in the risk assessment.

The USEPA performed indoor air sampling in April, 1998; the USEPA report is included as Appendix R to this report. In general, the USEPA sampling identified a similar profile of chemicals in indoor air. However, analytes were generally detected at concentrations considerably lower than those detected in the RI sampling effort. Most notably, toluene was detected at only 2 μ g/m³ in the USEPA samples compared to 1000 μ g/m³ in the RI samples, ethylbenzene at only 0.5 μ g/m³ in the USEPA samples compared to 470 μ g/m³ in the RI samples; nonane and octane were not detected in the USEPA samples, whereas other fuel-related constituents (e.g., decane, dodecane, and trimethylbenzene) were detected at concentrations below 1 μ g/m³. Methylpentane was detected at 170 μ g/m³ in the USEPA samples, but was not detected in the RI samples. Methylnaphthalene and naphthalene, which were associated with rejected data in the RI samples, were detected at concentration up to 1.4 μ g/m³ in the USEPA samples.

Harding Lawson Associates

W003971.080

Given this information, it appears that the RI sampling effort provided a conservative assessment of the potential concentrations of indoor air constituents. Evaluation of the maximum detected concentrations in indoor air samples collected during the RI sampling effort as the EPCs in the risk assessment provided a conservative assessment that is unlikely to underestimate potential risks to future pupils that may use the school.

Toxicity Assessment

Arsenic was selected as a CPC detected in all media at AOC 69W. Use of the CSF for arsenic to estimate ELCRs is thought to overestimate the true risk by perhaps an order of magnitude or more (USEPA, 1998). The oral CSF for inorganic arsenic is based on dose/response data for skin cancer incidence obtained by Tseng et al. (1968). Individuals in this study were exposed to high levels of inorganic arsenic in drinking water (170 micrograms per milliliter $[\mu g/mL]$). Arsenic exposure was approximated based on estimates of water intake. Other exposure pathways contributing to total exposure, such as ingestion of fish, livestock, and plants, were not assessed, potentially resulting in an underestimate of arsenic exposure. The oral slope factor was calculated using a model that assumes the dose/response curve is linear at low doses. Recent evidence suggests that arsenic, at low doses, may be largely detoxified by methylation, producing a non-linear dose/response curve. In the study of Tseng et al. (1968), the overwhelming of the normal detoxification pathways, coupled with an underestimate of exposure, may have resulted in an overestimate of cancer risk. These uncertainties have caused the USEPA to report that, "the uncertainties associated with ingested inorganic arsenic are such that estimates could be modified downwards as much as an order of magnitude, relative to risk estimates associated with most other carcinogens" (USEPA, 1998).

Use of the USEPA provisional oral RfD for iron to estimate noncancer risk may result in an overestimate of risk. The oral RfD for iron is based on average daily intake data for humans from the NHANES II data base. The provisional RfD represents an average dietary intake of iron that is protective against iron deficiency anemia, but is insufficient to cause toxic effects. Therefore, the data do not represent a threshold for adverse effects due to over-exposure. Toxic effects would occur following chronic oral exposures to higher iron levels than those represented by the provisional oral RfD. In addition, the forms of iron administered to individuals reported in the NHANES II data base were bioavailable forms (ferrous

Harding Lawson Associates

form) contained in iron supplements. Iron in environmental media is typically present in the elemental or ferric form, which have substantially lower bioavailability than the ferrous form. These two uncertainties relative to the provisional iron RfD likely result in an overestimate of noncancer risk.

No inhalation dose-response values for acetone and arsenic were published in sources recognized by USEPA Region I (i.e., IRIS, HEAST, NCEA). Therefore, non-cancer inhalation risks were not quantified for these CPCs. However, MADEP has published RfCs or RfC surrogate values for these CPCs. The RME HQ associated with particulate inhalation exposure to arsenic for the excavation worker (the most-exposed receptor to particulate emissions) is 0.001, which does not add measurable risk to the excavation worker RME inhalation HI of 0.8. The RME HQ associated with inhalation exposure to acetone in indoor air for the pupil is 0.07, which increases the pupil RME inhalation HI from 0.4 to 0.5; the HI for combined media exposures remains 1.

Risk Characterization

Given the uncertainties discussed in this section, it appears that the risk estimates reported in this risk assessment overestimate risks rather than underestimate risks. The primary sources of uncertainty the lend to a general overestimation of risks include:

- I. Including in the risk assessment inorganic analytes that are detected in Site media at concentrations consistent with background
- II. Frequency and contact rate assumed for exposures to soil, surface water, and sediment
- III. Likelihood of groundwater use as drinking water
- IV. Conservativeness of the arsenic and iron dose-response values
- V. Evaluation of inhalation exposures to constituants that are unlikely to be present in indoor as a result of subsurface fuel-oil contamination.
- VI. In addition, the quantitative evaluation of total petroleum hydrocarbon data is not required by USEPA Region I, but has been performed in this risk assessment to estimate possible risks associated with fuel oil-related constituents in environmental media. Therefore, non-cancer risks associated with CPCs which are quantitatively evaluated under USEPA Region I guidance are less than those reported in this risk assessment.

Harding Lawson Associates

9.2.6 Summary and Conclusions

Possible health risks were evaluated for the current and anticipated future land uses at AOC 69W. Since the former elementary school at AOC 69W is presently closed, exposures and risks for current site use were evaluated for a site maintenance worker (possible exposure to surface soil), and a trespasser ages 6 through 18 (possible exposure to sediment and groundwater discharge to surface water). The possible health risks associated with future site use were evaluated assuming that the school will be re-opened, and included evaluation of a pupil ages 6 through 18 (possible exposure to surface soil, sediment, groundwater discharge to surface water, and indoor air), and an excavation worker (possible exposure to surface soil and subsurface soil). In addition, future use of the groundwater as a potable water source was evaluated. Since groundwater at and beneath AOC 69W is not used as a source of drinking or industrial water, evaluation of potable use represents a hypothetical worst-case evaluation of potential exposures and risks.

The risk assessment evaluated post-removal action conditions for surface soil and subsurface soil, and pre-removal action conditions for groundwater, sediment, and indoor air. CPCs were identified in surface soil, subsurface soil, sediment, groundwater, and indoor air and included metals, SVOCs, VOCs, and petroleum-related compounds including TPHCs, EPHs, VPHs, and PAHs. Among these CPCs, only the petroleum-related compounds are directly associated with the release of fuel-oil at AOC 69W.

Possible health risks were quantified for carcinogenic and non-carcinogenic effects, for both reasonable maximum and central tendency exposure assumptions. Estimated cancer and non-cancer risks associated with possible current and future land use exposures to surface soil, subsurface soil, sediment, groundwater discharge to surface water, and indoor air were within acceptable levels established by the USEPA. Risks for multi-media exposures for the site maintenance worker, child, trespasser, future pupil, and future excavation worker were within acceptable levels. Estimated cancer and non-cancer risks associated with hypothetical exposures to AOC 69W groundwater used as a residential drinking water source exceeded levels generally considered acceptable by USEPA. However, these risks are primarily due to the presence of arsenic in the groundwater, which is not interpreted to be directly related to the release of fuel oil at AOC 69W.

Harding Lawson Associates

Based on the conclusions of the risk assessment, there are no unacceptable human health risks associated with soils, sediment, groundwater discharge, or indoor air. Moreover, the soil removal action at AOC 69W significantly reduced fuel oil contamination in soil thereby mitigating possible exposures to fuel-related CPCs in soil, reducing the CPC concentrations in groundwater, and eliminating a possible source of fuel-related vapors. Therefore, the risk estimates presented in this risk assessment worst-case estimates that are unlikely to be exceeded under conceivable future land use conditions. The only risks that exceed USEPA thresholds are associated with hypothetical future use of the groundwater as a source of residential drinking water.

9.3 ECOLOGICAL RISK ASSESSMENT

This Baseline Ecological Risk Assessment (BERA) evaluates actual and potential adverse effects to ecological receptors associated with exposure to contamination from Area of Contamination (AOC) 69W at Devens, Massachusetts. Fuel spills originating from an underground storage tank and associated piping at AOC 69W (the former Fort Devens Elementary School) in the 1970's have resulted in the presence of petroleum-related contaminants in site media. This BERA utilizes surface soil, sediment, groundwater, and toxicity test data to evaluate potential risks to ecological receptors.

The BERA for AOC 69W was completed in accordance with current guidance materials for BERAs at Superfund sites including the following:

- Risk Assessment Guidance for Superfund, Environmental Evaluation Manual (USEPA, 1989a);
- Ecological Assessment of Hazardous Waste Sites, A Field and Laboratory Reference (USEPA, 1989b);
- Ecological Assessment of Superfund Sites, An Overview (USEPA, 1991a);
- Framework for Ecological Risk Assessment (USEPA, 1992a);
- USEPA Region I, New England "Risk Updates" (issued since 1992);
- Tri-Service Procedural Guidelines for Ecological Risk Assessments, Volumes I and II (Wentsel et. al., 1996); and
- Proposed Guidelines for Ecological Risk Assessment (USEPA, 1996);

Harding Lawson Associates

W003971.080

and

 Ecological Risk Assessment Guidance for Superfund: Process for Designing and Conducting Ecological Risk Assessments (USEPA, 1997).

Recent risk assessment guidance including the USEPA "Eco Update" bulletins (issued since 1991) and publications (e.g., Maughan, 1993; Suter, 1993) were also consulted.

Discussions of the general site history and layout are provided in Section 2.0, and are briefly summarized at the beginning of Section 9.0. A discussion of historical and current analytical data is provided in the contaminant assessment (Section 5.0). The AOC 69W BERA includes a Site Characterization (Subsection 9.2.1), Problem Formulation (Subsection 9.2.2), Hazard Assessment and Selection of CPCs (Subsection 9.2.3), Exposure Assessment (Subsection 9.2.4), Ecological Effects Assessment (Subsection 9.2.5), Risk Characterization (Subsection 9.2.6), Uncertainty Analysis (Subsection 9.2.7), and Summary (Subsection 9.2.8).

9.3.1 Site Characterization

AOC 69W is located at the intersection of Antietam Street and MacArthur Boulevard (Figure 5-3). The area south and west of AOC 69W is primarily suburban/residential. The Wherring Housing area is located directly north of the site and Shepley's Hill sits immediately behind the site to the east and northeast.

AOC 69W is comprised of the former Fort Devens Elementary School (Building #215), a parking lot and an adjacent lawn located on the northern side of the school, a man-made drainage ditch, and a persistent emergent wetland bordering Willow Brook (Figure 5-3). The site extends from the school, west to a man-made drainage ditch that transitions into Willow Brook, and northwest to a persistent emergent wetland approximately 250 feet away. The original school was built in 1951 and an addition was constructed in 1972. During construction of the addition, the old UST was removed and a new 10,000-gallon UST was installed north of the school in the middle of the current parking lot. In addition, a 250-gallon UST was located about 300 feet to the northeast of the school and about 30 feet in from the edge of the delineated bordering vegetated wetland. Both the 10,000-gallon and 250-gallon USTs have been removed from the site. The 250-gallon UST was part of the recovery system installed after 7,000 to 8,000 gallons of oil leaked from faulty pipes in 1972.

Harding Lawson Associates

A site investigation completed by ADL (1994) identified contamination in surface soil, subsurface soil and groundwater form the school out towards Willow Brook. The contamination consist of oil leaked from piping associated with the 10,000-gallon underground storage tank. Most of the contamination appears to be beneath the paved area located north of the school. Ecological receptors would not be exposed to contaminants beneath the paved area; however, potential risks to ecological receptors do exist from contaminated surface soil and sediment in the vicinity of Willow Brook and groundwater discharging to the drainage ditch and Willow Brook.

In October of 1995, HLA ecologists visited AOC 69W to characterize habitats associated with the drainage ditch and Willow Brook. In addition, ecological receptors that would likely use the site were identified for evaluation in the BERA. Habitat types in the ditch and brook were characterized based on plant species present, topography, and hydrology. Ecologists divided brook habitats into two portions - upper and lower. The upper portion begins at the outfall of the drainage ditch that goes under Antietam Street. The lower portion begins 300 to 350 yards downgradient of the culvert where the drainage ditch transitions into a small persistent emergent wetland. The following paragraphs summarize the results of the qualitative survey. More information regarding the ecological survey is provided in Appendix P.

9.3.1.1 Vegetative Cover. The area between the upper portion of the ditch and MacArthur Boulevard consists of mowed grass. The dominant tree species along the northwest side of the drainage ditch is Norway spruce (*Picea abies*). The southwest side of the ditch is dominated by red maple (*Acer rubrum*). Other tree and shrub species prevalent along the western side of the ditch include: slippery elm (*Ulmus rubra*), pin oak (*Quercus palustris*), shagbark hickory (*Carya ovata*), American chesnut (*Castanea dentata*), white mulberry (*Morus alba*), red-osier dogwood (*Cornus stolonifera*), arrow-wood (*Viburnum recognitum*), and swamp azalea (*Rhododendron viscosum*). Herbaceous species present along this portion of the site include: sedges (*Cyperus sp.*), rice-cutgrass (*Leersia oryzoides*), nightshade (*Solanum dulcamara*), boneset (*Eupatorium perfoliatum*), cow vetch (*Vicia cracca*), jewelweed (*Impatiens capensis*), common dodder (*Cuscuta gronovii*), buttercup (*Ranunculus sp.*), and cocklebur (*Xanthium strumarium*).

Harding Lawson Associates

W003971-080

9.3.1.2 Wetland and Drainage Ditch Habitat. The man-made ditch, which leads into Willow Brook, is primarily a stormwater and meltwater conveyance feature and is not likely to support aquatic life for most of the year, because it typically does not contain standing water. However, it is possible that groundwater seasonally discharges to the lower portion of the ditch and may provide habitat for juvenile amphibians and some macroinvertebrates. The ditch is approximately 3 to 5 feet wide with an average depth of 3 feet, and has steep abrupt banks. The bottom of the ditch contains remnants of asphalt underlain by a medium to coarse sand with little to no organic matter.

Approximately 300 to 350 yards below the culvert at Antietam Road, the drainage ditch turns slightly to the northwest. The area immediately adjacent to the northeast side of the drainage ditch transitions into a small triangular-shaped persistent emergent wetland, which is about 0.25 acres in size. The wetland does not contain standing water; however, it is likely that water is at or near the surface throughout much of the year. The wetland area is dominated by herbaceous species, and has an open canopy. Tree and shrub species in this wetland include red maple, white pine (*Pinus strobus*), gray birch (*Betula populifolia*), arrow-wood, highbush blueberry (*Vaccinium corymbosum*), and buckthorn (*Rhamnus frangula*). Herbaceous species reported include tussock sedge (*Carex stricta*), wooly sedge (*Scirpus cyperinus*), bluejoint (*Calamagrostis canadensis*), broad-leaved cattail (*Typha latifolia*), Joe-pye weed (*Eupatorium maculatum*), arrow-leaved tearthumb (*Polygonum sagittatum*), and hardhack (*Spiraea tomentosa*).

9.3.1.3 Wildlife Habitat Characterization. The wetland area and open mowed areas may provide habitat to various small mammals and birds, predatory mammals, and soil invertebrates. Small mammal receptors may include the meadow vole (*Microtus pennsylvanicus*), short-tailed shrew (*Blarina brevicauda*), and the white-footed mouse (*Peromyscus leucopus*). Small birds found at AOC 69W may include the American robin (*Turdus migratorius*), a habitat generalist typically found in forested areas, and the red-winged blackbird (*Agelaius phoeniceus*), found in riparian and field habitats. The fragmented habitat at AOC 69W may not be suitable for many predatory mammals and birds. However, the raccoon (*Procyon lotor*) may exist along the wooded and wetland areas of the site where water may be present.

9.3.1.4 Rare, Threatened, and Endangered Species. The presence or absence of rare and endangered flora and fauna at the site is reviewed in this subsection. Under

Harding Lawson Associates

W003971.080

contract to the U.S. Army Corps of Engineers, HLA developed a database of all flora and fauna known to seasonally or permanently occur at Fort Devens (ABB-ES, 1993a), with particular emphasis on rare and endangered biota. The Fort Devens biological and endangered species baseline study (BESBS) contains information from the Massachusetts Natural Heritage Program (MNHP) and the United States Fish and Wildlife Service (USFWS) regarding all rare and endangered species known to occur at Fort Devens; in addition, more information has been requested for recently documented occurrences. The BESBS has been checked for known occurrences of rare and endangered biota in the vicinity of AOC 69W; no state or federally listed rare and endangered species occur at AOC 69W. AOC 69W may provide suitable habitat for species that are listed by the state as species of special concern or are on the state watch list, including the spotted salamander (*Ambystoma maculata*), northern leopard frog (*Rana pipiens*), olive-sided flycatcher (*Contopus borealis*), eastern bluebird (*Sialia sialis*), and Vesper sparrow (*Pooecetes gramineus*).

According to the MNHP (MNHP, 1997), several species protected by the Massachusetts Endangered Species Act and other implementing regulations have been documented as occurring within 1 mile of AOC 69W. The actual occurrence of these species at the site is unknown. The eastern box turtle (*Terrapene carolina*) (special concern) may be found in the wooded portions of AOC 69W. The following species listed by the MNHP may be found in the upland sandy soils or disturbed portions of AOC 69W: Houghton's flatsedge (*Cyperus houghtonii*) (endangered), New England blazing star (*Liatris scariosa* var. *novae-angliae*) (special concern), and wild senna (*Senna hebecarpa*) (endangered).

9.3.2 Problem Formulation

Problem formulation is the initial step of the BERA process whereby receptors, exposure pathways, and the assessment and measurement endpoints are selected for evaluation.

9.3.2.1 Identification of Receptors. Mammals, birds, reptiles, amphibians, terrestrial plants, and terrestrial invertebrates are expected to be found in the terrestrial habitats of AOC 69W. Semi-terrestrial wildlife (including some mammals and birds) and aquatic receptors (including some macroinvertebrates and amphibians) are expected to inhabit areas around Willow Brook and in the wetland.

Harding Lawson Associates

9.3.2.2 Identification of Exposure Pathways. Exposure pathways are identified for four groups of ecological receptors (wildlife, terrestrial plants, terrestrial invertebrates, and aquatic receptors). The exposure pathway includes a source of contamination, potentially contaminated media, and an exposure route. The exposure pathways from the AOC 69W contaminant source to ecological receptors are depicted in the contaminant pathway model in Figure 9-3. All of the potential exposure pathways are shown in Figure 9-3; those pathways that are quantitatively evaluated in the AOC 69W BERA are indicated by shading. This limitation is necessary to focus the BERA on the pathways for which: (1) contaminant exposures are the highest and most likely to occur, and (2) there are adequate data pertaining to the receptors, contaminant exposures, and toxicity for completion of risk analyses. Exposure pathways evaluated include portions of food chains (e.g., surface soil \rightarrow primary consumer \rightarrow secondary consumer \rightarrow tertiary consumer), as well as other direct and indirect exposures.

The general classes of contaminants include petroleum related compounds associated with the fuel oil release. Potentially affected media include surface soil, subsurface soil, sediment, and groundwater. The likely ecological effects associated with this type of contamination are identified for terrestrial wildlife, terrestrial plants, terrestrial invertebrates, and aquatic receptors. Given the relatively low concentrations of contaminants detected in site media, it is unlikely that there would be any acute toxicological effects in ecological receptors at AOC 69W. The effects on terrestrial wildlife from exposure to fuel oil contamination may include reduced fecundity, reduced growth, reduced activity, and mortality. The effects on terrestrial plants may include decrease in growth and propagation. The effects on terrestrial invertebrates may include reduced cocoon production, decreased growth, and mortality. The effects on aquatic receptors may include reduced growth and reproduction, and behavioral and developmental effects.

<u>Terrestrial Wildlife</u>. The wildlife exposure routes that are believed to contribute the highest potential contaminant exposures include incidental ingestion of site media and ingestion of food items that have bioaccumulated contaminants from site media. Dermal exposures to wildlife are not evaluated in the AOC 69W BERA because there are few data relating dermal exposures to toxic responses in wildlife. Dermal exposure may be an ecologically significant exposure pathway for amphibians and for young, hairless mammals in subterranean dens; however, in general, an assumption is made that fur, feathers, or chitinous exoskeleton limit the transfer of contamination

Harding Lawson Associates

across the dermis. The reported spills occurred a number of years ago; consequently, in the sandy soils of this region, most deposited VOCs would have already volatized or leached. Therefore, the inhalation exposure pathway was not quantitatively evaluated for ecological receptors. Potential food chain exposures for amphibians and reptiles exist at AOC 69W, but are not evaluated due to a lack of data relating contaminant exposures to adverse responses for these taxa. This data gap is furthered discussed in the uncertainty section.

Wildlife and semi-terrestrial wildlife ingestion exposures from surface soil contaminants in the upland and wetland portions of AOC 69W are evaluated in the BERA.

<u>Terrestrial Plants and Invertebrates</u>. Terrestrial plants and soil invertebrates may be exposed to contamination in surface soil by direct contact, root uptake (plants), or ingestion (invertebrates) of soil.

Aquatic Receptors. Exposure pathways for aquatic receptors at AOC 69W include direct contact with and ingestion of sediment in the downstream portions of the ditch and Willow Brook. Although conservative, this exposure pathway was evaluated because the ditch and Willow Brook may support aquatic life at some time during the year (i.e., when water is present). However, it is more likely that sensitive aquatic receptors would not even be present in the ditch due to the sporadic presence of water in the ditch, and limited availability of habitat (i.e., organic matter and decomposing leaf litter) required by sensitive aquatic receptors to survive during periods of no water. In addition, aquatic receptors may be indirectly exposed to contaminants in groundwater as it discharges to the surface; consequently, potential risks to aquatic receptors from exposure to contaminants in groundwater are evaluated. Exposures to aquatic receptors from sediment in the upstream portions of the ditch were not evaluated because this area is upgradient of the contaminant source, and because the ditch does not provide habitat suitable for supporting aquatic life.

9.3.2.3 Identification of Endpoints. The assessment and measurement endpoints selected for the AOC 69W BERA are listed in Table 9-13. Assessment endpoints represent the ecological component to be protected, whereas the measurement endpoints provide an operational definition of the assessment endpoint. The assessment endpoint selected for this BERA is the survival and propagation of

Harding Lawson Associates

W003971.080

receptor populations at AOC 69W. The specific objectives of the AOC 69W BERA are to determine whether the chemical concentrations detected in surface soil, sediment and groundwater are likely to result in population decline of ecological species.

Measurements of actual toxicity and adverse effects to survival and growth were completed for two benthic invertebrates, the midge (*Chironomus tentans*) and amphipod (*Hyalella azteca*). Assays with these invertebrates were intended to provide more realistic, site-related measures of adverse impact than generic screening values. Site-specific toxicological data are not available for surface soil or groundwater; therefore, the measurement endpoints used to gauge the likelihood of population-level effects are toxicological benchmark values based on laboratorymeasured survival, growth, and reproductive effects.

9.3.3 Hazard Assessment and Selection of Contaminants of Potential Concern (CPCs)

The Hazard Assessment includes a review of analytical data and selection of CPCs. CPCs are the analytes detected in environmental media that are considered in the AOC 69W BERA and could present a potential risk to ecological receptors. The process for selecting CPCs is depicted in Figure 9-4.

Historical (i.e., data collected before 1995) surface soil, sediment, and groundwater data were not included in the AOC 69W BERA as they do not represent current conditions at the site. Only current analytical data (i.e., those data collected in 1995, 1996 or 1997) were utilized in the AOC 69W BERA. All of the analytical data are provided in Section 7.0 of the Remedial Investigation (RI) and in electronic format in Appendix L. All samples were analyzed for PAL metals, pesticides/PCBs, PAL SVOCs, PAL VOCs, TPHC, EPH, and VPH. Off-site laboratory analyses for PAL organics and inorganics were considered approximately equivalent to USEPA Level III quality data. TPHC data are not evaluated in this BERA because methods to relate wildlife exposures to TPHC with adverse responses are not available. In addition, wet chemistry data (e.g., alkalinity, chloride, nitrite, nitrogen, phosphate, sulfate, TDS, TSS, and hardness) are available for groundwater. As discussed in Subsection 3.2.6, analytical data for AOC 69W were evaluated to determine their validity for use in the BERA.

Harding Lawson Associates

The following data sets are evaluated in the AOC 69W BERA:

- Surface soil (a combined data set of upland and wetland soils),
- Downgradient sediment (collected from the bottom of the ditch), and
- Groundwater.

Analytes were not retained as CPCs if the maximum detected concentration is less than the background concentration. The background surface soil and groundwater data sets consist of chemical data gathered from locations designed to establish background concentrations of inorganic analytes for Group 1A sites. The values approximately represent the 68th percentile upper bound limits (the mean values plus one standard deviation) of these chemicals (ABB-ES, 1993b). Sediment data were screened against soil background concentrations since no background database exists for sediment and the upgradient sediment may be contaminated from road run-off. The upgradient sediment data set consists of data collected in 1995 at sampling stations ZWD-95-01X, ZWD-95-04X, and ZWD-95-05X and will be qualitatively compared to downgradient concentrations.

The essential nutrients (e.g., calcium, magnesium, potassium, and sodium) were excluded as CPCs for all media, and iron was excluded as a wildlife CPC for foodchain exposures of surface soils and sediment. These analytes are considered to be essential nutrients; evidence suggests that there is little potential for toxic effects resulting from over-exposure to these essential nutrients. The highly controlled physiological regulatory mechanisms of these inorganics suggest that there is little, if any, potential for bioaccumulation, and available toxicity data demonstrate that high dietary intakes of these nutrients are well-tolerated (NAS, 1977; National Research Council [NRC], 1982; 1984).

All analytes detected in surface soil, sediment, and groundwater are presented in tables that include the following summary statistics: frequency of detection, range of detection limits, range of detected concentrations, and background values. For those analytes that were retained as CPCs in the BERA, the following information is also provided: average of all concentrations, and RME and average exposure point concentrations. Ninety-fifth percent UCLs were not calculated for any data sets as there are fewer than 10 samples in all the data sets. For groundwater, there were 13 samples analyzed for nitrogen, phosphate, hardness, total dissolved solids, total suspended solids, and TPHC. However, these analytes were not quantitatively

Harding Lawson Associates

evaluated, and 95 percent UCLs are not presented. A discussion of how EPCs were calculated is provided in Subsection 9.2.4.1.

9.3.3.1 Surface Soil. Summary statistics for six surface soil samples (ZWS-95-35X, ZWS-95-39X, ZWS-95-42X, and ZWS-95-45X through ZWS-95-47X) collected at AOC 69W (shown in Figure 5-3 and 5-5) are presented in Table 9-14. All organic analytes including acetone, toluene, trichlorofluoromethane, xylenes, eight PAHs, and TPHC were retained as CPCs in the BERA. Eight of the 19 inorganic analytes detected in soil (beryllium, cobalt, copper, lead, mercury, nickel, selenium, and zinc) were retained as CPCs because their maximum detected concentrations exceed background and none are essential nutrients as discussed above.

9.3.3.2 Upgradient Sediment. Summary statistics for three upgradient sediment samples (ZWD-95-01X, ZWD-95-04X, and ZWD-95-05X) (shown in Figure 5-3) are presented in Table 9-15. Information from this table was used for a qualitative comparison with the downgradient sediment samples.

9.3.3.3 Downgradient Sediment. Summary statistics for three sediment samples (ZWD-95-02X, ZWD-95-03X, and ZWD-95-06X) (shown in Figure 5-3) are presented in Table 9-16. All organic analytes including trichlorofluoromethane, 4 PAHs, the DDT_R family (i.e., DDD, DDE, and DDT), TPHC (gas fraction), diesel fuel, and TPHC (total) were retained as CPCs in the BERA. Three of 16 inorganic analytes (cobalt, copper, and nickel) detected in downgradient sediment were retained as CPCs because their maximum detected concentrations exceed background concentrations and they are not essential nutrients. TOC concentrations were low (ranging from 2,400 to 5,400 μ g/g).

9.3.3.4 Groundwater. Summary statistics for ten groundwater samples (69W-94-10 through 69W-94-14, ZWM-95-15X, ZWM-95-16X, and ZWM-95-18X through ZWM-95-20X) collected at AOC 69W (shown in Figure 5-3) are presented in Table 9-17. All organic analytes including three VOCs, seven SVOCs, two pesticides, and TPHC were retained as CPCs. Three inorganic analytes (arsenic, iron, and manganese) detected in filtered and unfiltered groundwater samples were retained as CPCs because their maximum detected concentrations exceeded background concentrations.

9.3.4 Exposure Assessment

Harding Lawson Associates

Exposure assessment is the process of estimating or measuring the amount of a CPC to which an ecological receptor may be exposed. The following sections briefly describe how contaminant exposures were estimated or measured for wildlife, terrestrial plants, terrestrial soil invertebrates, and aquatic receptors at AOC 69W. The contaminant pathway model (Figure 9-3) provides a summary of the potential exposure pathways that exist at AOC 69W for each group of receptors.

9.3.4.1 Calculation of Exposure Point Concentrations. RME and average EPCs were chosen for all CPCs in surface soil, sediment, and groundwater to evaluate exposures to receptors. An RME concentration represents the highest concentration of an analyte that ecological receptors would likely encounter at the site, whereas average EPCs represent typical site concentrations. For each data set, the RME concentration is equal to the maximum detected concentration because the 95th percent UCL was not calculated when there are fewer than 10 samples in the data set. The average of all samples was used to represent the average EPC unless it exceeds the maximum EPC, in which case the maximum EPC is used for both scenarios. The average of all concentrations assigns a value of one-half of the sample quantitation limit (SQL) to all samples in which the analyte is not detected.

RME and average EPCs are presented in Tables 9-14 through 9-17 for surface soil, sediment, and groundwater. A tiered approach was used to efficiently evaluate exposure and risk at AOC 69W; if no risk was calculated from exposure to the RME concentrations, then average exposure scenarios were not evaluated.

9.3.4.2 Terrestrial Wildlife. Exposure routes for wildlife receptors include direct or indirect ingestion of AOC 69W soil, sediment, and ingestion of contaminated food. To evaluate exposures at AOC 69W, representative wildlife species were selected for evaluation in food chain models which estimate contaminant exposures to wildlife species respective to their position in the food chain. Ecological exposures for the AOC 69W BERA are assumed to occur within the top two feet of soil and the top 6 inches of sediment. Contaminant exposures for terrestrial wildlife are related to the foraging characteristics of the species; therefore, terrestrial and semi-terrestrial receptors were chosen to represent the trophic levels typically found in upland suburban settings, ephemeral ditches, and forested wetland habitats. The following representative wildlife species (summarized in Table 9-18) were selected for evaluating food-chain exposures in the AOC 69W BERA:

Harding Lawson Associates

- White footed mouse (*Peromyscus leucopus*). The white-footed mouse is a small granivorous mammal (i.e., feeding primarily on seeds and young grass shoots) that inhabits wooded or scrub/shrub habitats. Invertebrates also make up a small portion of this receptor's diet. The white-footed mouse represents granivorous mammals found in terrestrial areas at AOC 69W.
- Short-tailed shrew (Blarina brevicauda). The short-tailed shrew finds suitable habitat in forests, fields, marshes, and brush. It primarily feeds on earthworms, snails, centipedes, insects, small vertebrates, and slugs (DeGraaf and Rudis, 1986). Relative to other small mammals, insectivorous species may receive high doses of contamination through bioaccumulation in prey. The shrew represents small mammal omnivores that may be found in the ditch/Willow Brook and wetland area at AOC 69W.
- American robin (*Turdus migratorius*). The robin is often seen perched in open woodlands and foraging in developed areas such as maintained grassy lawns. The robin represents avian receptors that consume earthworms, insects, and plants, and was selected to represent avian omnivores in terrestrial areas at AOC 69W.
- Red-winged blackbird (Agelaius phoeniceus). The red-winged blackbird is often seen in riparian habitats and fields. The red-winged blackbird represents avian receptors that consume insects, spiders, and seeds, and nests in shrubs, within sedges or grass, or in other emergent vegetation (Ehrlich et al., 1988). The blackbird represents avian omnivores that may be found in the ditch/Willow Brook or wetland areas at AOC 69W.
- Raccoon (*Procyon lotor*). The raccoon represents an opportunistic species that is commonly found in virtually every aquatic habitat and in developed areas. Raccoons are primarily active from sunset to sunrise (USEPA, 1993a). Raccoons consume a variety of food items including fleshy fruits, nuts, acorns, grains, insects, frogs, crayfish, and eggs (USEPA, 1993). The raccoon represents higher trophic level omnivorous mammals found in the ditch/Willow Brook and the wetland at AOC 69W.

Exposure assumptions (e.g., body weights, food ingestion rates, relative consumption of food items) for each of the representative wildlife species for AOC 69W are

Harding Lawson Associates

W003971.080

provided in Appendix P, Table P-1.

The Site Foraging Frequency (SFF) approximates the frequency a receptor feeds within a potentially contaminated site area and is calculated by the ratio of the site area to the receptor's home range. By definition, the SFF cannot exceed 1. For instance, the potentially contaminated surface soil area (calculated to be approximately 1 acre) is larger than the home ranges of the white-footed mouse, short-tailed shrew, American robin, and the red-winged blackbird. Therefore, it is assumed that the SFF for these receptors in these areas is 1.

To estimate receptor exposures to contaminants in site media and contaminated food items, a Potential Dietary Exposure (PDE) (or body dose) was estimated for all representative wildlife species for each CPC in all media according to the equations in Table 9-19. The PDEs calculated from exposure to AOC 69W surface soil and sediment CPCs for each receptor are presented in Appendix P, Tables P-8 through P-10.

Tissue concentrations of CPCs in prey items were estimated using bioaccumulation factors (BAFs) for surface soil and sediment. BAFs for most receptors were extrapolated from literature values or estimated using regression equations from scientific literature. Based on the lack of scientific data for VOC bioaccumulation and evidence provided in several reference materials (Suter, 1993; Maughan, 1993), an assumption was made that VOCs do not bioaccumulate in biological tissue. The general approach used to select BAFs for AOC 69W is summarized in Table 9-20. BAFs reported in the scientific literature for avian and mammalian receptors are defined as the reported ratios of CPC concentrations in the tissues of these receptors (mg contaminant/kg tissue wet weight) to the concentrations of CPCs in their food items (mg contaminant/kg tissue wet weight).

BAFs for invertebrate and plant prey items are defined as the ratio of the CPC concentration in plant or invertebrate tissue (mg contaminant/kg tissue wet weight) to the CPC concentration in solid media (mg contaminant/kg dry weight). BAFs for each of the CPCs evaluated at AOC 69W are included in Appendix P, Table P-2.

Although indirect exposures to wildlife from groundwater CPCs exist, this exposure pathway is not a significant route of exposure and is unlikely to result in risk. Groundwater discharges to the surface (i.e., the ditch) for a only a short period

Harding Lawson Associates

during the year, limiting the amount of time an ecological receptor may be exposed to groundwater. The analytes detected in groundwater would likely attenuate to sediments prior to discharge, decreasing the concentration of a contaminant to which an ecological receptor may be exposed. In addition, groundwater would be diluted at the point of discharge, as it is likely that groundwater would be discharging to the ditch when the local water table was high and surface water would most likely be present. Consequently, risks to wildlife from exposure to groundwater CPCs were not quantitatively evaluated at AOC 69W.

9.3.4.3 Terrestrial Plants and Invertebrates. Terrestrial plants and soil invertebrates may be exposed to CPCs via direct contact with, root uptake (plants), or ingestion (invertebrates) of CPCs measured in AOC 69W surface soil. For the purposes of the AOC 69W BERA, exposures to terrestrial plants and invertebrates are assumed to occur within the top 2 feet of surface soil.

9.3.4.4 Aquatic Receptors. Aquatic organisms may be exposed to CPCs via direct contact with sediment and groundwater as it discharges to the surface. Surface water EPCs were estimated based on the detected analyte concentrations in groundwater; however, this exposure assumption may be overly conservative because analytes may attenuate before reaching a discharge area, or may become diluted at the discharge area.

Subchronic toxicity tests were performed using the midge (*Chironomus tentans*) and the amphipod (*Hyalella azteca*). These benthic and epibenthic (respectively) invertebrates were exposed to sediment samples collected from three downgradient locations at AOC 69W (ZWD-95-02X, ZWD-95-03X, and ZWD-95-06X), shown in Figure 5-3. Sediment samples for toxicity testing were collected concurrently at AOC 69W and AOC 57, and it was determined that sample location 57D-95-08X (collected from Cold Spring Brook, a perennial stream at AOC 57) would be used as a reference sample for AOC 69W and AOC 57. However, sediment at AOC 69W contains trace organic matter (< 1 percent) and is described as sandy and gravelly with some cobbles, whereas sediment from the AOC 57 reference sample has sandy texture and finer grain sizes with about 26 percent organic matter. Consequently, the AOC 57 sample location was not an ideal reference for AOC 69W, however it was the only reference location available at the time of the testing. Amphipod survival may be attributable to characteristics of the substrate and percent organic matter rather than site contamination. The uncertainties associated with using 57D-95-08X

Harding Lawson Associates

as a reference station are also discussed in Subsection 9.2.7.

Sediment toxicity studies were conducted in accordance with the guidelines Methods for Measuring the Toxicity and Bioaccumulation of Sediment-Associated Contaminants with Freshwater Invertebrates (USEPA, 1994). Eight replicates of laboratory-raised midges and vendor-supplied amphipods were maintained for each sediment sample and the control. Replicates were exposed to whole sediment in 10-day static renewal toxicity tests. After exposure, surviving organisms were counted and weighed. The results of these tests were used to evaluate potential toxicity from exposure to downgradient sediment contamination. A summary of results are presented in Subsection 9.3.5.3. Detailed results of the AOC 69W sediment toxicity testing are presented in Appendix P.

9.3.5 Ecological Effects Assessment

As stated in the problem formulation, the assessment endpoints of the BERA are the survival and propagation of ecological receptor populations at AOC 69W. The ecological effects assessment discusses what measurement endpoints were used to represent the assessment endpoints evaluated in this BERA. Terrestrial and semi-terrestrial wildlife receptors, terrestrial plants, terrestrial invertebrates, and aquatic organisms are potentially exposed to CPCs detected in AOC 69W site media; the measures of adverse ecological effects for these receptor groups are discussed separately.

9.3.5.1 Terrestrial and Semi-terrestrial Wildlife. Because no long-term wildlife population data are available at Devens, a direct measurement of the survival and propagation of wildlife populations at AOC 69W is not possible. The literature-derived results of laboratory toxicity studies that relate the dose of a contaminant in an oral exposure with an adverse response to growth, reproduction, or survival of a test population (avian or mammalian species) were used in food-web models as a measure of the assessment endpoint. Lethal and sublethal wildlife ingestion toxicity data (which are used to derive reference toxicity values (RTVs) for evaluating risk) are presented in Appendix P, Table P-3. Wildlife effects from exposure to CPCs in surface soil and downgradient sediment were evaluated in the AOC 69W BERA.

For each CPC identified and each representative wildlife species selected, two RTVs were identified. A lethal RTV represents the threshold for lethal effects and is based

Harding Lawson Associates

on oral LD_{50} data (oral dose [in mg/kg body weight-day] lethal to 50 percent of a test population). The lethal RTV is equal to one-fifth of the lowest reported LD_{50} for the most closely related test species; this is considered to be protective against lethal effects for 99.9 percent of individuals in a test population (USEPA, 1986). When LD_{50} data were not available, a LOAEL for lethal effects was selected. A sublethal RTV is selected to represent a threshold body weight-normalized dose for adverse effects related to reproduction or growth. A summary of RTVs selected from the ingestion toxicity data are provided in Appendix P, Table P-4.

In some instances, one-fifth of the LD⁵⁰ (the lethal RTV) is less than the sublethal LOAEL or NOAEL. Therefore, the RTV used for evaluating adverse effects to wildlife is conservatively selected as the lessor of the lethal or sublethal RTVs derived from the literature. If neither lethal nor sublethal toxicity information were available for a taxonomic group, RTVs from another taxonomic group were used as surrogates. The uncertainties associated with using inter-taxonomic surrogates are discussed in Subsection 9.3.7.

9.3.5.2 Terrestrial Plants and Invertebrates. Site-specific toxicity data for plants and invertebrates are not available for AOC 69W; therefore, the results of toxicity studies from the literature that relate the soil concentrations of a contaminant to adverse growth, reproduction, or survival effects of a test population were used as a measure of the assessment endpoint. These study results are summarized in Appendix P, Tables P-5 (plants) and P-6 (invertebrates). Terrestrial plant and invertebrate effects from exposure to upland and wetland surface soils are evaluated in the AOC 69W BERA.

For plants, the effects primarily considered were measures of growth or yield as these response parameters are most common in phytotoxicity studies. For invertebrates, the effects primarily considered were measures of reproduction or mortality; when LC_{50} data were used, one-fifth of the LC_{50} was used to be protective of 99.9 percent of the population (USEPA, 1986).

9.3.5.3 Aquatic Receptors. Aquatic organism effects from exposure to sediment are evaluated in the AOC 69W BERA. Potential adverse ecological effects associated with CPCs in sediment were evaluated by comparing the CPC concentrations with literature-derived benchmarks. In addition, adverse effects to aquatic receptors from direct contact with CPCs in sediment were evaluated based on the results of the

Harding Lawson Associates

sediment toxicity study using the midge and amphipod. The results of the sediment toxicity test for the amphipod and midge are presented in Table 9-21 and Appendix P, and are summarized below.

Literature values that relate the concentration of a contaminant with an effect level (derived from data for adverse growth, reproduction, or survival effects of test populations) are used as a measure of the assessment endpoint. Sediment benchmarks selected for comparison to detected sediment concentrations included National Oceanic and Atmospheric Administration (NOAA) Effects Range-Low (ER-L) and Effects Range-Median (ER-M) sediment guidelines (Long et al., 1994) based on the National Status and Trends Program approach; USEPA Sediment Quality Guidelines (SQG) based on equilibrium partitioning (USEPA, 1988a; USEPA, 1993b); Ontario Ministry of the Environment (OME) Lowest Effect Level (LEL) provincial sediment quality guidelines (Persaud et al., 1996) based on the Apparent Effects Threshold (AET) approach; and New York State Department of Environmental Conservation (NYSDEC) sediment quality criteria (NYSDEC, 1994).

Midge survival and growth results for AOC 69W sediments were not significantly different from either the control or the reference samples. Growth results for amphipods exposed to AOC 69W sediment samples were also not significantly different from the control or reference samples. Amphipod survival in sample ZWD-95-06X (36 ± 23 percent) was significantly less than the control (64 ± 18 percent). Amphipod survival was also significantly less in samples ZWD-95-02X (55 ± 24 percent) and ZWD-95-06X compared to the reference sample (80 ± 21 percent).

Results appear to indicate that midge survival and growth, as well as amphipod growth, are not adversely affected from exposure to downgradient sediment at AOC 69W. Amphipod survival may be adversely affected from exposure to sediment at sample locations ZWD-95-02X and ZWD-95-06X. However, there is uncertainty associated with the amphipod control results because the control sample did not meet the 80 percent acceptance criteria for survival, as defined in the *Methods for Measuring the Toxicity and Bioaccumulation of Sediment-Associated Contaminants with Freshwater Invertebrates* (USEPA, 1994). This deviation alters the reliability of the data analysis; however, inferences about this toxicity test can still be discussed. The control sample for the midge did meet its 70 percent acceptance criteria.

The effects of groundwater CPCs on aquatic organisms were evaluated by comparing

Harding Lawson Associates

groundwater EPCs to surface water RTVs. Surface water RTVs selected for comparison to groundwater exposure concentrations include Federal chronic Ambient Water Quality Criteria (AWQC) (USEPA, 1991b; USEPA, 1988b) and aquatic toxicity information from the USEPA AQUIRE database (AQUIRE, 1996). Effects concentration data obtained from AQUIRE are included in Appendix P, Table P-7. Chronic AWQC are concentrations that, if not exceeded by the four-day average chemical concentration more than once every three years, are considered protective of most species of aquatic life and its uses (USEPA, 1983). When criteria were lacking, lowest observed effects concentrations measuring survival, growth, reproduction, and biodiversity endpoints were derived from the AQUIRE database (AQUIRE, 1996).

9.3.6 Risk Characterization

This subsection discusses how risks were characterized for ecological receptors exposed to contaminated media at AOC 69W. A comparison of exposure information with the appropriate concentration-response toxicity data is the basis for risk characterization.

9.3.6.1 Terrestrial and Semi-terrestrial Wildlife. Risks for the representative wildlife species associated with ingestion and bioaccumulation of CPCs in surface soil, sediment, and prey items are quantitatively evaluated using HQs, which are calculated for each CPC by dividing the PDE based on RME concentrations by the selected lethal or sublethal RTV. HIs are determined for each receptor by summing the HQs for all CPCs. When the estimated PDE is less than the RTV (i.e., the HQ < 1), it is assumed that chemical exposures are not associated with adverse effects on survival, growth, or reproduction, and that no risks to wildlife populations exist. When an HI is greater than 1, a discussion of the ecological significance of the HQs comprising the HI is provided, and risks from exposure to average concentrations of CPCs are evaluated.

This hazard ranking scheme evaluates potential ecological effects to individual organisms and does not evaluate potential population-wide effects. Contaminants may cause population reductions by affecting birth and mortality rates, immigration, and emigration (USEPA, 1989a). In many circumstances, lethal or sublethal effects may occur to individual organisms with little population- or community-level impacts; however, as the number of individual organisms experiencing toxic effects increases,

Harding Lawson Associates

W003971,080

the probability that population effects will occur also increases. The number of affected individuals in a population presumably increases with increasing HQ or HI values; therefore, the likelihood of population-level effects occurring is generally expected to increase with higher HQ or HI values.

The HQs and HIs calculated based on RME and average EPCs for each representative wildlife species are provided in Appendix P, Tables P-8 through P-10. A summary of risks to representative wildlife receptors is provided in Table 9-22 and in the following paragraphs. There are no toxicity data available relating wildlife exposures to TPHC with adverse responses; therefore, TPHC exposures were not included in the food-web model, and potential adverse effects from TPHC exposure remain an uncertainty.

Surface Soil. The HQs and HIs calculated for each representative wildlife species are provided in Appendix P, Tables P-8 and P-9; a summary of risks is provided in The summary HI for the short-tailed shrew exposed to RME Table 9-22. concentrations in soil is 2.8. The primary risk contributor to the shrew is lead, which was detected at a maximum concentration (238 $\mu g/g$) that exceeds the background concentration for lead (48 μ g/g) by nearly five times. The summary HI for the shrew based on average EPCs in soil (which are more representative of site conditions) is 0.88, suggesting that small mammals are generally not at risk from exposure to lead in surface soil. Summary HIs for all other wildlife receptors exposed to RME and average EPCs in soil are less than 1. These results suggest that adverse effects to the small mammal population are negligible. Several site-related factors suggest that the estimated risk to RME lead concentrations may be overstated. These factors include the unlikelihood that receptors would consistently be exposed to these RME concentrations and the (generally) greater availability of inorganics in laboratory dietary studies. Furthermore, the low magnitude of estimated risk from RME concentrations of lead indicates a low probability that small mammals would experience adverse effects. Risks are not expected for small mammals based an HI of 2.8

<u>Sediment</u>. HQs and HIs calculated for each representative wildlife are provided in Appendix P, Table P-10, and a summary of risks is provided in Table 9-22. Summary HIs for all wildlife receptors exposed to RME EPCs in sediment are less than 1. Therefore, it is unlikely that adverse effects to wildlife receptors would occur from exposure to RME concentrations of analytes in AOC 69W sediment.

Harding Lawson Associates

W003971.080

9.3.6.2 Terrestrial Plants. Risks for terrestrial plants were evaluated by comparing the selected phytotoxicity benchmarks (Appendix P, Table P-5) to RME and average EPCs of analytes detected in soil. The results of the surface soil evaluation for AOC 69W are presented in Table 9-23 and are discussed below. There are no data available relating plant exposures to acetone, trichlorofluoromethane, or TPHC with adverse responses; therefore, risks to plants from exposure to these chemical could However, the concentrations of acetone not be evaluated. and trichlorofluoromethane were extremely low. Furthermore, acetone is a common laboratory contaminant, and the chlorofluorocarbons are highly non-toxic to animal receptors. Consequently, risks attributable to these analytes are considered negligible. Risks resulting from TPHC exposure could not be estimated due to the limited availability of toxicity data relating TPHC concentrations in soil and phytotoxicity.

RME EPCs for lead and zinc in AOC 69W surface soil exceed phytotoxicity benchmarks (Table 9-23). All other CPC exposure concentrations are less than phytotoxicity benchmarks, suggesting that plants are not at risk from exposure to VOCs, SVOCs, pesticides, beryllium, cobalt, copper, mercury, nickel, or selenium in AOC 69W surface soil. The RME and average concentration of lead (238 and 61.2 $\mu g/g$, respectively) exceed the background concentration and the phytotoxicity benchmark values (48 and 50 $\mu g/g$, respectively) by less than a factor of five and two, respectively. The RME concentration of zinc (71.7 $\mu g/g$) slightly exceeds the background (44 $\mu g/g$) and benchmark value (50 $\mu g/g$) by less than a factor of two. The average concentration of zinc (32.5 $\mu g/g$) is less than background and the benchmark value. Based on the slight exceedances of the RME concentration, it is likely that risks to plants from exposure to zinc would be negligible.

Maximum lead and zinc concentrations occurred in a single sample location (ZWS-95-42X). It is possible that plants in the vicinity of sample location ZWS-95-42X may exhibit phytotoxic effects from exposures to lead. However, concentrations of lead and zinc in the other six samples were below background concentrations and benchmarks. Given this isolated area of contamination, it is unlikely that populationlevel effects to plants from these analytes would occur. Risks to plants based on average concentrations of lead and zinc, which are more representative of actual exposures to plant populations at AOC 57 are negligible. In addition, no signs of phytotoxicity were noted during the site survey.

Harding Lawson Associates

9.3.6.3 Terrestrial Invertebrates. Risks for terrestrial invertebrates were evaluated by comparing the selected invertebrate benchmarks (Appendix P, Table P-6) to RME and average EPCs. The results of the surface soil evaluations for AOC 69W are presented in Table 9-23 and are discussed below. There are no toxicity data available relating invertebrate exposures to acetone, trichlorofluoromethane, beryllium, cobalt, selenium, or TPHC with adverse responses; therefore, invertebrate exposures to these analytes were not evaluated and potential adverse effects remain an uncertainty. Because of low concentrations, low toxicity, and/or presence as a laboratory contaminant, as discussed above, neither trichlorofluoromethane nor acetone are significant risk issues. RME and average EPCs of all other analytes detected in AOC 69W surface soil are less than available invertebrate benchmarks. The available data suggest that, among organics, TPHC may present a significant risk to certain invertebrate communities. Risks related to inorganics exposure are negligible, based on the comparison of EPCs with RTVs (Table 9-23).

9.3.6.4 Aquatic Organisms. Risks to aquatic receptors from exposure to sediment at AOC 69W were characterized based on the toxicity test evaluation performed for the midge and amphipod and on a comparison of sediment EPCs with the toxicity benchmarks discussed in subsection 9.2.5.3. In addition, a qualitative evaluation of upgradient and downgradient sediments is performed.

<u>Sediment - Toxicity Test Results</u>. Risks for aquatic macroinvertebrates were characterized based on the results of sediment toxicity tests. The sediment analytical and toxicity test samples were collected concurrently; therefore, the analytical results for the sediment samples were used to interpret the contaminant exposures and responses of the test species (midges and amphipods) in the toxicity tests.

As previously discussed in Subsection 9.2.5.3, amphipod survival was significantly lower in sediment sample ZWD-95-06X than in the control sample and the reference sample (57D-95-08X), and amphipod survival was significantly lower in sediment sample ZWD-95-02X than in the reference sample. No other statistically significant differences in midge or amphipod survival and growth were observed between the reference or control samples and the sediment collected at AOC 69W.

Since the survival of amphipods exposed to the control sediment was below the 80 percent acceptance criteria, comparisons to the control may not be appropriate. Furthermore, 57D-95-08X may not be a suitable reference station for the ditch

Harding Lawson Associates

sediments at AOC 69W. However, sediment samples for toxicity testing were collected concurrently for AOC 57 and AOC 69W, and it was assumed that the reference location identified at AOC 57 would be suitable for a reference at AOC 69W. The unacceptable control survival, and the physical differences between the reference sediment and site sediment, limit the inferences that can made regarding the data.

<u>Sediment - Benchmark Comparison</u>. The comparison of sediment concentrations with toxicity benchmarks values is provided in Table 9-24. Sediment concentrations of two metals, three pesticides, and four semivolatiles exceeded sediment benchmark values.

RME concentrations of copper (23.4 μ g/g) and nickel (18.1 μ g/g) slightly exceed OME LEL and NYSDEC LEL values of 16 μ g/g. These sediment benchmarks are representative of concentrations that can be tolerated by the majority of benthic organisms (Persaud, et al, 1996). Average concentrations of these analytes, which are more representative of site conditions, are below benchmark values. This evidence suggests that copper and nickel may not cause adverse effects to the majority of aquatic organisms. RME and average concentrations of pesticides (DDD, DDE, and DDT) exceeded toxicity benchmarks; however, these analytes are not likely related to the fuel oil spill at AOC 69W. As with metals, RME and average concentrations of semivolatiles slightly exceed the NOAA ER-Ls and the OME LELs; these slight exceedances suggest that in sediment, PAHs may not cause adverse population level effects for aquatic organisms.

Relating the results of the benchmark comparison for chemicals detected in downgradient sediment to the adverse responses observed in the toxicity tests may be difficult. Of the analytes that only slightly exceeded sediment benchmarks, the four PAHs and copper were detected at their highest concentrations at ZWD-95-06X (where *H. azteca* survival was significantly lower than in the control and reference samples), and nickel was detected at its highest concentration at ZWD-95-02X (where *H. azteca* survival was significantly lower than in the reference sample). These results suggest adverse effects to populations of these test organisms. However, *H. azteca* are a highly sensitive macroinvertebrate species, and effects on the more diverse and tolerant community structure occurring in the natural environment are unclear. In addition, sandy nature (i.e., low organic carbon content) of the ditch substrate may have contributed to the reduced survival observed in the amphipod

Harding Lawson Associates

toxicity tests as compared to the reference sample collected from Cold Spring Brook. In addition, the highest concentration of any DDT_{R} compound was detected at ZWD-95-03X, where no adverse effects were observed in test organisms.

<u>Comparison to Upgradient Sediment</u>. Eight pesticides, five semivolatiles, two volatiles, five metals, and TPHC were selected as CPCs for upgradient sediment. In general, concentrations of contaminants (that were found in both upgradient and downgradient samples) were higher in -upgradient samples. Alpha- and gamma-chlordane, Aroclor-1260, dieldrin, Endosulfan II, and chrysene were detected in upgradient samples, but not in downgradient samples. Lead and zinc were selected as CPCs for upgradient sediment, but not for downgradient sediment. Contamination in upgradient sediments does not appear to be site-related. Upgradient sediment samples were collected in high automotive traffic areas, which may account for higher levels of contamination (particularly TPHC, PAHs, and possibly lead). Pesticides are most likely not associated with site activities. It is likely that concentrations detected in upgradient sediment would influence concentrations observed in downgradient sediment.

<u>Groundwater - Benchmark Comparison</u>. Risks to aquatic organisms from exposures to AOC 69W groundwater are evaluated by comparing exposure concentrations to surface water toxicity benchmarks. This comparison is presented in Table 9-25. Maximum and average concentrations of iron and manganese (filtered and unfiltered) exceeded aquatic benchmarks. However, the benchmarks for these analytes are protective of species (i.e., fish) that would not likely inhabit the ditch at AOC69W. These maximum values were associated with one sample location (69W-94-10) located near the school (i.e., furthest from the discharge point). In addition a soil removal action at the site has mitigated the reducing conditions that may have mobilized these metals in groundwater.

Two pesticides (gamma-chlordane and heptachlor epoxide) were detected in only one of eight samples (at sample location 69W-94-10). Concentrations of both of these analytes exceed toxicity benchmarks. However, these analytes are not associated with past fuel oil spill activities. Bis(2-ethylhexyl)phthalate and phenanthrene were detected in two of eight samples at maximum concentrations well above benchmark values. These maximum concentrations were also associated with 69W-94-10. Given the low frequency of detection and that contamination is associated with a single sampling location, exposure to these analytes at AOC 69W may not result in adverse

Harding Lawson Associates

population-level effects.

9.3.7 Uncertainty Analysis

The objective of the uncertainty analysis is to discuss the assumptions of the BERA process that may influence the risk assessment results and conclusions. General uncertainties inherent in the risk assessment process and in the AOC 69W BERA are included in Table 9-26.

Specific uncertainties associated with the assessment of risks associated with contamination at AOC 69W include the following.

- Risks to terrestrial receptors associated with exposure to TPHC in surface soil, sediment, and groundwater may be underestimated. Although selected as a CPC for these media, TPHC was not evaluated in the BERA because of the lack of toxicological benchmarks.
- Risks to avian species may be over- or underestimated because bioaccumulation and toxicity data for this taxonomic group are generally lacking in the literature. To conservatively estimate risks to avian species at AOC 69W, mammalian data were used as surrogate values when avian data were lacking. Significant metabolic, life stage, and physiological differences between these two taxa provide considerable extrapolation uncertainty. Furthermore, there is considerable uncertainty extrapolating between laboratory text organisms and wildlife.
- Risks to terrestrial wildlife receptors from food chain exposures to lead may have been underestimated because a BAF of zero was used for uptake into plants. This FAB of zero was obtained from the literature; however, the literature varies on the subject. The risk conclusions are not likely to be changed substantially by using other lead FAVs that will only slightly raise the estimated risks from lead exposure.
- Risks to plants and invertebrates may be under- or overestimated because phytotoxicity and invertebrate benchmarks for most analytes are either lacking or are based on a limited number of studies. Specifically, potential risks to plants from exposure to acetone, trichlorofluoromethane, and TPHC, and

Harding Lawson Associates

potential risks to invertebrates from exposure to beryllium, cobalt, selenium, acetone, trichlorofluoromethane, and TPHC, could not be evaluated.

There is uncertainty associated with using unfiltered groundwater data for evaluating risk to aquatic organisms from exposure to inorganic analytes. Risks to aquatic organisms may be overestimated because unfiltered data represent the total fraction of analytes that occur in groundwater, including those that are sorbed to particulates. The bioavailable (i.e., filtered) fraction provides a better measure of toxicity; however, particulate matter can also be hazardous because of mechanical obstruction with physiological processes.

- No sediment benchmarks exist for cobalt, trichlorofluoromethane, diesel fuel, or TPHC gas fraction; therefore, risks to aquatic organisms from exposure to these analytes remain an uncertainty.
- Risks to aquatic organisms from exposure to several metals (e.g., aluminum, arsenic, barium, cobalt, iron, magnesium, manganese, potassium, and vanadium) in sediment may have been underestimated because these analytes were screened out of the ERA using background values for surface soil. However, none of these metals have sediment toxicity benchmarks (except for arsenic); therefore, it is possible that these analytes do not contribute significantly to adverse effects for aquatic organisms. The maximum detected concentration of arsenic (10 μ g/g) only slightly exceeds the most conservative sediment guideline (OME LEL=6 μ g/g); therefore, potential risks from exposure to arsenic in sediment are unlikely.
- Risks to amphibians may be underestimated, as this exposure pathway was not evaluated due to the paucity of toxicological data relating dermal exposure and toxicological effects to amphibians.
- There is uncertainty associated with the control results in the amphipod toxicity test. Only 64 percent survival was observed in the control samples, as opposed to the recommended minimum of 80 percent. These low survival rates may be attributable to shipment stress or contaminated control sediments. The poor control results may have impacted the results of the evaluation. There are also uncertainties associated with the sample to sample variation as indicated by large standard deviations. In addition, there is

Harding Lawson Associates

uncertainty associated with using a sample from AOC 57 as a reference sample for AOC 69W, as the habitats for these areas are different. Sediment at AOC 69W contains trace organics (< 1 percent) and is described as sandy and gravely with some cobbles, whereas sediment from the AOC 57 reference sample has sandy texture and finer grain sizes, with about 26 percent organic matter.

Risks for analytes detected in the method blanks, trip blanks, or rinsate blanks, and in the field samples, may have been overestimated. In particular, phthalates, acetone, and chlorinated solvents were detected in method blanks associated with water samples, and phthalates, TPHC, and various VOCs were detected in method blanks associated with solid media. In addition, several chlorinated solvents, bis(2-ethylhexyl)phthalate, mercury, lead, iron, potassium, and manganese were detected in rinsate blanks, and several chlorinated solvents were detected in trip blanks. Although analytes were detected in blanks, none of the blanks were grossly contaminated. In addition, all of the analytes detected were common laboratory contaminants, most were not analytes of concern, and all were detected at very low levels.

9.3.8 Summary of BERA for AOC 69W

Potential risks for ecological receptors were evaluated for CPCs in surface soil, sediment, and groundwater at AOC 69W. The following items summarize the results of the AOC 69W BERA:

- Small birds, small mammals, and predatory mammals exposed to maximum concentrations of analytes in surface soil would not likely be at risk. HIs for these taxa were all less than one (except for the shrew which had a low HI of 2.8). Uncertainty is greater for avian populations because of a generally inadequate toxicological reference base.
- It is unlikely that wildlife receptors exposed to maximum concentrations of analytes in sediment would be at risk. HIs for all receptors were less than one.
- Terrestrial plants may be at risk from exposure to lead in surface soil; however, elevated lead concentrations were localized. In addition, no stressed

Harding Lawson Associates

vegetation was observed during site walkovers conducted by HLA ecologists.

• It is unlikely that terrestrial invertebrates exposed to maximum detected concentrations of analytes in soil would be at risk. All of the analytes detected in soil were below invertebrate benchmark values or were detected at very low concentrations.

- Given the results of the toxicity tests, benthic macroinvertebrates (e.g., amphipods) may be at risk from exposure to sediment at AOC 69W. However, there are many uncertainties associated with the reliability of the statistical evaluation for amphipods, and the suitability of the ditch to provide adequate aquatic habitat.
- Based on the benchmark comparisons for aquatic receptors, risks from exposure to copper, nickel, and PAHs in sediment and in groundwater that discharges to the surface at AOC 69W are negligible. The monitoring well where maximum concentrations of all groundwater contaminants were detected is located close to the school. Concentrations would likely be lower due to attenuation before discharging to the surface. In addition, the seasonal groundwater discharge to the ditch or the wetland (at least 300 feet away) may be subject to dilution from precipitation, overland flow, and mixing with surface water.

Harding Lawson Associates

10.0 CONCLUSIONS AND RECOMMENDATIONS

RI activities have been conducted by HLA at AOC 69W to evaluate the nature and distribution of the groundwater and soil contamination detected during previous investigations. Based on a review of RI data, a soil removal action was implemented to eliminate the source of contamination at AOC 69W. Conclusions developed from the RI findings are presented in the following subsections.

10.1 CONCLUSIONS

The following conclusions are based on interpretation of data collected from previous investigations, the RI, and the removal action completed at AOC 69W. Tables 10-1 and 10-2 summarize the results of the Human Health and Ecological Risk Assessments.

- The geologic setting at AOC 69W is comprised of dark yellowishbrown, fine to coarse sands, gravely sands, and silty sands which underlie and are interbedded with organic wetland deposits. The near surface organic deposits are more predominant near Willow Brook and within the grassy area north of the paved parking lot. Soils surrounding and underlying Willow Brook consist of dark grayishbrown sandy silt 4 to 5 feet in thickness.
- For the purpose of this RI the hydrogeologic condition at AOC 69W is dominated by the overburden aquifer. The water table is found in the overburden sands, gravely sands, and silty sands at depths ranging from 7 to 1 foot bgs. Local groundwater flow is primarily to the north-northwest. Groundwater discharges to Willow Brook in the vicinity of AOC 69W at times of high water levels.

Estimates of hydraulic conductivity range between 1.3 x 10⁻³ cm/sec (3.7 ft/day) and 3.0 x 10⁻² cm-sec (85 ft/day) with a geometric mean 1.6 x 10⁻² cm/sec (45 ft/day). A groundwater flow velocity of 0.7 ft/day was calculated using the geometric mean of estimated hydraulic conductivities and horizontal hydraulic gradients.

Harding Lawson Associates

SECTION 10

- The primary contaminant source was identified as contaminated subsurface soil in the vicinity of the new boiler room. The contaminated subsurface soils were attributed to a 1972 fuel oil release from ruptured piping. A 1978 fuel oil release due to poorly joined piping in the vicinity of the old boiler room does not appear to have resulted in a significant contaminant source.
- The largest area of soil contamination extended from the new boiler room to a 250-gallon UST in the wooded area approximately 300 feet northwest of the school. The oil recovery system installed in response to the 1972 fuel oil release consisted of an underground pipe leading from the school to the 250-gallon UST and appeared to have acted as a conduit for contaminant migration. Detected contaminants were primarily TPHC, PAHs, and inorganics at approximately 6 to 10 feet bgs adjacent to the school building and 0 to 4 feet bgs downgradient in the grassy area and vicinity of the 250 gallon UST.
- Based on a review of the RI chemical data, a soil remedial action was completed in the winter of 1997 and 1998 to remove contaminated subsurface soil associated with the 1972 fuel oil spill and downgradient piping leading to the 250-gallon UST, thus, removing the majority of the source of soil and groundwater contamination.
- Confirmatory subsurface sample results indicate that EPH/VPH concentrations immediately adjacent to the school still exceed MCP S-1/GW-1 soil standards, but that subsurface soil concentrations in other portions of the area of the soil remedial action are generally low.
- A second area of soil contamination was identified adjacent to the school building outside of the old boiler room. Analytical data indicates that the contaminants are primarily TPHC at depths of 4 to 7 feet bgs. The observed contamination does not appear to be migrating downgradient based upon the results of numerous TerraProbeSM and monitoring well analytical samples. Further contaminant migration is unlikely based upon the age of the release and the paved parking lot inhibiting recharge.
- Fuel related VOCs, SVOCs, and TPHC comprise the observed groundwater contaminants at AOC 69W. Varying degrees of

Harding Lawson Associates

W003971.080

groundwater contamination, as identified by field and off-site analysis, were observed to extend from the new boiler room towards the 250gallon UST located approximately 300 feet to the northwest. The area of groundwater contamination was coincident with the underground piping for the oil recovery system. Contaminant concentrations were highest between the new boiler room and monitoring well 69W--94-13, which was also the location of the most contaminated soils. The soils around 69W--94-10 and 69W--94-13 were removed as part of the soil removal action.

- The soil removal action has removed the majority of the source of fuel oil contamination to groundwater, and has likely mitigated the reducing conditions that have caused metals to be mobilized in groundwater. Chemical concentrations in groundwater are expected to improve.
- Soil and groundwater contaminant distributions were vertically located in the vicinity of the water table and above.
- There is not any significant groundwater contamination associated with the 1978 fuel oil release. A soil removal conducted following the release likely mitigated the potential for groundwater contamination.
- TPHC data from groundwater samples collected prior to the soil removal action are highly variant from round to round and even between samples and duplicates.
- Contaminant distribution in sediment samples suggests that the petroleum contamination may have contributed SVOCs to deep (2 feet bgs) Willow Brook sediments. Sediment VOCs may have also come from asphalt-lining in Willow Brook. Pesticides, PCBs, and TPHCs were detected primarily in upgradient samples suggesting that their source is not site related.

Human Health

• Chemicals of potential concern for the human health risk assessment were identified in surface soil and subsurface soil representative of post-removal action conditions, sediment, groundwater and indoor air. They included

Harding Lawson Associates

W003971.080

SECTION 10

metals, semi-volatile organic compounds, volatile organic compounds, and petroleum-related compounds including total petroleum hydrocarbons, extractable petroleum hydrocarbons, volatile petroleum hydrocarbons, and polynuclear aromatic hydrocarbons. Among these CPCs, only the petroleumrelated compounds are directly associated with the release of fuel oil at AOC 69W.

- The presence of the constituents detected in the building air is likely due to the presence of numerous ambient sources within and outside the building, and not due to fuel-related constituents that may be present in soil and groundwater beneath the building. Nevertheless, circumstantial evidence indicated that a possible association between low level detections of xylenes, ethylbenzene, and methylheptane detected in air samples and soil samples collected in the vicinity of the northwestern portion of the building could not be ruled out. However, even the maximum detected concentrations of these three constituents were below levels that would include them as CPCs in the risk assessment. This indicates that pupils and school staff members who may hypothetically be exposed to these constituents in building air would not incur health risks even close to those considered to be of concern by the USEPA.
- Estimated cancer and non-cancer risks associated with possible current and future land use exposures to surface soil, subsurface soil, sediment, groundwater discharge to surface water and indoor air are within acceptable levels established the USEPA.
- Estimated cancer and non-cancer risks associated with hypothetical exposures to AOC 69W groundwater used as a residential drinking water source exceed levels generally considered acceptable by USEPA. However, these risks represent a *hypothetical worst-case* evaluation of potential exposures and risks because groundwater at and beneath AOC 69W is not used as a source of drinking or industrial water. Moreover, the risks are primarily due to the presence of arsenic in groundwater, which is not interpreted to be directly related to the release of fuel-oil at AOC 69W.
- Based on the conclusions of the risk assessment, there are no unacceptable human health risks associated with non-groundwater media. Future use of the site as a school is not associated with any unacceptable risks. Moreover, the soil removal action at AOC 69W significantly reduced fuel oil contamination in soil thereby mitigating possible exposures to fuel-related CPCs in soil,

Harding Lawson Associates

W003971.080

reducing the CPC concentrations in groundwater, and eliminating a possible source of fuel-related vapors. Therefore, the risk estimates presented in this risk assessment worst-case estimates that are unlikely to be exceeded under conceivable future land use conditions.

Potential risks for ecological receptors were evaluated for CPCs in surface soil, sediment, and groundwater at AOC 69W. The following items summarize the results of the AOC 69W BERA:

Ecological

- There are no anticipated risks to terrestrial wildlife receptors or soil invertebrates exposed to chemicals in surface soil. Likewise, there are no risks to semi-terrestrial wildlife receptors exposed to chemicals in Willow Brook sediment.
- Risks to terrestrial plants may occur at one surface soil sample location (ZWS-95-42X) due to the presence of lead. However, lead at this location may be associated with road run-off or lawn mower maintenance due to the proximity of this sample with MacArthur Avenue. Risks to plants would be localized, and are not likely to result in population-level effects since the average concentration of lead in soil only slightly exceeded toxicity benchmarks.
- Risks to aquatic organisms were identified for certain metals; however, the soil removal action has likely mitigated the reducing conditions in the subsurface soils that may have mobilized the metals in groundwater. Adverse effects were observed for aquatic organisms exposed to sediment in toxicity tests; however, these adverse effects are likely related to the marginal aquatic habitat and substrate quality, rather than the presence of site-related chemicals. Furthermore, although significantly reduced amphipod survival was observed in toxicity tests, the reliability of these tests are questionable (because of the poor control results). The midge toxicity test results indicate a minimal likelihood of risk to aquatic receptors. Exposure point concentrations for chemicals detected in sediment only slightly exceeded sediment benchmarks, suggesting that risks are unlikely.
- Based on the conclusions of the ERA, there are no unacceptable risks associated with site-related fuel oil contamination at AOC 69W.

Harding Lawson Associates

10.2 RECOMMENDATIONS

Based upon the results and interpretation of the RI, supplemental sampling, and the soil removal action it is recommended that AOC 69W be proposed for limited action consisting of long term monitoring of downgradient groundwater quality with no additional investigation or remedial action. This course of action is supported by the following:

- The soil removal significantly reduced surficial and subsurface contaminants that were acting as a source for groundwater contamination
- No unacceptable risks posed to human health or the environment from site related fuel contamination
- Estimated risks associated with hypothetical exposure to groundwater used as a residential drinking water source exceeded levels generally considered acceptable by the USEPA. However, groundwater at AOC 69W is not used as a source of drinking or industrial water. Estimated risks for groundwater as a drinking water source are primarily due to the presence of arsenic. The presence of the arsenic is attributed to the mobilization of naturally occurring arsenic by reducing conditions in the aquifer brought on by the aerobic degradation of fuel related contaminants. The soil removal will act to lessen reducing conditions in the aquifer and therefore decrease arsenic concentrations in the groundwater.
- No chemicals of potential concern related to historical fuel releases were identified for air samples collected from inside the school
- Because the soil removal eliminated the majority of source area contaminants, estimated risks and interpretations represent worst-case estimates that are unlikely to be exceeded under future land use conditions.

Harding Lawson Associates

W003971.080

ABB-ES	ABB Environmental Services, Inc.
ADL	Arthur D. Little, Inc.
AOC	area of contamination
AREE	area requiring environmental evaluation
ARF	analysis request form
ARAR	applicable or relevant and appropriate requirements
AST	aboveground storage tank
ATEC	ATEC Environmental Consultants, Inc.
bgs	below ground surface
BNA	base neutralized acids
BRAC	Base Realignment and Closure
BRAC	Base Realignment and Closure Environmental Evaluation
BTEX	benzene, toluene, ethylbenzene, and xylene
°C	degrees Celsius
cm/sec	centimeters per second
CENAE	U.S. Army Corps of Engineers, New England District
CERCLA	Comprehensive Environmental Response, Compensation and Liability Act
CFR	Code of Federal Regulations
CFS	cubic feet per second
CLP	Contract Laboratory Program
CMR	Code of Massachusetts Regulations
COC	chain-of-custody
COR	Contracting Officer's Representative
CPC	chemical of potential concern
CRL	certified reporting limits
CSF	cancer slope factor
1,2-DCA	1,2-dichloroethane
DCE	1,1-dichloroethene
DDD	dichlorodiphenyl dichloroethane
DDE	dichlorodiphenyl dichloroethylene
DDT	dichlorodiphenyl trichloroethane
DOT	Department of Transportation
DQO	data quality objective
DRO	diesel range organics

Harding Lawson Associates

W003971.080

DWEL Drinking \	Water Equivalency Level
ECD electron ca	apture detector
	Environment, Inc.
	ntal evaluation
	ental Engineering and Geotechnics
	conductivity detector
and the second	time cancer risk
EMO Environme	ental Management Office
	point concentration
	e petroleum hydrocarbons
	Risk Assessment
	ental Science and Engineering, Inc.
	g Technologies Associates
ft/ft feet per fo	ot
ft/min feet per m	inute
ft/day feet per da	ly
ft ² /day square fee	t per day
°F degrees fal	
FFA Federal Fa	acilities Agreement
FID flame ioniz	zation detector
FS Feasibility	Study
FSP Field Samp	pling Plan
	atograph/mass spectrometer
g/mL grams per	milliliter
gpm gallons per	r minute
GPR ground-per	netrating radar
GRO gasoline ra	inge organics
H _c Henry's La	w Constant
HASP Health and	l Safety Plan
HEAST Health Eff	ects Assessment Summary Table
HI hazard ind	ex
HQ hazard que	otient
HLA Harding L	awson Associates
HSA hollow-ster	n augers

Harding Lawson Associates

W003971.080

IAG	Inter Agency Agreement
ID	inside diameter
IDW	investigation-derived waste
IR	infrared spectrophotometer
IRDMIS	Installation Restoration Data Management Information
	System
IRIS	Integrated Risk Information System
kg	kilograms
K _{oc}	organic carbon partition coefficient
m ³	cubic meters
MAAF	Moore Army Air Field
MADEP	Massachusetts Department of Environmental Protection
MCL	maximum contaminant level
MCLG	maximum contaminant level goals
MCP	Massachusetts Contingency Plan
MDL	method detection limits
MEP	Master Environmental Plan
mg/kg	milligrams per kilogram
mg/L	milligrams per liter
mL	milliliter
MMCL	Massachusetts Maximum Contaminant Level
mph	miles per hour
MS	matrix spike
MSD	matrix spike duplicate
NCEA	National Center for Environmental Assessment
NCP	National Contingency Plan
ND	non-detect
NDIR	non-dispersed infrared
NFA	no further action
NWR	National Wildlife Refuge
OSHA	Occupational Safety and Health Administration
OD	outside diameter
РАН	polynuclear aromatic hydrocarbon
PAL	project analyte list

Harding Lawson Associates

W003971.080

PARCC	precision, accuracy, representativeness, completeness, and
DC	comparability
PC	personal computer
PCB	polychlorinated biphenyl
PCE	tetrachloroethene
PID	photoionization detector
POP	Project Operations Plan
ppb	parts per billion
ppm	parts per million
PQL	practical quantitation limit
PRE	preliminary risk evaluation
PRI	Potomoc Research, Inc.
PVC	polyvinyl chloride
QA	quality assurance
QAPP	Quality Assurance Project Plan
QC	quality control
RBC	risk-based concentration
RCRA	Resource Conservation and Recovery Act
RfD	reference dose
RFTA	Reserve Forces Training Area
RI	Remedial Investigation
RME	reasonable maximum exposure
ROD	Record of Decision
RPD	relative percent difference
SA	study area
SAP	Sampling and Analysis Plan
SARA	Superfund Amendments and Reauthorization Act
SCS	Soil Conservation Service
SDWA	Safe Drinking Water Act
SI	Site Investigation
SMCL	secondary maximum contaminant level
SQL	sample quantitation limit
SVOA	semivolatile organic analysis
SVOC	semivolatile organic compound
3100	sennvolatile organic compound
TBC	to be considered

Harding Lawson Associates

W003971.080

1,1,1-TCA	1,1,1,-trichloroethane	
1,1,2,2-TCA	1,1,2,2-trichloroethane	
TCE	trichloroethene	
TCL	target compound list	
TDS	total dissolved solids	
TEX	toluene, ethylbenzene and xylenes	
TIC	tentatively identified compounds	
TPHC	total petroleum hydrocarbons	
TOC	total organic carbon	
TSCA	Toxic Substance Control Act	
TSS	total suspended solids	
μg/g	micrograms per gram	
µg/kg	micrograms per kilogram	
$\mu g/L$	micrograms per liter	
$\mu g/m^3$	micrograms per cubic meter	
µg/ml	micrograms per milliliter	
μL	microliter	
UCL	upper confidence limit	
USACE	U.S. Army Corps of Engineers	
USAEC	U.S. Army Environmental Center	
USATHAMA	U.S. Army Toxic and Hazardous Materials Agency	
USDA	U.S. Department of Agriculture	
USEPA	U.S. Environmental Protection Agency	
USFWS	U.S. Fish and Wildlife Service	
UST	underground storage tank	
VC	vinyl chloride	
VPH	volatile petroleum hydrocarbons	
VOA	volatile organic analysis	
VOC	volatile organic compound	
WPA	Works Progress Administration	
yd ³	cubic yards	

Harding Lawson Associates

W003971.080

- ABB Environmental Services (ABB-ES), Inc., 1997a. "Final RI/FS Work Plan Addendum for Supplemental Air Sampling for AOC 69W"; Prepared for U.S. Army Corps of Engineers and for Commander, U.S. Army Environmental Center; Prepared by ABB Environmental Services, Inc.; Portland, Maine; July, 1997.
- ABB-ES, 1997b. "The Analytical Approach for AOC 69W Groundwater Sampling"; Prepared for U.S. Army Corps of Engineers and for Commander, U.S. Army Environmental Center; Prepared by ABB Environmental Services, Inc.; Portland, Maine; November, 1997.
- ABB-ES, 1997c. "Final RI/FS Work Plan Addendum for Supplemental Air Sampling, Devens Elementary School"; Prepared for U.S. Army Corp of Engineers; Prepared by ABB Environmental Services, Inc., Wakefield, Massachusetts.
- ABB-ES, 1996a. "Final Task Order Work Plan Area of Contamination (AOC) 57, AOC 63AX, and AOC 69W;" Data Item A002; Prepared for Commander. U.S. Army Environmental Center; Prepared by ABB Environmental Services, Inc., Portland, Maine; January 1996.
- ABB-ES, 1996b. "Final Task Order Work Plan Addendum for Area of Contamination 57;" Task Order No. 1, Modification No. 1; Prepared for Commander. U.S. Army Environmental Center; Prepared by ABB Environmental Services, Inc., Wakefield, Massachusetts; August 1996.
- ABB-ES, 1995a. "Revised Final Site Investigation Report Groups 2, 7, and Historic Gas Stations, Fort Devens, Massachusetts"; Data Item A009; prepared for Commander, U.S. Army Environmental Center; prepared by ABB Environmental Services, Inc., Portland, Maine; October 1995.
- ABB-ES, 1995b. "Final Project Operations Plan for Site Investigations and Remedial Investigations, Fort Devens, Massachusetts"; Data Item A005/A008; prepared for Commander, U.S. Army Environmental Center; prepared by ABB Environmental Services, Inc., Portland, Maine; May 1995.

W003971.080

- ABB-ES, 1995c. "Lower Cold Spring Brook Site Investigation Report, Fort Devens, Massachusetts"; Data Item A009; prepared for Commander, U.S. Army Environmental Center; prepared by ABB Environmental Services, Inc., Portland, Maine; December 1995.
- ABB-ES, 1993a. Biological and Endangered Species Baseline Study, Fort Devens, Massachusetts; prepared for U.S. Army Corps of Engineers New England Division Waltham, Massachusetts.
- ABB-ES, 1993b. Draft Remedial Investigation Addendum Report Data Item A009, Fort Devens Group 1 A, Massachusetts; prepared for U.S. Army Environmental Center, Aberdeen Proving Ground, Maryland.
- Alden, W.C., 1925. "Physical Features of Central Massachusetts." In Contributions to the Geography of the United States, 1923-1924. U.S. Geological Survey Bulletin 760, pp. 13-106.

American Cancer Society (ACS), 1997, "Cancer Facts and Figures".

- Arthur D. Little, Inc. (ADL), 1994a. Final Storm Sewer System Evaluation (AREE 70) Evaluation. Base Realignment and Closure Environmental Evaluation. Fort Devens, Massachusetts. Prepared for the U.S. Army Environmental Center (AEC). June 1994.
- Arthur D. Little, Inc. (ADL), 1994b. Addendum 1 AREE River Evaluation. Base Realignment and Closure Environmental Evaluation. Fort Devens, Massachusetts. Prepared for the U.S. Army Environmental Center (AEC). Rev 2 June.
- Arthur D. Little, Inc. (ADL), 1994c. "Past Spill Sites (AREE 69) Supplemental Site Evaluation Data Package", Base Realignment and Closure Environmental Evaluation. Fort Devens, Massachusetts. (BRAC EE). October.
- Aquatic Information Retrieval System (AQUIRE), 1996. USEPA On-line Aquatic Database Record Search; April, (Toxicity Data) and July, (Bioconcentration Factors).

Harding Lawson Associates

W003971.080

- ATSDR, 1991. "Toxicological Profile for Manganese". Agency for Toxic Substances and Disease Registry, U.S. Public Health Service (Draft).
- Audubon, 1994. <u>National Audubon Society Field Guide to North American Birds;</u> Borzoi Books, Chanticleer Press, Inc., New York, NY.
- Baes, C.F. III, R.D. Sharp, A.L. Sjoreen, and R.W. Shor, 1984. "A Review and Analysis of Parameters for Assessing Transport of Environmentally Released Radionuclides through Agriculture"; Oak Ridge National Laboratory; ORNL-5786; September 1984.
- Bailey, N.J.L, A.M. Jobson, and M.A. Rogers, 1973. "Bacterial Degradation of Crude Oil: Comparison of field and Experimental Data"; *Chemical Geology*; Vol. II; pp. 203-211.
- Baker, R.H., 1983. Michigan Mammals; Michigan State University Press; 642 pp.
- Barbour, F. A., c. 1941. "Fort Devens, Mass. General Layout Plan"; Plan 6101-710.1B; prepared for Construction Division, Office of Quartermaster General; Scale approximately 1:7,000.
- Barnthouse, L.W., J.E. Breck, T.D. Jones, G.W. Suter II, C. Easterly, L.R. Glass, B.A. Owen, A.P. Watson, 1988. <u>Relative Toxicity Estimates and</u> <u>Bioaccumulation Factors for the Defense Priority Model</u>, Environmental Sciences Division, Oak Ridge National Laboratory, Publication No. 2997, October 1988.
- Base Realignment and Closure Commission Plan (BRAC 91), 1994. PL 101-510. May 1994.
- Beyer, W.N., 1990. "Evaluating Soil Contamination"; U.S. Fish and Wildlife Service, Biological Report 90(2).
- Biang, C.A., R.W. Peters, R.H. Pearl, and S.Y. Tsai, 1992. "Master Environmental Plan for Fort Devens, Massachusetts"; prepared for U.S. Army Toxic and Hazardous Materials Agency; prepared by Argonne National Laboratory, Environmental Assessment and Information Sciences

Division; Argonne, Illinois; April.

- Bouwer, H., and R.C. Rice, 1976. "Slug Test Method for Determining Hydraulic Conductivity of Unconfined Aquifers with Completely or Partially Penetrating Wells"; *Water Resources Research*; Vol. 12, No. 3; pp. 423-428.
- Brackley, R.A. and B.P. Hansen, 1977. "Water Resources of the Nashua and Souhegan River Basins, Massachusetts." U.S. Geological Survey Hydrologic Investigations Atlas HA-276.
- Briggs, G.G., R.H. Bromilow, and A.A. Evans, 1982, Relationships Between Lipophilicity and Root Uptake and Translocation of Non-ionised Chemicals by Barley; Pestic. Sci. (14) 495-504; January 1982.
- Briggs, G.G., R.H. Bromilow, A.A. Evans, and M. Williams, 1983. "Relationships Between Lipophilicity and the Distribution of Non-ionised Chemicals in Barley Shoots Following Uptake by the Roots"; <u>Pestic. Sci</u>. (14) 492-500; February, 1983.
- Burt, W.H., and R.P. Grossenheider, 1976. <u>A Field Guide to the Mammals</u>; Houghton and Mifflin Co., Boston, MA; 289 pp.
- Chadwick, M., Branfman, A.K. and Silveria, D.M. 1982. Dose-dependence of and Effect of Prior Exposure on the Metabolism of DEPH Administered in the Diet to Rats. Report to Chemical Manufacturers Association. Arthur D. Little, Inc.
- Chang, L.H. 1943. The Fecal Excretion of Polycyclic Hydrocarbons Following Their Administration to the Rat. J. Biol. Chem. 151: 93-102.
- DeGraaf, R.M., and D.D. Rudis, 1986. New England Wildlife: Habitat, Natural History, and Distribution; USDA Forest Service, Northeastern Forest Experiment Station, General Technical Report NE-108.
- Detrick, C.E., 1991. "Installation Assessment, Fort Devens, Ayer, Massachusetts"; prepared for U.S. Army Toxic and Hazardous Materials Agency; prepared by Environmental Photographic Interpretation Center, Environmental

Harding Lawson Associates

W003971.080

Monitoring Systems Laboratory, Warrenton, VA; September.

Ecology and Environmental, Inc. (E&E), 1994. Final Remedial Investigations Report, Functional Area I, Fort Devens, Massachusetts (Volumes I through IV). Prepared by E&E for the U.S. Army Environmental Center (AEC). August 1994.

- Engineering Technologies Associates, Inc. (ETA), 1992. "Ground Water Flow Model at Fort Devens, Massachusetts." Prepared for Commander, U.S. Army Toxic and Hazardous Materials Agency. Ellicott City, Maryland: ETA. October 30.
- Engineering Technologies Associates, Inc. (ETA), 1996. "Analysis of Well Head Protection Area Delineation of Fort Devens Water Supply Wells, Addendum One of the Detailed Flow Model for the Main and North Post, Fort Devens, Massachusetts". Prepared for Commander, U.S. Army Environmental Center. Ellicott City, Maryland: ETA.
- Environmental Criteria and Assessment Office (ECAO), 1994a. "Derivation of a Provisional Oral RfD for Aluminum." U.S. Environmental Protection Agency, Office of Research and Development, Cincinnati, Ohio.
- Environmental Criteria and Assessment Office (ECAO), 1994b. "Derivation of a Provisional Oral RfD for Iron." U.S. Environmental Protection Agency, Office of Research and Development, Cincinnati, Ohio.
- Environmental Criteria and Assessment Office (ECAO), 1994c. "Derivation of a Provisional Oral RfD for Naphthalene." U.S. Environmental Protection Agency, Office of Research and Development, Cincinnati, Ohio.
- Ehrlich, P.R., Dobkin, D.S., and Wheye.d., 198. The Birder's Handbook: A Field Guide to the Natural History of North American Birds; The Essential Companion to Your Identification Guide, New York, NY.
- Finley, B., D. Proctor, P. Scott, N. Harrington, D. Paustenback, P. Price, 1994. Recommended Distributions for Exposure Factors Frequently Used in Health Risk Assessments. Risk Analysis, Volume 14, No. 4, August 1994.

Harding Lawson Associates

- Fitchko, J., 1989. "Criteria for Contaminated Soil/Sediment Cleanup"; Pudvan Publishing Co., Inc.; Northbrook, IL.
- Fort Devens Dispatch, 1992. "The Army in New England: 75 Years in the Making." Special Edition. Vol. 52, No. 34. September 10.
- Freeze, 1979. Freeze, R. Allan, J. A. Cherry, 1979. Groundwater: Englewood Cliffs, N.J., Prentice Hall, p. 29.
- Goyer, R.A., 1991. "Toxic Effects of Metals in Casarett & Doull's Toxicology." In The Basic Science of Poisons; 4th Ed.; Eds. Amdur, M.O., J. Doull and C.D. Klaassen; New York: Pergamon Press, Inc.
- Hodson and Williams. 1988. Average of values from The Estimation of the Adsorption coefficient (K_{∞}) for Soils by High Performance Liquid Chromatography. Chemosphere. 17(1):67-77.
- Huling, S.G. 1989. "Facilitated Transport"; EPA/540/4-89/003, 5 p.
- Hvorslev, M.J., 1951. "Time Lag and Soil Permeability in Groundwater Investigations"; U.S. Army Corps of Engineers Waterways Experiments Station; Bulletin 36; Vicksburg, MS.
- Jahns, R.H., 1953. Surficial Geology of the Ayer Quadrangle, Massachusetts. Scale 1:31,680. U.S. Geological Survey.
- Jamison, V.W., R.L. Raymond, and J.O. Hudson, 1975. "Biodegradation of High Octane Gasoline in Groundwater"; *Developments in Industrial Microbiology*; Vol. 16; pp. 305-311.
- Keely, J.F. 1989. "Performance Evaluations of Pump-and-Treat Remediations"; U.S. EPA/540/4-89/005. 19 p.
- Kenaga, S.E., and C.A.I. Goring, 1978. "Relationship Between Water Solubility, Soil Sorption, Octanol-Water Partitioning, and Bioconcentration of Chemicals in Biota"; American Society of Testing Material; Third Aquatic Toxicity Symposium; New Orleans, Louisiana; p. 63.

W003971.080

- Kondakis, X.G., et al., 1989. "Possible Health Effects of High Manganese Concentration in Drinking Water." Arch. Environ. Health. 44:175-178.
- Koteff, C., 1966. "Surficial Geologic Map of the Clinton Quadrangle, Worcester County, Massachusetts." U.S. Geological Survey Map GQ-567.
- Kurz Associates, Inc., 1991. "Underground Storage Tank Removal and Related Environmental Investigations, Fort Devens Military Reservation, Harvard, Massachusetts"; prepared for Franklin Environmental Services, Inc., Wrentham, MA; Bridgewater, MA; January.
- Levine, M.B., A.T. Hall, G.W. Barrett, and D.H. Taylor, 1989. "Heavy Metal Concentrations During Ten Years of Sludge Treatment to an Old-Field Community"; J. Environ. Qual., Vol. 18, pgs. 411-418.
- Long, E.R., D.D. MacDonald, S.L. Smith, and F.D. Calder, 1994. Incidence of Adverse Biological Effects within Ranges of Chemical Concentrations in Marine and Estuarine Sediments; National Oceanic and Atmospheric Administration; September 15.
- Lyman, W. J., Reehl, W. F., and Rosenblatt, D. H. 1982. "Handbook of Chemical Property Estimation Methods." McGraw-Hill, Inc., New York.
- Mabey, W. R., Smith, J. H., Podoll, R. T., Johnson, H. L., Mill, T., Chou, T. W., Gates, J., Patridge, I. W., Jaber, H., and Vandenberg, D. 1982. "Aquatic Fate Process Data for Organic Priority Pollutants." Prepared by SRI International. Prepared for Monitoring and Data Support Division, Office of Water Regulations and Standards. Washington, D.C. EPA Contract Nos. 68-01-3867 and 68-03-2981.
- Mackay, D. and Shiu, W. Y. 1981. "A Critical Review of Henry's Law Constants for Chemicals of Environmental Interest." J. Phys. Chem. Ref. Data. 10:1175-1199.
- Mackay, D., Shiu, W. Y., Sutherland, R. P. 1979. Determination of air-water Henry's Law constants for hydrophobic pollutants. Environ. Sci. Technol. 13:333-337.

W003971.080

- MADEP, 1997. Characterizing Risks posed by Petroleum Contaminated Sites: Implementation of MADEP VPH/EPH Approach. Public Comment Draft. September 25.
- Marcoa Publishing, Inc., 1990. "Welcome to Fort Devens A Community of Excellence." San Diego, California.
- Massachusetts Department of Environmental Protection (MADEP), 1994. "Interim Final Petroleum Report: Development of Health-Based Alternative to the Total Petroleum Hydrocarbon (TPH) Parameter", Prepared for: Bureau of Waste Site Cleanup, Massachusetts Department of Environmental Protection, Boston, Ma, August.
- Massachusetts Department of Environmental Protection (MADEP, 1995). "Revised Massachusetts Contingency Plan, 310 CMR 40.000. January 1995.
- Massachusetts Department of Environmental Protection (MADEP), 1995a. "Method for the Determination of Volatile Petroleum Hydrocarbons (VPH)"; Division of Environmental Analysis; Office of Research and Standards; Bureau of Waste Site Cleanup; August 1995.
- Massachusetts Department of Environmental Protection (MADEP), 1995b. "Method for the Determination of Extractable Petroleum Hydrocarbons (EPH)"; Division of Environmental Analysis; Office of Research and Standards; Bureau of Waste Site Cleanup; August 1995.
- Massachusetts Department of Environmental Protection, 1996. "Drinking Water Standards and Guidelines for Chemicals in Massachusetts Drinking Waters"; Office of Research and Standards, Boston, MA. May 1996.
- Massachusetts Natural Heritage & Endangered Species Program, 1997. Letter from Andrea Arnold to Nancy Roka regarding state-listed rare species in the vicinity of AOC 57 and 59W.
- Maughan, J.T., 1993. Ecological Assessment of Hazardous Waste Sites; New York, NY; Van Nostrand Reinhold; p. 352.

W003971.080

- McMaster, B.N., J.D. Bonds, J.H. Wiese, K.L. Hatfield, J.B. Holly, L.C. Carter, E.A. Knauft, and K.A. Civitarese, 1982. Installation Assessment of Headquarters Fort Devens, Report No. 326. Prepared for Commander, Headquarters Fort Devens and for U.S. Army Toxic and Hazardous Materials Agency. Prepared by Environmental Science and Engineering, Inc.. Gainesville, FL. August.
- Montgomery, J.H., 1991. "Groundwater Chemical Field Guide". Lewis Publishers, Chelsea, MI. 271 p.
- National Academy of Sciences (NAS), 1977. "Drinking Water and Health"; Safe Drinking Water Committee, Washington, D.C.
- National Research Council (NRC), 1982. "Nutrient Requirements of Minks and Foxes. Second Revised Edition, 1982"; Subcommittee on Furbearer Nutrition, National Academy Press, Washington, D.C.
- National Research Council (NRC), 1984. <u>Nutrient Requirements of Poultry</u>; Subcommittee on Poultry Nutrition; National Academy Press, Washington, D.C.
- NYSDEC, 1994. "Sediment Criteria"; Bureau of Environmental Protection, Division of Fish and Wildlife, NYSDEC; July, 1994.
- OHM Remediation Services Corp., 1995. Draft Removal Action Report Study Area 57 - Area 2. Fort Devens, Massachusetts. Prepared for U.S. Army Corps of Engineers, New England Division, Waltham, Massachusetts, October 1995.
- Ohno, Y., Kawanishi, T. & Takahashi, A. 1986. Comparisons of the Toxicokinetic Parameters in Rats Determined for Low and High Doses of gamma-Chlordane. J. Toxicol. Sci. 11: 111-124.
- Peck, J.H., 1975. "Preliminary Bedrock Geologic Map of the Clinton Quadrangle, Worcester County, Massachusetts." Scale 1:24,000; text and three maps. U.S. Geological Survey Open-File Report 75-658.

W003971.080

- Peck, J.H., 1976. "Silurian and Devonian Stratigraphy in the Clinton Quadrangle, Central Massachusetts." In Contributions to the Stratigraphy of New England. Geological Society of America Memoir 148.
- Persaud, D., R. Jaagumagi, and A. Hayton, 1996. "Guidelines for the Protection and Management of Aquatic Sediment Quality in Ontario"; Ontario Ministry of the Environment, Queen's Printer for Ontario; 23 pp.
- Potomac Research Inc. (PRI), 1993. User's Manual, IRDMIS PC Data Entry and Validation Subsystem. Version 5.0. Prepared for U.S. Army Toxic and Hazardous Materials Agency, Aberdeen Proving Ground, Maryland. February.
- Rand, G.M. and S.R. Petrocelli, 1985. <u>Fundamentals of Aquatic Toxicology:</u> <u>Methods and Applications</u>; Hemisphere Publishing Corporation, New York, 1985, 666 pp.
- Rathburn, R. E. and Tai, D. Y. 1982. "Volatilization of Ketones from Water." Water, Air and Soil Poll. 281-293
- Robinson, G.R., Jr., 1978. "Bedrock Geologic Map of the Pepperell Shirley, Townsend Quadrangles, and Part of the Ayer Quadrangle, Massachusetts and New Hampshire." Miscellaneous Field Studies Map MF-957. U.S. Geological Survey.
- Robinson, P. and R. Goldsmith, 1991. "Stratigraphy of the Merrimack Belt, Central Massachusetts." In *The Bedrock Geology of Massachusetts*, U.S. Geological Survey Professional Paper 1366-G. pp. 61-637.
- Rojko, A.M., 1990. "Heavy Metals in the Sediments of Massachusetts Lakes and Ponds." Unpublished Master's Thesis; Northeastern University: Department of Civil Engineering; June.
- Rowen, D.J., and J.B. Rasmussen, 1992. "Why Don't Great Lakes Fish Reflect Environmental Concentrations of Organic Contaminants? An Analysis of Between-lake Variability in the Ecological Partitioning of PCBs and DDT." J. Great Lakes Res.; 18(4): 724-741.

W003971.080

- Russell, S.L. and R.W. Allmendinger, 1975. "Interim Geologic Map of the Shirley Quadrangle, Massachusetts." U.S. Geological Survey Open File Report 76-267.
- Ryan, J.A., R.M. Bell, J.M. Davidson, and G.A. O'Connor, 1988. "Plant Uptake of Non-ionic Organic Chemicals from Soils"; <u>Chemosphere</u>, Vol. 17, No. 12, pp. 2299-2323; 0045-6535/88; Great Britain; 1988.
- Shen, T. J. 1982. Air quality assurance for land disposal of industrial wastes. Environmental Management. 6:297-305.
- Soil Conservation Service, undated. "General Soil Map of Middlesex County, Massachusetts"; U.S. Department of Agriculture; Middlesex Conservation District; unpublished.
- Soil Conservation Service, 1985. "Soil Survey of Worcester County, Massachusetts, Northeastern Part"; U.S. Department of Agriculture.
- Soil Conservation Services, 1989. Untitled Middlesex County field sheet #19; U.S. Department of Agriculture; Middlesex Conservation District; January 5.
- Soil Conservation Service, 1991. "Middlesex County Massachusetts Interim Soil Survey Report"; U.S. Department of Agriculture; Middlesex Conservation District; Westford, MA (includes Soil Sheet #19).
- Suter, Glen W., 1993. Ecological Risk Assessment; Lewis Publishers, Chelsea Michigan.
- Tinsley, I.J., 1979. Chemical Concepts in Pollutant Behavior; John Wiley and Sons; New York, New York.
- Todd, D.K., 1980. <u>Groundwater Hydrology</u>; University of California, Berkeley and David Keith Todd, Consulting Engineers, Inc.; John Wiley & Sons, New York, New York. 1980.

W003971.080

- Travis, C.C., and A.D. Arms, 1988. "Bioconcentration of Organics in Beef, Milk, and Vegetation"; <u>Environ. Sci. Tech.</u>; Vol. 22; 271-274.
- Tseng, W.P., H.M. Chu, S.W. How, J.M. Fong, C.S. Lin, and S. Yen, 1968. "Prevalence of Skin Cancer in an Endemic Area of Chronic Arsenicism in Taiwan." J. Natl. Cancer Inst. 40:453-463.
- Tseng, W.P., H.M. Chu, S.W. How, J.M. Fong, C. S. Lin, and S. Yen. 1968. "Prevalence of Skin Cancer in an Endemic Area of Chronic Arsenicism in Taiwan"; J. Natl. Cancer Inst., Vol. 40, No. 3, pp. 453-463.
- U.S. Army Environmental Center (USAEC), 1987. "Geotechnical Requirements for Drilling, Monitoring Wells, Data Acquisition, and Reports"; March 1987.
- U.S. Army Environmental Center, 1990. "Quality Assurance Manual"; Aberdeen Proving Ground, MD; January.
- U.S. Army Environmental Center (USAEC), 1990. "USAEC QA Program"; January 1990.
- U.S. Army Environmental Center, 1993. Fact Sheet: Fort Devens Installation Restoration Program (IRP) and BRAC 91 Environmental Restoration Impact. March.
- U.S. Army Toxic and Hazardous Materials Agency (USATHAMA), 1990. Quality Assurance Program. USATHAMA PAM 11-41; Aberdeen Proving Ground, MD; January 1990.
- U.S. Army Toxic and Hazardous Materials Agency (USATHAMA), 1990. Quality Assurance Program. PAM-41. January.
- U.S. Department of the Army, 1979. Environmental Impact Statement, Fort Devens Mission Activities, Fort Devens, Massachusetts. Headquarters, U.S. Army Forces Command. June 30; Revised May 1, 1980.

W003971.080

- U.S. Engineer Office, 1952. "Master Plan, Fort Devens, Ayer, Mass., Detail Site Plan, Building Use Map"; Sheets Nos. 6 through 21; Boston, MA; Files X100-109/705 through X100-109/720.
- U.S. Environmental Protection Agency (USEPA), 1983. "Methods for the Chemical Analysis of Water and Wastes" Environmental Monitoring and Support Laboratory; USEPA 600-4-79-020; Cincinnati, OH; March 1983.
- U.S. Environmental Protection Agency (USEPA), 1983. Guidelines for Deriving Numerical Water Quality Criteria for the Protection of Aquatic Life and Its Uses. Draft. July.
- U.S. Environmental Protection Agency (USEPA), 1984. "Guidance for Preparation of Combined Work/Quality Assurance Project Plans for Environmental Monitoring"; OWRS; QA-1; May 1984.
- U.S. Environmental Protection Agency (USEPA), 1986. "Test Methods for Evaluating Solid Waste:; Laboratory Manual Physical/Chemical Methods; Office of Solid Waste and Remedial Response; Washington, D.C., SW-846; November 1986.
- U.S. Environmental Protection Agency (USEPA), 1986. *Quality Criteria for Water* 1986. Office of Water Regulations and Standards, Washington, D.C., EPA 440/5-86-001. May 1, 1986.
- U.S. Environmental Protection Agency (USEPA), 1986. Superfund Public Health Evaluation Manual. Exhibit C-1.
- U.S. Environmental Protection Agency (USEPA), 1986 Hazard Evaluation Division Standard Evaluation Procedure: Ecological Risk Assessment; Office of Pesticide Programs; EPA 540/9-85-001; Washington, D.C.
- U.S. Environmental Protection Agency (USEPA), 1988. "Region 1 Laboratory Data Validation Functional Guidelines For Evaluating Organic Analyses"; Hazardous Site Evaluation Division; November 1988.

W003971.080

- U.S. Environmental Protection Agency (USEPA), 1988a. "Region 1 Laboratory Data Validation Functional Guidelines for Evaluating Organic Analyses"; Hazardous Site Evaluation Division; November 1988.
- U.S. Environmental Protection Agency, 1988b. "Guidance for Conducting Remedial Investigations and Feasibility Studies Under CERCLA"; USEPA Office of Emergency and Remedial Response, EPA/540/G-89/004, OSWER Directive 9355.3-01; October.
- U.S. Environmental Protection Agency, 1988c. "Agency Policy on the Carcinogenic Risk Associated with Inorganic Arsenic"; memorandum from Lee M. Thomas, Office of the Administrator for Pesticides and Toxic Substances; Washington, D.C.; June 21.
- U.S. Environmental Protection Agency (USEPA), 1989a. "Risk Assessment Guidance for Superfund, Volume 1, Human Health Evaluation Manual (Part A)"; Office of Emergency and Remedial Response; EPA-540/1-89/002 (interim final); Washington, D.C., December.
- U.S. Environmental Protection Agency (USEPA), 1989b. Exposure Factors Handbook; Office of Health and Environmental Assessment; EPA-600/8-89/043; Washington, D.C.
- U.S. Environmental Protection Agency (USEPA), 1989a. "Transport and Fate of Contaminants in the Subsurface". EPA/625/4-89/019, 148 p.
- U.S. Environmental Protection Agency (USEPA), 1989b. "Risk Assessment Guidance for Superfund". Volume I: Human Health Evaluation Manual. Interim Final. OSWER Directive 9285.7-01a. September 29, 1989. Office of Emergency and Remedial Response, Washington, D.C. EPA/540/1-89/001.
- U.S. Environmental Protection Agency (USEPA), 1989c. Exposure Factors Handbook. Office of Health and Environmental Assessment, Washington, D.C. July.

W003971.080

- U.S. Environmental Protection Agency (USEPA), 1989d. "Region 1 Laboratory Data Validation Functional Guidelines for Evaluating Inorganic Analyses;" Hazardous Site Evaluation Division; February 1989.
- U.S. Environmental Protection Agency (USEPA), 1990. "Basics of Pump-and-Treat Groundwater Remediation Technology"; Environmental Research Laboratory, Ada, OK, EPA/600/8-90/003.
- U.S. Environmental Protection Agency (USEPA), 1990. Code of Federal Regulations, Title 40, Part 300, National Oil and Hazardous Substances Pollution Contingency Plan; Federal Register; March 8.
- U.S. Environmental Protection Agency (USEPA), 1991. "Risk Assessment Guidance for Superfund, Volume I: Human Health Evaluation Manual, Supplemental Guidance, Standard Default Exposure Factors"; Office of Emergency and Remedial Response, Toxics Integration Branch; OSWER Directive 9285.6-03 (interim final); Washington, D.C.
- U.S. Environmental Protection Agency (USEPA), 1991a. "National Functional Guidelines for Organic Data Review"; Washington, D.C.
- U.S. Environmental Protection Agency (USEPA), 1991b. Risk Assessment Guidance for Superfund. Volume I: Human Health Evaluation Manual Supplemental Guidance. Standard Default Exposure Factors. Interim Final. Washington, D.C. OSWER Directive 9285.6-03. March 25, 1991.
- U.S. Environmental Protection Agency (USEPA), 1991c. Risk Assessment Guidance for Superfund. Volume I: Human Health Evaluation Manual (Part B, Development of Risk-based Preliminary Remediation Goals). Publication 9298.7-013. December 1991 and Revision October 1992.
- U.S. Environmental Protection Agency (USEPA) Region I and the U.S. Department of the Army, 1991e. In the Matter of: The U.S. Department of the Army, Fort Devens Army Installation, Fort Devens, Massachusetts. Federal Facility Agreement Under the Comprehensive Environmental Response, Compensation, and Liability Act, Section 120. May 13.

W003971.080

- U.S. Environmental Protection Agency (USEPA), 1992a. "Risk Updates"; USEPA Region I, Waste Management Division; December.
- U.S. Environmental Protection Agency (USEPA), 1992b. "Supplemental Guidance to RAGS, Calculating the Concentration Term"; Office of Solid Waste and Emergency Response; Intermittent Bulletin, Vol. 1, No. 1; Washington, D.C.; May.
- U.S. Environmental Protection Agency (USEPA), 1992c. "Guidance for Data Useability in Risk Assessment (Parts A and B)"; Office of Emergency and Remedial Response; Publication 9285.7-09A; Washington, D.C.; April.
- U.S. Environmental Protection Agency (USEPA), 1992d. Dermal Exposure Assessment: Principles and Applications; Office of Research and Development; EPA/600/9-91/011B; Washington, DC
- U.S. Environmental Protection Agency (USEPA), 1992a. Guidelines for Exposure Assessment. Federal Register 57:22888-22938.
- U.S. Environmental Protection Agency (USEPA), 1992b. Supplemental Guidance to RAGS: Calculating the Concentration Term. Office of Solid Waste and Emergency Response. Publication 9285.7-081. May 1992.
- U.S. Environmental Protection Agency (USEPA), 1992c. Dermal Exposure Assessment: Principles and Applications. Interim Report. Office of Research and Development. EPA/600/8-91/011B. January.
- U.S. Environmental Protection Agency (USEPA), 1993. Data Quality Objective Process for Superfund, Interim Final Guidance, EPA/540/G-93/071, Publication 9355.9-01, September 1993.
- U.S. Environmental Protection Agency (USEPA), 1994a. Region I Risk Updates Number 2. USEPA Region One Waste Management Division. August 1994.
- U.S. Environmental Protection Agency (USEPA), 1994b. "Revised Interim Soil Lead Guidance for CERCLA Sites and RCRA Corrective Action

W003971.080

Facilities"; Office of Solid Waste and Emergency Response, OWSER Directive 355.4-12; July 1994.

- U.S. Environmental Protection Agency (USEPA), 1995. "Risk Updates"; USEPA Region I, New England, Waste Management Division; No. 3; August.
- U.S. Environmental Protection Agency (USEPA), 1995a. Risk Updates. Number 3. EPA New England. August 1995.
- U.S. Environmental Protection Agency (USEPA), 1995b. "Health Effects Assessment Summary Tables (HEAST)", Annual FY 1995. Office of Solid Waste and Emergency Response EPA/R-95-036, and the November 1995 Supplemental Update.
- U.S. Environmental Protection Agency (USEPA), 1995c. Supplemental Guidance to RAGS: Region 4 Bulletins, November.
- U.S. Environmental Protection Agency (USEPA), 1996. "Test Methods for Evaluating Solid Waste"; Laboratory Manual Physical/Chemical Methods; Office of Solid Waste and Emergency Response; Washington, DC; SW-846; November 1986; Revision 3- January 1995; Revision 4 - December 1996.
- U.S. Environmental Protection Agency (USEPA), 1996. "Risk Updates"; USEPA New England, Waste Management Division; No. 4; November.
- U.S. Environmental Protection Agency (USEPA), 1996. "Region 1 EPA-NE Data Validation Guidelines For Evaluating Environmental Analyses"; Quality Assurance Unit Staff; Office of Environmental Measurement and Evaluation; December 1996
- U.S. Environmental Protection Agency (USEPA), 1996a. Risk-Based Concentration Table. Prepared by Roy L. Smith, USEPA Region III Technical Program Support Branch. April 19, 1996.
- U.S. Environmental Protection Agency (USEPA), 1996b. Drinking Water Regulations and Health Advisories. Office of Water, Washington, DC. February 1996.

Harding Lawson Associates

W003971.080

- U.S. Environmental Protection Agency (USEPA), 1996c. "Integrated Risk Information System (IRIS); On-Line Service. June 1996.
- U.S. Environmental Protection Agency (USEPA), 1996d. <u>Soil Screening</u> <u>Guidance: Technical Background Document</u>. Office of Solid Waste and Emergency Response, EPA/540/R-95/128, May.
- U.S. Environmental Protection Agency (USEPA), 1997a. Drinking Water Regulations and Health Advisories. Office of Water. EPA822-B-96-002. May.
- U.S. Environmental Protection Agency (USEPA), 1997b. Risk-Based Concentration Table. Prepared by Roy L. Smith, USEPA Region III Technical Program Support Branch. January, 1997
- U.S. Environmental Protection Agency (USEPA), 1997c. USEPA Region III Risk-Based Concentration Table. September.
- U.S. Environmental Protection Agency (USEPA), 1997d. "Integrated Risk Information System (IRIS); On-Line Service. January 1997.
- U.S. Environmental Protection Agency (USEPA), 1997e. Health Effects Assessment Summary Tables (HEAST), Annual Update: Office of Solid Waste and Emergency Response, EPA 540/R/97/036, PB97-921199
- U.S. Environmental Protection Agency (USEPA), 1997. Ecological Risk Assessment Guidance for Superfund: Process for Designing and Conducting Ecological Risk Assessments; Interim Final U.S. EPA Document; Environmental Response Team, Edison, New Jersey; June 2, 1997
- U.S. Environmental Protection Agency (USEPA), 1998. "Integrated Risk Information System (IRIS)"; on-line data-base search; February.
- U.S. Fish and Wildlife Service (USFWS), 1992. "Survey and Evaluation of Wetlands and Wildlife Habitat, Fort Devens, Massachusetts. House of Representatives Appropriations Committee. p. 1 - 10.

W003971.080

- Vahter, M. 1983. Metabolism of Arsenic. In: Fowler, B.A., ed. Biological and Environmental Effect of Arsenic. NY: Elsevier. pp. 171-198.
- Vanasse Hangen Brustlin, Inc., 1994. Devens Reuse Plan. Prepared for the Boards of Selectmen of the Towns of Ayer, Harvard, Lancaster, and Shirley and the Massachusetts Government Land Bank. November 14, 1994.
- Walton, W.C., 1988. "Practical Aspects of Groundwater Modeling". 3rd Edition. NWWA, 587 p.
- Weiner, J., USFWS, Lacrosse, Wisconsin, 1993. "BAFs for Inorganic Analytes in Freshwater Systems, Plants, and Animals"; Personal communication with Cornell Rosiu of ABB-ES; May.
- Wentsel, R.S., T.W. La Point, M. Simini, R.T. Checkai, D. Ludwig, and L.W. Brewer, 1996. <u>Tri-Service Procedural Guidelines for Ecological Risk</u> <u>Assessments, Volumes I and II</u>; written for the Air Force Center for Environmental Excellence (AFCEE), the Army Environmental Center (AEC), and the Naval Facilities Engineering Service Center (NFESC); June, 1996.
- Wones, D.R. and R. Goldsmith, 1991. Intrusive Rocks of Eastern Massachusetts. U.S. Geological Survey Professional Paper 1366-I.
- Zen, E-an, Ed., 1983. "Bedrock Geologic Map of New England." U.S. Geological Survey; Scale 1:250,000; three sheets.

W003971.080

TABLE ES-1 SUMMARY OF HUMAN HEALTH RISK ASSESSMENT AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUGETTS

		CENTRAL T	and a set of the local data was set of the l	RMI		ARE SITE RISKS UNACCEPTABLE?	
EXPOSURE MEDIUM	RECEPTOR	Total Cancer	Totel Hazerd	Total Canoar	Total Hazard	Cencer Risk (greater than	Non-Cencer Rie (greater than
		Riek	Index	Risk	index	1x10 ⁴ - 1x10 ⁴) 1	$HI = 1)^2$
CHILD TRESPASSER: Current Lend	Use						
SURFACE	SOIL:	3x10*	0.1	6x10 ⁴	0.2	NO	NO
SEDIMEN	<u>n</u>	5x10 ⁻⁷	0.05	1x10 ⁴	0.07	NO	NO
GROUND	NATER (Discharge to Surface Water) :	1×10*	0.2	2x10*	0.2	NO	NO
	TOTAL CHILD TRESPASSER RISK:	6x10*	0.4	1x10 ⁴	0.6	NO	NO
SITE MAINTENANCE WORKER: Cu	ment Land USe						
SURFACE	SOIL:	1 x 10 ⁻⁶	0.07	5x10 ⁴	0,1	NO	NO
PUPIL: Future Land Vee							
SURFACE	BOIL:	5x10 ⁻⁴	0.3	9x10 ⁻⁴	0.3	NO	NO
SEDIMEN	<u>r:</u>	5x10-7	0.05	1x10 ⁴	0.07	NO	NO
GROUND	NATER (Discharge to Surface Water) :	1×10 ⁻⁵	0.2	2x10*	0.2	NO	NO
INDOOR A	MR:	NC	0.4	NC	0.4	NO	NO
	TOTAL PUPIL RISK:	6x10 ⁻⁶	1	1x10*	1	NO	NO
EXCAVATION WORKER: Future La	ind Use						
SURFACE	BOIL:	1x10 ⁻⁷	0.1	3x10"	0.2	NO	NO
SUBSURF	ACE SOIL:	6x10 ⁻⁸	0,8	1x10'7	0.9	NO	NO
	TOTAL EXCAVATION WORKER RISK:	2x10 ⁻⁷	- A	4x10 ⁻⁷	1	NO	NO
ADULT RESIDENT: Future Lend Ue	•		-				
GROUND	NATER HYPOTHETICAL POTABLE USE 3	1×10 ⁻⁴	4	3x10 ⁻³	25	YES	YES
CHILD RESIDENT: Future Land Use							
GROUND	NATER HYPOTHETICAL POTABLE USE 2	8x10 ⁴	8	2x10 ⁻³	67	YES	YES
	TOTAL RESIDENT RISK:	2x104	+	3x10 ⁻³	-	YES	YES

NOTES:

1 According to the National Contigency Plan for Superfund Sites, the acceptable cancer risk range is within or below 1 in 10,000 (1x10⁴) to 1 in 1 million (1x10⁴).

2 According to the National Contigency Plan for Superfund Sites, the acceptable non-cancer risk is a chemical does that will not

result in adverse hearth effects to sensitive subpopulations; this is often interpreted by the USEPA to be a Hi of not greater than 1.

3 Groundwater is not presently, nor will be in the future, used as a source of residential or industrial supply water.

Therefore, this evaluation represents a theortical exposure which does not and will not occur. 4 Includes 2-methylheptane, nepththalone, extractable petroleum hydrocarbone, volatile petroleum hydrocarbone

5 Includes bis(2-sthythexyl)phthelate, sluminum, Iron, manganese, chloroform, trichloroethene

RME - Researable Maximum Exposure bgs - below ground surface HI - Hazard Index

Table ES-2 Ecological Risk Assessment Summary AOC 69W Remedial Investigation Devens, Massachusetts							
Receptor		Medium					
	Surface Soil	Groundwater	Sediment				
Small Mammals	Negligible	NA	None				
Small Birds	None	NA	None				
Predatory Mammals	None	NA	None				
Terrestrial Plants	Pb at ZWS-95-42X? No signs of stressed vegetation	NA	NA				
Soil Invertebrates	None	NA	NA				
Aquatic Organisms	NA	Fe and Mn ¹ . Negligible risk from other analytes	Negligible. Adverse effects observed in toxicity tests may be associated with low habitat quality				

3

.

÷.,

TABLE 3-1 TARGET COMPOUNDS AND REPORTING LIMITS 1995 AND 1996 FIELD SCREENING PROGRAM AOC 69W

Vear	Target Analyte	Soil µg/g	Water µg/L
995	Benzene	0.002	2
	Toluene	0.002	2
	Ethylbenzene	0.002	2
	m/p-xylene	0.002	4
	o-xylene	0.002	2
	Tetrachloroethene	0.002	2
	Trichloroethene	0,002	2
	cis-1,2-dichloroethene	0.002	2
	trans-1,2-dicholorethene	0.002	2
	1,1-dichloroethene	0.002	5
	Chloroform	0.002	2
	1,1,1-trichloroethane	0.002	2
	Carbontetrachloride	0.002	2
	Chlorobenzene	0.002	2
	TPH-GRO	0.1	100
	TPH-DRO	100	NA
	TPHC-IR	50	NA
996	Benzene	0.25	2
	Toluene	0.25	2
	Ethylbenzene	0.25	2 4
	m/p-xylene	0.5	
	o-xylene	0.25	2
	Tetrachloroethene	0.25	2
	Trichloroethene	0.25	2
	cis-1,2-dichloroethene	0.25	2 2
	trans-1,2-dicholorethene	0.25	2
	Vinyl chloride	0.25	2
	1,1-dichloroethene	0.25	2
	Chloroform	0.25	2 2
	1,1,1-trichloroethane	0.25	2
	Carbontetrachloride	0.25	2
	Chlorobenzene	0.25	2
	Naphthalene*	0.25	2
	1,2-dichlorobenzene*	0.25	2
	1,3-dichlorobenzene*	0.25	2
	1,4-dichlorobenzene*	0.25	2
	TPHC-IR	50	NA

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes:

NA = soil not analyzed

 $\mu g/g = microgram per gram$

 $\mu g/L = microgram per liter$

* = added to list during 1996 field program

TABLE 3-2 USAEC DATA FLAGS AND QUALIFIERS AOC 69W

Measurement Boolean	Concentration	(Upper case letters)	Flagging Codes (Lower case letters or #)	
ND	300	J	adf	
easurement Boole	1010			
easin entent book	< = Concentratio ND = Not detecta	ble above the ind n the maximum c	certified concentration	
Data Qualifier:	8			
	 ? = Control char has been uploade I = The low spik M = The high spi J = The low spik K = Missed holdi L = Missed analy N = The high spi O = Low spike red 	d to the database, the recovery for the the recovery for the recovery for the recovery for the recovery for the sis holding time the recovery for the coveries excessive	this lot was high is lot was low iction or preparation this lot was low vely different	
	R = Data is reject	ted and is not use:	eable	
Flagging Code	1 = Result was les 2 = Ending calibra 3 = Internal stand 7 = Low spike recov exceeded the certi a = Analyte found b = Analyte found c = Analyte found c = Analyte found c = Analyte found c = Analyte found h = Lot out of confi i = Interences in t j = Value is estim k = Reported resu I = Out of control m= High duplicato n = Tentatively-id p = Value is less th q = Confirmatory r = Non-target and Analyte was not p s = Non-target confirmation u = Analyte is un v = Sample was not p	ation not within a ard not within acc overy not within ery outside certifi ified range by <1. I in trip blank as v I in method blank confirmed by a di lysis iltered prior to and lysis iltered prior to and lysis trol but data acce he sample caused hated hts affected by int . Data rejected du e spike not within entified compound han the method re analysis was perf alyte analyzed for erformance demo mpound analyzed inconfirmed. Confi tot correctly prese	sceptable limits a control limits fied range but within acceptable limits. This code is used when analyte concentrations L5 % and the laboratory felt a dilution was not warranted well as the sample k or QC sample as well as the sample. lifferent column or technique. halysis sate blank as well as the sample epted due to high recoveries d the quantitation and/or identification to be suspect terferences or high background. An elevated quantitation limit is reported hue to low recoveries	d

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Federal Standards and Guidance									
Analyte			Ambient							
	Safe Drinking Wate	er Act (SDWA) ^(a)		otection an Health	For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)				
	ARAR Drinking Water MCL (μg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)					
Volatile Organics						1				
acetone	-	4			-/-	3,700N				
benzene	5	zero	0.66	40	5,300/-2	0.36C				
carbon tetrachloride	5	zero	0.4	6.94	35,200/-2	0.16C				
chloroform (THM)	100/803	zero	0.19	15.7	28,900/1,240 ²	0.15C				
ethylbenzene	700	700	1,400	3,280	32,000/-2	1,300N				
styrene	100	100	4	-	-/-	1,600N				
1,1,2,2-tetrachloroethane			0.17	10.7	$-/2,400^{2}$	0.052C				
tetrachloroethylene	5	zero	0.8	8.85	5,280/840 ²	1.10C				
toluene	1,000	1,000	14,300	424,000	17,500/-2	750N				
1,1,1-trichloroethane	200	200	18,400	1,030,000	-/-	1,300N				
trichloroethylene	5	zero	2.7	80.7	45,000/21.900 ²	1.6C				
trichlorofluoromethane				4	-/-	1,300N				
xylenes (total)	10,000	10,000			-/-	12,000N				

+

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Federal Standards and Guidance								
Analyte			Clean Water Act (CWA) Ambient Water Quality Criteria (AWQC) ^(b)						
	Safe Drinking Wat	er Act (SDWA) ^(a)	For Protection of Human Health		For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)			
	ARAR Drinking Water MCL (μg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (μg/L)	1.9.21			
Semivolatile Organics				E.					
acenaphthene	-	-	4		-/-	4			
anthracene				-	-/-	11,000N			
bis(2-ethylhexyl)phthalate	6	zero			-/-	4.80C			
benzo(a)anthracene	1			1	-1-	0.092C			
benzo(a)pyrene	2	zero		-	-/-	0.092C			
benzo(b)fluoranthene					-/-	0.092C			
benzo(g,h,i)perylene	1			2	-/-				
benzo(k)fluoranthene	-	I			-/-	0.092C			
benzyl alcohol	-				-	11,000N			
carbazole	4	- 1	1	÷	-/-	3.4C			
chrysene	-	-	-		-/-	9.2C			
dibenzofuran	-				-/-	150N			
di-n-butyl phthalate		-			*	3,700N			

continued

	Federal Standards and Guidance								
Analyte			Clean Water Act (CWA) Ambient Water Quality Criteria (AWQC) ^(b)						
	Safe Drinking Water Act (SDWA) ^(a)		For Pro of Huma	otection In Health	For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)			
	ARAR Drinking Water MCL (μg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)	(All of the second s			
fluoranthene	-	4	-	4	-/-	1,500N			
fluorene			s	1997 - A.	-/-	1,500N			
indeno(1,2,3-c,d)pyrene	-	-	· · ·	-	-/-	0.092C			
2-methylnaphthalene	-	4	-	4	-/-	1			
naphthalene	1			-	2,300/620 ²	1,500N			
n-nitrosodiphenylamine	4	(*)	4.9	16.1	-/-	14C			
phenanthrene		2		-	30/6.35	1			
pyrene	-		4	-1. k.	· -/-	1,100N			
Inorganics									
aluminum	-	50 to 200 ⁸			-/-	37,000N			
antimony	6	6 ⁵	146	45,000	88/30 ⁵	15N			
arsenic	50 ¹		0.0022	0.0175	360/1902,7	11N/0.038C			
barium	2,000	2,000	1,000		-/-	2,600			
beryllium	44	4	0.0037	0.0641	130/5.3 ²	0.016C			

Table 4-1 Federal ARARS and TBC Guidance - Groundwater and Surface Water AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Federal Standards and Guidance								
			Clean Water Act (CWA) Ambient Water Quality Criteria (AWQC) ^(b)						
Analyte	Safe Drinking Wat	er Act (SDWA) ^(a)		otection In Health	For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)			
	ARAR Drinking Water MCL (µg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)				
cadmium	5	5	10		3.9/1.14	18N			
calcium	* +				-1-				
chromium (total)	100	100		4	1,700/2104,9	180			
cobalt				4	÷/-	220			
copper	TT ¹⁰	1,300			18/124	1,400N			
iron	- S	300 ⁸	300	-	-/1,000				
lead	TT'	zero	50	•	83/3.24				
magnesium	-			4	-/-	-			
manganese	14	50 ⁸	50	100	-1-	1,800N			
mercury	2	2	0.144	0.146	2.4/0.012	11N			
nickel	1005	100 ⁵	13.4	100	1,400/1604	730N			
potassium -	-			÷	-/-				
selenium	50	50	10	-	20/5	180N			
silver	-	1008	50		4.1/0.124,6	180N			

continued

٠

4

.

TABLE 4-1 FEDERAL ARARS AND TBC GUIDANCE - GROUNDWATER AND SURFACE WATER AOC 69W

	Federal Standards and Guidance								
Analyte			Clean Water Act (CWA) Ambient Water Quality Criteria (AWQC) ^(b)						
	Safe Drinking Wat	er Act (SDWA) ^(a)	For Protection of Human Health		For Protection of Aquatic Life	TBC Region III Tap Water (µg/L)			
	ARAR Drinking Water MCL (µg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)				
sodium			6	-	-/-				
vanadium -						260N			
zinc		5,000 ⁸	-	-	12/1104	11,000N			
Pesticide/PCBs									
DDT			0.000024	0.000024	1.1/0.001	0.2C			
DDD	4		4	1	-1-	0.28C			
DDE			18		1,050/-2	0.2C			
endrin	2	2	1.0	-	0.18/.0023	11N			
alpha chlordane	213	zero ¹³	0.0004613	0.0004813	2.4/0.0043 ¹³	0.05213			
gamma chlordane	213	zero ¹³	0.0004613	0.00048 ¹³	2.4/0.004313	0.052C ¹³			
heptachlor	0.4	zero	0.00028	0.00029	0.52/.0038	0.0023C			
PCB 1248	0.514	zero ¹⁴	0.00007914	0.00007914	2.0/0.01414	0.0087C			
PCB 1254	0.514	zero ¹⁴	0.00007914	0.00007914	2.0/0.01414	0.73C ¹⁴			
PCB 1260	0.514	zero ¹⁴	0.00007914	0.00007914	2.0/0.01414	0.0087C ¹⁴			

	Federal Standards and Guidance								
Analyte	Safe Drinking Water Act (SDWA) ^(a)		Ambient						
				otection an Health	For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)			
	ARAR Drinking Water MCL (µg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)				
Explosives									
cycloetramethylenetetranitramine (HMX)	¢.	•		4	-	(19)			
cyclonite (RDX)				-	-	4.00			
2,4-dinitrotoluene		-		-	4	73N			
2,6-dinitrotoluene						37N			
nitroglycerine		-	-		7	e			
2,4,6-trinitrotoluene		4				2.2C			
Cations/Anions									
chloride	12	250,000 ⁸	18		860K/230K				
phosphate		-			-/-				
sulfate	A.	500,000 ¹⁵ /250,000 ⁸	200		-/-	-			
alkalinity	1	*	-		-/20,000	An Anna			

141

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

			Federal Standards	and Guidance		
Analyte				Clean Water Act (CWA) Ambient Water Quality Criteria (AWQC) ^(b)		
	Safe Drinking Water Act (SDWA) ^(a)		For Protection of Human Health		For Protection of Aquatic Life	TBC Region III Tap Water (μg/L)
	ARAR Drinking Water MCL (µg/L)	TBC Drinking Water MCLG (µg/L)	ARAR Water and Fish Consumption (µg/L)	ARAR Fish Consumption Only (µg/L)	ARAR Fresh Water Acute/Chronic (µg/L)	
Other						-
nitrate/nitrite as N	10,000/1,00012		10,000/-			58,000N/3,100N
ТРН	-					-
	summary" Office of Science and Tech					
1994 Proposed rule for Disinfecta Hardness dependent criterion (100 Standard is being remanded. Proposed level, freshwater acute - Values presented are for trivalent Non-enforceable secondary regula Values presented are for hexavale Treatment technique action level 1 Treatment technique action level 1	nt Level	 Treatment te No federal o Noncarcinog Carcinogenic carcinogenic bestrved Effect Level (LOEL). otal for all THMs combined w odor, taste). 	chnique required. r state guidance criteria or stan enic effects effects	ndards exist.	own for carcinogenspresent a	one-in-a-million increm

continued

TABLE 4-2 STATE ARARS AND TBC GUIDANCE - GROUNDWATER AOC 69W

	MASSACHUSETTS STANDARDS AND GUIDANCE					
ANALYTE	MMCL/ORSG DRINKING WATER ^(A) (µg/L)	CLASS I GROUNDWATER ⁽⁸⁾ (µg/L)				
Volatile Organics						
acetone	3000 ²					
benzene	5					
carbon tetrachloride	5					
chloroform	52	100 ³				
ethylbenzene	700					
styrene	100					
1,1,2,2-tetrachloroethane	÷					
tetrachloroethylene	5	14				
toluene	1000					
1,1,1-trichloroethane	200					
trichloroethylene	5					
trichlorofluoromethane						
xylenes (total)	10,000					
Semivolatile Organics						
acenaphthylene	-					
anthracene						
bis(2-ethylhexyl)phthalate	6	1				
benzo(a)añthracene						
benzo(a)pyrene	0.2	1				
benzo(b)fluoranthene	*					
benzo(g,h,i)perylene		1				
benzo(k)fluoranthene	÷.	-				
benzyl alcohol						
carbazole		-				
chrysene		4				
dibenzofuran	8	*				
di-n-butyl phthalate						
fluoranthene						
fluorene						
indeno(1,2,3-c,d)pyrene	-	4				
2-methylnaphthalene						
naphthalene						
n-nitrosodiphenylamine						
phenanthrene						

TABLE 4-2 STATE ARARS AND TBC GUIDANCE - GROUNDWATER AOC 69W

	MASSACHUSETTS STANDARDS AND GUIDANCE			
ANALYTE	MMCL/ORSG DRINKING WATER ^(A) (µg/L)	CLASS I GROUNDWATER ^(B) (µg/L)		
pyrene		4 · · · ·		
Inorganics	*			
aluminum	50 to 200 ⁽¹⁰⁾	-		
antimony	6	4		
arsenic	50	50		
barium	2,000	1,000		
beryllium	4			
cadmium	5	10		
calcium	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	÷.		
chromium (total)	100	50		
cobalt		÷.		
copper	1,300	1,000		
iron	30010	300		
lead	15	50		
magnesium		÷		
manganese	50(10)	50		
mercury	2	2		
nickel	100	*		
potassium				
selenium	50	10		
silver	10010	50		
sodium	20.000			
vanadium				
zinc	5,00010	5,000		
Pesticides/PCBs				
DDT	÷.			
DDD		10.000		
DDE				
endrin	2 ⁸	0.2		
alpha chlordane	2			
gamma chlordane	2			
heptachlor	0.4	4		
PCB 1248	0.59			
PCB 1254	0.59			
PCB 1260	0,59			

TABLE 4-2 STATE ARARS AND TBC GUIDANCE - GROUNDWATER **AOC 69W**

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	MASSACHUSETTS STANDARDS AND GUIDANCE			
ANALYTE	MMCL/ORSG DRINKING WATER ^(A) . (µg/L)	CLASS I GROUNDWATER ⁽⁸⁾ (µg/L)		
Explosives	*			
cyclotetramethylenetetranitramine (HMX)				
cyclonite (RDX)				
2,4-dinitrotoluene		÷		
2,6-dinitrotoluene	A	4		
nitroglycerine	a la	R		
2,4,6-trinitrotoluene				
Cations/Anions				
chloride	250,00010	+		
phosphate		6		
sulfate	250,00010	250,000		
alkalinity				
Other				
nitrate/nitrite (total)	10,000	10,000*		
трн				

Notes:

- (A) MADEP - Office of Research and Standards; Massachusetts Drinking Water Standards and Guidelines, (310 CMR 22.00) Massachusetts MCLs; Autumn 1994.
- MADEP Division of Water Pollution Control; Massachusetts Surface Water Quality Standards, (314 CMR 6.06) Minimum GW Quality Criteria (B) - Class I; promulgated December 31, 1986.
- DWS Drinking Water Standards
- MCLG 3 Maximum Contaminant Level Goal
- MMCL Massachusetts Maximum Contaminant Level =
- Office of Research and Standards Guideline (Massachusetts) -
- $\mu g/L$ = micrograms per liter
- = Standard not established
- MMCL established for 1,4-dichlorobenzene isomer (more stringent than for 1,2- isomer). Reported values are totals (isomers not distinguished). Value is an Office of Research and Standards guideline. (1) (2) (3)
- Standard indicated is concentration of total trihalomethanes (i.e., the sum of concentrations of chloroform, bromodichloromethane, dibromochloromethane, and bromoform). Defers to USEPA DWS; see federal MCLs/MCLGs.
- (4) (5) (6) (7) (8) (10)
- Mean value per any set of samples.
- Numerical standard does not exist. MMCL is based on presence or absence of coliform.
- Nitrate as nitrogen.
- Value reported for endrin; CAS No. 72208.
- Value reported for PCBs; CAS No. 1336363.
- SMCLs = Secondary Maximum Contaminant Levels.

ANALYTE	SOIL ING	ESTION		TBC USEPA SQC ^c (mg/kg organic Carbon)
	TBC REGION III/RESIDENTIAL ^A (mg/kg)	TBC Region III/Commercial /Industrial [*] (mg/kg)	TBC NOAA SEDIMENT ⁸ Effects Range - Low (mg/kg)	
Volatile Organics				
acetone	7,800N	200,000N	-	4
benzene	22C	200C	-	÷
carbon tetrachloride	4.9C	44C	-	
chloroform	100C	940C	4	÷
ethylbenzene	7,800N	200,000N	×.	
styrene	16,000N	410,000N	-	
tetrachloroethylene	12C	110C		*
toluene	16,000N	410,000N		
1,1,1-trichloroethane	7,000N	180,000N		(#*)
1,1,2,2-tetrachloroethane	3.2C	29C	-	
trichloroethylene	58C	520C	-	÷
trichlorofluoromethane	23,000N	610,000N	-	÷.
xylenes (total)	160,000N	1E+06N	-	(*)
Semivolatile Organics				
acenaphthene		1.2		(†)
anthracene	23,000N	610,000N		-
bis(2-ethylhexyl)phthalate	46C	410C		-
benzo(a)anthracene	0.88C	7.8C		1,317

Remedial Investigation Report Devens, Massachusetts

Analyte	Soil Ing	ESTION	TBC NOAA SEDIMENT ^B EFFECTS RANGE - LOW (mg/kg)	TBC USEPA SQC ^c (mg/kg organic Carbon)
	TBC REGION III/RESIDENTIAL ^A (mg/kg)	TBC Region III/Commercial /Industrial [*] (mg/kg)		
benzo(a)pyrene	0.088C	0.78C	0.4	1,063
benzo(b)fluoranthene	0.88C	7.8C	7	
benzo(g,h,i)perylene				-
benzo(k)fluoranthene	8.8C	78C	÷	
benzyl alcohol	23,000N	610,000N	+	-
carbazole	_ 32C	290C		-
chrysene	88C	780C	0.4	•
dibenzofuran	310N	8200N	-	4
di-n-butyl phthalate	7,800N	200,000N		
fluoranthene	3,100N	82,000N	0.6	1,883
fluorene	3,100N	82,000N	0.035	
indeno(1,2,3-c,d)pyrene	0.88C	7.8C		
2-methylnaphthalene	-	-		0.065
naphthalene	3,100N	82,000N	0.34	
n-nitrosodiphenylamine	130C	1200C	-	
phenanthrene			0.225	139
pyrene	2,300N	61,000N	0.35	1,311
total PAHs	-	*	4.0	

.

	Soil Ing	ESTION		TBC USEPA SQC ^c (mg/kg organic Carbon)
ANALYTÉ	TBC REGION III/RESIDENTIAL ^A (mg/kg)	TBC Region III/Commercial /Industrial ^{&} (mg/kg)	TBC NOAA SEDIMENT ^b Effects Range - Low (mg/kg)	
Inorganics				
aluminum	78,000N	1E+06N	•	- (
antimony	31N	820N	0.002	*
arsenic	0.37C/23N	3.3C/610N	0.033	+
barium	5,500N	140,000N		
beryllium	0.15C	1.3C	*	+
cadmium	39N	1,000N	0.005	-
calcium	-	-		÷
chromium	390N ¹	10,000N ¹	0.080	
cobalt	4,700N	120,000N	- 4	
copper	2,900N	76,000N	0.070	
iron		-	-	÷
lead	•		0.035	+
magnesium				
manganese	390N	10,000N	+	*
mercury	23N	610N	0.00015	
nickel	1,600N	41,000N	0.030	-
potassium		-	÷	-
selenium	390N	10,000N		

Analyte	SOIL ING	ESTION	TBC NOAA SEDIMENT ^B EFFECTS RANGE - LOW (mg/kg)	TBC USEPA SQC ^c (mg/kg organic carbon)
	TBC REGION III/RESIDENTIAL ⁴ (mg/kg)	TBC Region III/Commercial /Industrial [*] (mg/kg)		
silver	390N	10,000N	0.001	
sodium		÷.		
vanadium	550N	14,000N		
zinc	23,000N	610,000N	0.120	-
Pesticides/PCBs				
DDT	1.9C	17C	0.001	0.828
DDD	2.7C	24C	0.002	+
DDE	1.9C	17C	0.002	***
endrin	23N	610N	0.00002	.0332
alpha chlordane	0.49C ³	4.4C ³	0.0005 ³	4
gamma chlordane	0.49C ³	4.4C ³	0,0005 ³	
heptachlor	0.14C	1.3C		0.110
PCB 1248	0.083C ⁴	0.74C ⁴	0.054	÷
PCB 1254	1.6N	41N	0.054	19.5
PCB 1260	0.083C ⁴	0.74C ⁴	0.054	
Explosives				
cycloetramethylenetetranitramine (HMX)	-	-		-
cyclonite (RDX)		4		1.5
2,6-dinitrotoluene	78N	2000N	4	+

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

ANALYTE	Soil Ing	ESTION		TBC USEPA SQC ^c (mg/kg organic carbon)
	TBC REGION III/RESIDENTIAL ^A (mg/kg)	TBC Region III/Commercial /Industrial [*] (mg/kg)	TBC NOAA SEDIMENT ^B Effects Range - Low (mg/kg)	
2,4,6-trinitrotoluene	21C	190C		
nitroglycerine				
Other				
nitrate/nitrite	130,000N/7,800N	1E+06N/200,000N	+	
ТРН	-	-		+

Notes:

(1) Chromium IV values.

(2) Dry weight.

(3) Values reported for chlordane (CAS # 57-74-9).

(4) Values reported for total polychlorinated biphenyls (CAS # 1336-36-3).

(a) US Environmental Protection Agency (USEPA) Region III, January 1995 Memorandum from Roy L Smith to RBC (Risk-Based Concentration) Table Mailing List, Subject: Risk-Based Concentration Table.

(b) National Oceanic and Atmospheric Administration (NOAA), March 1990 "The Potential for Biological Effects of Sediment-sorbed Contaminants Tested in the National States and Trends Program"; NOAA Technical Memorandum NOS OMA52 (Edward R Long and Lee G Morgan, authors).

(c) USEPA, May 1988 "Interim Sediment Quality Criteria Values for Nonpolar Hydrophobic Organic Contaminants".

С

(d) USEPA OSWER Directive 193554-02.

mg/kg = milligrams per kilogram

SQC = Sediment Quality Criteria

NOAA = National Oceanic and Atmospheric Administration N = Non-carcinogenic effects representing an HQ of 10 = No federal or state guidance criteria or standards exist
 = Carcinogenic effects representing an incremental lifetime cancer risk of 1 in a million

1E+06 = 1,000,000

TABLE 4-4 POTENTIAL LOCATION-SPECIFIC ARARS AT DEVENS AOC 69W

STATUTORY, REGULATORY BASIS	CITATION	DESCRIPTION
Resource Conservation and Recovery Act	40 CFR Sec. 264.18	Prohibits or restricts siting of hazardous waste management units in certain sensitive areas (100-year floodplain, active seismic area, wetlands).
Migratory Bird Treaty Act of 1972	16 USC Sec. 703-712 50 CFR Parts 10, 20, 21	If migratory birds are present, provides protection of almost all species of native birds in the U.S. from unregulated activi- ties. Unregulated activities can include poisoning at hazardous waste sites.
Endangered Species Act	16 USC 1531 50 CFR Parts 81, 225, 402	Provides for protection and conservation of various species of fish, wildlife, and plants.
Massachusetts Endan- gered Wildlife and Wild Plants	321 CMR 8.00	The Commonwealth of Massachusetts has authority to research, list, and protect any species deemed endangered, threatened, or of special concern. These species are listed as either endangered, threatened, or species of special concern in the regula- tions. The Massachusetts lists may differ from the federal lists of endangered spe- cies.
Clean Water Act, Section 404	33 USC 1251, Sec. 404; 40 CFR Part 230 33 CFR 320-330	Prohibits discharge of dredged or fill mate- rial into wetlands without a permit. Provides for management of dredged mate- rial; establishes requirements for structures affecting navigable waterways; and pro- vides for certain permitting requirements.
Fish and Wildlife Coor- dination Act	16 USC 661	All agencies regulating activities that may have an effect on either fish or wildlife must notify and allow input by agencies overseeing fish and wildlife habitats in the area of the proposed activities.

TABLE 4-4POTENTIAL LOCATION-SPECIFICARARS AT DEVENSAOC 69W

STATUTORY, REGULATORY BASIS	CITATION	DESCRIPTION
Protection of Wetlands- Executive Order (EO) 11990	40 CFR Part 6, Appendix A	Appendix A of 40 CFR 6 sets forth policy for carrying out provisions of the Protec- tion of Wetlands Executive Order. Under this order, federal agencies are required to minimize the degradation, loss, or destruc- tion of wetlands, and to preserve the natu- ral and beneficial values of wetlands. Appendix A requires that no remedial alternatives adversely affect a wetland if another practicable alternative is available. If no alternative is available, effects from implementing the chosen alternative must be mitigated.
Wetlands Protection	310 CMR 10.00	Establishes State of Massachusetts regu- lations for protection of coastal and inland wetlands, including compliance with the Massachusetts Environmental Policy Act.

Standard, Requirement, Criteria, or Limita- tion	Citation	Description	Comments/ Applicability
Solid Waste Disposal Act	42 USC 6901	Resource Conservation and Recovery Act.	
Criteria for Classification of Solid Waste Disposal Facilities and Practices	40 CFR Part 257	Establishes criteria for use in determining which solid waste disposal facilities and practices pose a reasonable probability of adverse effects on health or the environ- ment and thereby prohibit open dumps.	Not Applicable.
Criteria for Municipal Solid Waste Disposal Facilities	40 CFR Part 258	Establishes minimum federal criteria for design, construction, operation, and per- mitting of municipal solid waste landfills.	Not Applicable.
Identification and Listing of Hazardous Waste	40 CFR Part 261	Defines those solid wastes which are subject to regulation as hazardous waste.	Applicable for defining whether or not hazardous waste is generated or disposed.
Standards Applicable to Generators of Haz- ardous Waste	40 CFR Part 262	Establishes standards for generators of hazardous waste.	Relevant and appropriate. If remedial action causes hazardous waste to be generated, then these stan- dards will be applicable.
Standards Applicable to Transporters of Hazardous Waste	40 CFR Part 263	Establishes standards which apply to persons transporting hazardous waste within the U.S	Applicable. If hazardous waste is transported off- site, these standards will apply.

Standard, Requirement, Criteria, or Limita- tion	Citation	Description	Comments/ Applicability	
Standards for Owners and Operators of Hazardous Waste Treatment, Storage, and Disposal Facilities	40 CFR Part 264	Establishes minimum national standards which define the acceptable management of hazardous waste for owners and opera- tors of facilities which treat, store, or dispose of hazardous waste.	Relevant and appropriate for hazardous waste man- agement activities during remediation.	
Standards for Management of Specific Hazardous Wastes and Specific Types of Hazardous Waste Management Facilities	40 CFR Part 266	Establishes requirements which apply to recyclable materials used in a manner constituting disposal or hazardous waste burned for energy recovery.	If hazardous wastes are recycled on- or off-site, these standards are ap- plicable.	
Interim Standards for Owners and Operators of New Hazardous Waste Land Disposal Facilities	40 CFR Part 267	Establishes minimum national standards that define acceptable management of hazardous waste land disposal facilities.	Applicable if hazardous waste is present.	
Land Disposal Restrictions (LDR) Program	40 CFR Part 268	Sets treatment standards for hazardous wastes based on the levels achievable by current technology; sets two-year national variances from the statutory effective dates due to insufficient treatment capaci- ty.	Applicable for hazardous wastes destined for land disposal.	
Technical Standards and Corrective Action Requirements for Owners and Operators of Underground Storage Tanks (USTs)	40 CFR Part 280	Provides regulations pertaining to under- ground storage tanks.	Applicable if there is operation or removal of an UST.	

Standard, Requirement, Criteria, or Limita- tion	Citation	Description	Comments/ Applicability
Clean Water Act	33 USC 1251		
EPA-Administered Permit Programs: The National Pollutant Discharge Elimination System	40 CFR Part 122	Requirements for the discharge of pollutants from any point source into waters of the U.S.	Applicable if remedial action requires outfall discharge.
Criteria and Standards for the National Pollutant Discharge Elimination System	40 CFR Part 125	Provides discharge criteria, chemical standards, and permit forms for existing industrial operations.	Relevant and appropriate to remedial actions which cause discharge to waters of the U.S

Standard, Requirement, Criteria, or Limita- tion	Citation	Description	Comments/ Applicability
Clean Air Act	42 USC 7401		
Massachusetts Air Pollution Control Regulations	310 CMR 7.00	These regulations set emission limits necessary to attain ambient air quality standards.	Remedial activities will be conducted to meet the standards for visible emis- sions (310 CMR 7.06); dust, odor, construction, and demolition (310 CMR 7.09); noise (310 CMR 7.10); and volatile organic compounds (310 CMR 7.18).

TABLE 4-5 POTENTIAL ACTION-SPECIFIC FEDERAL ARARS AT DEVENS AOC 69W

RÉMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Standard, Requirement, Criteria, or Limita- tion	Citation	Description	Comments/ Applicability
Department of Transportation - Hazardous Materials Regulations			
Rules for the Transportation of Hazardous Materials	49 CFR Parts 107, 171, 172	Provides requirements for packaging, manifesting, and transportation of haz- ardous waste.	Applicable if off-site ship- ment of hazardous wastes occurs.

TABLE 5-1 SUMMARY OF INVESTIGATION ACTIVITIES AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

MONTH/ YEAR	ACTIVITY	CONTRACTOR	NUMBER COMPLETED	EXPLORATION/SAMPLE IDENTIFICATION	PURPOSE OF ACTIVITY
AREE 69 PAST	SPILL SITE INVESTIGATION PHASE I				
JULY 1993	DOCUMENT REVIEW AND SITE VISIT	ADL	NA	AREE 69W	DETERMINE STATUS OF 48 PAST SPILL SITES AND RECOMMEND RELEVANT ACTION OR NFA
AREE 69 PAST	SPILL SITE INVESTIGATION PHASE II (ARI	EE 69W)			
MARCH 1994	GEOPHYSICAL SURVEY	ADL	I SURVEY	AREE 69W	IDENTIFY POSSOBLE LOCATIONS OF UNDERGROUND STRUCTURES AND CLEAR BORING LOCATIONS FOR SUBSURFACE UTILITIES
MARCH 1994	SURFACE SOIL SAMPLING	ADL	6 SAMPLES	HA-1 THRU HA-6	COLLECT SURFICIAL SOIL SAMPLES FOR FIELD SCREENING. ONE SAMPLE COLLECTED FOR OFF-SITE CONFIRMATION
APRIL 1994	GEOPROBE TM SOIL AND GROUNDWATER SAMPLING (PHASE I)	ADL	16 BORINGS	GP-01 THRU GP-16	COLLECT SUBSURFACE SOIL AND GROUNDWATER SAMPLES FOR FIELD SCREENING. FOUR SOIL AND FIVE GROUNDWATER SAMPLES COLLECTED FOR OFF-SITE ANALYSIS
МАҮ 1994	GEOPROBE [™] SOIL AND GROUNDWATER SAMPLING (PHASE II)	ADL	9 BORINGS	GP-16 THRU GP-25	COLLECT SUBSURFACE SOIL AND GROUNDWATER SAMPLES FOR FIELD SCREENING.
JUNE 1994	MONITORING WELL BORING AND INSTALLATION	ADL	6 WELLS	69W94-09X THRU 69W94-14X	EVALUATE SUBSURFACE SOIL AND GROUNDWATER QUALITY AND CONFIRM TPHC SCREENING RESULTS
JUNE AND SEPT 1994	GROUNDWATER SAMPLING	ADL	6 WELLS	69W-94-09X THRU 69W-94-14X	MONITOR GROUNDWATER QUALITY (OFF-SITE ANALYSIS)
JUNE 1994	SURFACE WATER AND SEDIMENT SAMPLING (WILLOW BROOK)	ADL	2 SW/SED PAIRS	69W-94-15 AND 69W-94-16	EVALUATE IMPACT OF GROUNDWATER CONTAMINATION ON WILLOW BROOK (OFF-SITE ANALYSIS)
AOC 69W REM	EDIAL INVESTIGATION				
SEPT. 1995	GEOPHYSICAL SURVEY	ABB-ES	I SURVEY	PARKING LOT AND GRASSY AREA NORTH OF SCHOOL AND COURTYARD	PERFORM TERRAIN CONDUCTIVITY (EM-31), EM-61, AND GROUND PENETRATING RADAR SURVEYS TO IDENTIFY AND LOCATE POTENTIAL SUBSURFACE CONTAMINANT SOURCES AND UTILITIES.
	SEDIMENT SAMPLING AND TOXICITY TESTING	ABB-ES	9 SED AND 3 TOX TEST SAMPLES	ZWD-95-01X THRU ZWD-95-06X	COLLECT SAMPLES FOR OFF-SITE ANALYSIS AND TOXICITY TESTING TO ASSESS IMPACT TO ECOLOGICAL RECEPTORS IN WILLOW BROOK
	QUALITATIVE ECOLOGICAL SURVEY AND WETLANDS INVESTIGATION	ABB-ES	1 SURVEY	WILLOW BROOK AND ASSOCIATED WETLANDS	IDENTIFY POTENTIAL ECOLOGICAL RECEPTORS AND EXPOSURE PATHWAYS AT AOC 69W

TABLE 5-1 SUMMARY OF INVESTIGATION ACTIVITIES AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

MONTH/			NUMBER	EXPLORATION/SAMPLE	
YEAR	ACTIVITY	CONTRACTOR	COMPLETED	IDENTIFICATION	PURPOSE OF ACTIVITY
	TEST PITTING	ABB-ES	1 TEST PIT	ZWE-96-01X	INVESTIGATE IDENTIFIED GEOPHYSICAL ANAMOLIES
	TERRAPROBE SM BORINGS	ABB-ES	29 BORINGS	ZWR-95-26X THRU ZWR-95-28X, AND ZWR-95-30X THRU ZWR-95-55X	COLLECT FIELD AND OFF-SITE ANALYTICAL SOIL AND GROUNDWATER SAMPLES TO FURTHER DEFINE THE LATERAL AND VERTICAL CONTAMINANT DISTRIBUTION
	SOIL BORINGS	ABB-ES	2 SOIL BORINGS	ZWB-95-01X AND ZWB-95-02X	COLLECT OFF-SITE SOIL SAMPLES TO CONFIRM FIELD SCREENING RESULTS
	MONITORING WELL INSTALLATION	ABB-ES	4 WELLS	ZWM-95-15X THRU ZWM-95-18X	EVALUATE AND MONITOR GROUNDWATER QUALITY IN THE VICINTY OF AOC 69W
	PIEZOMETER INSTALLATION	ABB-ES	2 PIEZOS	ZWP-95-01X AND ZWP-95-02X	EVALUATE HYDROLOGIC AND HYDROGEOLOGIC CONDITIONS IN THE AREA OF WILLOW BROOK
OCT. AND NOV. 1995	GROUNDWATER SAMPLING ROUND 1	ABB-ES	11 WELLS	69W94-09X THRU 69W94-14X AND ZWM-95-14X THRU ZWM-95-18X	MONITOR GROUNDWATER QUALITY AT AOC 69W
NOV. 1995	IN-SITU HYDRAULIC CONDUCTIVITY TESTING	ABB-ES	11 WELLS	69W94-09X THRU 69W94-14X AND ZWM-95-14X THRU ZWM-95-18X	ESTIMATE HYDRAULIC CONDUCTIVITIES OF THE OVERBURDEN AQUIFER
FEB. 1996	GROUNDWATER SAMPLING ROUND 2	ABB-ES	11 WELLS	69W94-09X THRU 69W94-14X AND ZWM-95-14X THRU ZWM-95-18X	MONITOR GROUNDWATER QUALITY AT AOC 69W
AOC 57 REM	EDIAL INVESTIGATION MOD 1				the state of the second state of the second state
AUG. 1996	SOIL BORING	ABB-ES	1 BORING	ZWB-96-03X	COLLECT SUBSURFACE SOIL SAMPLES TO CONFIRM THE LIMITS OF CONTAMINATION
	MONITORING WELL INSTALLATION	ABB-ES	3 WELLS	ZWM-96-19X THRU ZWM-96-21X	EVALUATE SOIL AND GROUND WATER QUALITY WITHIN THE BOILER ROOMS AND COURTYARD
SEPT. AND OCT. 1996	GROUNDWATER SAMPLING	ABB-ES	6 WELLS	69W94-10, 69W94-11, ZWM-95-16X, AND ZWM-96-19X THRU ZWM-96-21X	MONITOR GROUNDWATER QUALITY AT AOC 69W SOURCE AREA
JAN. 1997	IN-SITU HYDRAULIC CONDUCTIVITY TESTING	ABB-ES	3 WELLS	ZWM-96-19X THRU ZWM-96-21X	ESTIMATE HYDRAULIC CONDUCTIVITIES OF THE OVERBURDEN AQUIFER
OCT, 1997	INDOOR AIR SAMPLING IN THE DEVENS ELEMENTARY SCHOOL	ABB-ES	12 AIR SAMPLES AND 1 WELL HEAD SAMPLE	ZWA-97-01X THRU ZWA-97-13X	AIR SAMPLING TO MEASURE THE LEVELS OF FUEL-RELATED CONTAMINANTS IN INDOOR AIR ASSOCIATED WITH SUBSURFACE AND GROUNDWATER CONTAMINATION

TABLE 5-1 SUMMARY OF INVESTIGATION ACTIVITIES AOC 69W

171

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

MONTH/ YEAR	ACTIVITY	CONTRACTOR	NUMBER COMPLETED	EXPLORATION/SAMPLE IDENTIFICATION	PURPOSE OF ACTIVITY
DEC. 1997	GROUNDWATER SAMPLING	ABB-ES	13 WELLS	69W94-09 THRU 69W94-14, ZWM-95- 15X THRU ZWM-95-18X, AND ZWM-96- 19X THRU ZWM-96-21X	An end of the second

GeoProbeTM and TerraProbeSM are proprietary direct push soil and groundwater sampling techniques.

TABLE 5-2 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

SAMPLE POINT	FIELD SCREENING SAMPLE ID	MEDIUM	DEPTH (feet below ground surface)	TPH (mg/kg)
GP-01	69W-94-01-2W	Groundwater	5-7	663
GP-02	69W-94-02-2W	Groundwater	4-6	354
GP-02(Dup)	69W-94-01-2WD	Groundwater	5-7	208
GP-03	69W-94-03-2W	Groundwater	5-7	110
GP-04	69W-94-04-2W	Groundwater	5-7	12
GP-05	69W-94-05-2W	Groundwater	5-7	126
GP-06	69W-94-06-2W	Groundwater	5-7	188
GP-07	69W-94-07-2W	Groundwater	4-6	3
GP-08	69W-94-08-2W	Groundwater	4-6	95
GP-08(Dup)	69W-94-08-2WD	Groundwater	4-6	75
GP-09	69W-94-09-2W	Groundwater	4-6	2
GP-10	69W-94-10-2W	Groundwater	4-6	25
GP-11	69W-94-11-2W	Groundwater	4-6	37
GP-11(Dup)	69W-94-11-2WD	Groundwater	4-7	2
GP-12	69W-94-12-2W	Groundwater	3-5	20
GP-13	69W-94-13-2W	Groundwater	3-5	460
GP-14	69W-94-14-2W	Groundwater	3-5	2
GP-15	69W-94-15-2W	Groundwater	3-5	438
GP-15(Dup)	69W-94-15-2WD	Groundwater	3-5	110
GP-16	69W-94-15-2WD		2-5	
		Groundwater	2-5	7 4
GP-16(Dup)	69W-94-16-2WD	Groundwater		
GP-17	69W-94-17-2W	Groundwater	5-7	1,015
GP-17(Filtered)	69W-94-17F	Groundwater	5-7	22
GP-18	69W-94-18-2W	Groundwater	5-7	19
GP-18(Filtered)	69W-94-18F	Groundwater	5-7	12
GP-19	69W-94-19-2W	Groundwater	5-7	15
GP-19(Filtered)	69W-94-19F	Groundwater	5-7	12
GP-20	69W-94-20-2W	Groundwater	5-7	650
GP-20(Filtered)	69W-94-20F	Groundwater	5-7	15
GP-21	69W-94-21-2W	Groundwater	5-7	119
GP-21(Filtered)	69W-94-21F	Groundwater	5-7	11
GP-22	69W-94-22-2W	Groundwater	5-7	1,087
GP-22(Filtered)	69W-94-22F	Groundwater	5-7	17
GP-23	69W-94-23-2W	Groundwater	5-7	45
GP-23(Filtered)	69W-94-23F	Groundwater	5-7	4
GP-24	69W-94-24-2W	Groundwater	5-7	10
GP-24(Filtered)	69W-94-24F	Groundwater	5-7	4
GP-25	69W-94-25-2W	Groundwater	5-7	15
GP-25(Filtered)	69W-94-25F	Groundwater	5-7	5
GP-1	69W-94-01-1	Soil	0-2	123
GP-1(Dup)	69W-94-01-1D	Soil	0-2	120
GP-1	69W-94-01-2	Soil	3-5	226
GP-2	69W-94-02-1	Soil	0-2	1,050
GP-2	69W-94-02-2	Soil	3-5	15,500
GP-3	69W-94-03-1	Soil	0-2	1,150
GP-3	69W-94-03-2	Soil	3-5	387

TABLE 5-2 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

SAMPLE	FIELD SCREENING SAMPLE ID	MEDIUM	DEPTH (feet below ground surface)	TPH (mg/kg)
GP-4	69W-94-04-1	Soil	0-2	1,230
GP-4	69W-94-04-2	Soil	3-5	1,010
GP-5	69W-94-05-1	Soil	0-2	201
GP-5	69W-94-05-2	Soil	3-5	102
GP-6	69W-94-06-1	Soil	0-2	806
GP-6	69W-94-06-2	Soil	3-5	53
GP-7	69W-94-07-1	Soil	0-2	38
GP-7(Dup)	69W-94-07-1D	Soil	0-2	38
GP-7	69W-94-07-2	Soil	3-5	11
GP-8	69W-94-08-1	Soil	0-2	291
GP-8(Dup)	69W-94-08-1D	Soil	0-2	272
GP-8	69W-94-08-2	Soil	3-5	48
GP-9	69W-94-09-1	Soil	0-2	219
GP-9	69W-94-09-2	Soil	3-5	11
GP-10	69W-94-10-1	Soil	0-2	19
GP-10	69W-94-10-2	Soil	3-5	51
GP-11	69W-94-11-1	Soil	0-2	20
GP-11	69W-94-11-2	Soil	3-5	74
GP-12	69W-94-12-1	Soil	0-2	40
GP-12	69W-94-12-2	Soil	3-5	8
GP-13	69W-94-13-1	Soil	0-2	25
GP-13	69W-94-13-2	Soil	3-5	157
GP-14	69W-94-14-1	Soil	0-2	51
GP-14	69W-94-14-2	Soil	3-5	8
GP-15	69W-94-15-1	Soil	0-2	95
GP-15	69W-94-15-2	Soil	3-5	13
GP-16	69W-94-16-1	Soil	0-2	607
GP-16	69W-94-16-2	Soil	3-5	18
GP-17	69W-94-17-1	Soil	3-5	51
GP-17(Dup)	69W-94-17-2D	Soil	3-5	53
GP-18	69W-94-18-2	Soil	3-5	22
GP-19	69W-94-19-2	Soil	3-5	12
GP-20	69W-94-20-2	Soil	3-5	420
GP-21	69W-94-21-2	Soil	3-5	1,497
GP-22	69W-94-22-2	Soil	3-5	14,484
GP-23	69W-94-23-2	Soil	3-5	11,404
GP-24	69W-94-24-2	Soil	3-5	1,520
GP-25	69W-94-25-2	Soil	3-5	21
59W-94-09	69W-94-09-1	Soil (Boring)	0-2	70
59W-94-09(Dup)	69W-94-09-1D	Soil (Boring)	0-2	70
59W-94-09(Dup)	69W-94-09-1D	Soil (Boring)	2-4	14
59W-94-09	69W-94-09-3	Soil (Boring)	4-6	14
59W-94-09	69W-94-09-4	Soil (Boring)	11-13	11
59W-94-10	69W-94-10-1	Soil (Boring)	0-2	664
69W - 94 - 10	69W-94-10-1	Soil (Boring)	2-4	631
69W-94-10	69W-94-10-2 69W-94-10-3	Soil (Boring)	4-6	7,675

TABLE 5-2 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

SAMPLE POINT	FIELD SCREENING SAMPLE ID	MEDIUM	DEPTH (feet below ground surface)	TPH (mg/kg)
69W-94-10	69W-94-10-4	Soil (Boring)	11-13	164
69W-94-11	69W-94-11-1	Soil (Boring)	0-2	1,735
69W-94-11	69W-94-11-2	Soil (Boring)	2-4	386
69W-94-11	69W-94-11-3	Soil (Boring)	4-6	125
69W-94-11	69W-94-11-4	Soil (Boring)	11-13	55
69W-94-12	69W-94-12-1	Soil (Boring)	0-2	11
69W-94-12	69W-94-12-2	Soil (Boring)	2-4	30
69W-94-12	69W-94-12-3	Soil (Boring)	4-6	667
69W-94-12	69W-94-12-4	Soil (Boring)	11-13	40
69W-94-13	69W-94-13-1	Soil (Boring)	0-2	1,140
69W-94-13	69W-94-13-2	Soil (Boring)	2-4	19
69W-94-13	69W-94-13-3	Soil (Boring)	4-6	2,900
69W-94-13	69W-94-13-4	Soil (Boring)	11-13	71
69W-94-14	69W-94-14-1	Soil (Boring)	0-2	25
69W-94-14	69W-94-14-2	Soil (Boring)	2-4	83
69W-94-14	69W-94-14-3	Soil (Boring)	4-6	1,990
69W-94-14	69W-94-14-4	Soil (Boring)	11-13	49
HA-1	69W-94-01	Soil	0-1	14
HA-1(Dup)	69W-94-01D	Soil	0-1	53
HA-2	69W-94-02	Soil	0-1	16
HA-3	69W-94-03	Soil	0-1	10
HA-4	69W-94-04	Soil	0-1	56
HA-5	69W-94-05	Soil	0-1	131
HA-6	69W-94-06	Soil	0-1	36

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes:

TPH = Total petroleum hydrocarbons mg/kg = milligrams per kilogram Dup = Duplicate sample

TABLE 5-3 ADL, AREE 69W SOIL OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Sile Er:		63W-\$4-\$1			69W-94-0			69W-94-03	1		¥94.83		69W-94-04			W-9445	E.	69W-94-69		6078-34.69		6997-94-14		6798-64-18
Field Sample Neurbert		5XC9W-10			BDC9W-34	k		8XC998.2L	· 1		C998-347		BXC#W-AJ			CC994-50		BXD9W-9U		BKD9W AL		EX09W-BU		EXD#W-CL
P		HA S			GP 2			GP 1			CP-S		CP-4			GP.33		8944-94-09		6911:94.89		69W 54-16		699V-94 16
Surple Dete: Depth:		91/30/94			05/02/94			8502/4		19 19	102/94		05/82/94			18/1/2/94		86/06/Ve		96/68/94		06/96/94		060641
Unite								PET			en ka		¥0/2			148/E		10/5		54/k			I	
L METALS																ot , o occosocoo		196.2				5612	opposite 4	
luminium		9690		1	7310	D	1	10600	1		7590	1	7710		-	10400	1	7260	1	5870		\$120		8440
mesút		12.4			11.5	D		13.7			12.6		11.6			13.8		9.08		11.4		10.2		25.9
arium		17.4		1.1	19.6	D		21.1			19.8		19.7			15.1		20.9	11.1	13.2		21.5		13.1
oren	(10		<	6.64	D	1	6.64		× .	6.64	1	6.64		<	6.64	<	6.64	<	6.64		13.6	- 1	11.8
alcium		680			1060	D		1060		11	1510		974			638	100	894	110	792		894		1570
in consisters		25.2			15.2	D		22.7			14.5		15.6			16.1		16.1		10.6		13.7		24.6
balt		5.49			5,24	D		5,98			4.73		4.23		<	2.5		5.47	1	3.04	I	3.8	- a	4.16
opper		13.2			9.1	D		9.07			7.8		7.81			8.94		10.8	511.	7.78		7.86	114	6.28
00		15300			13200	D		23300		1	4100		12700			15700		11300		8400		11300		14200
aid		56			12	D		13.4			9.41		8.78			12.1		2.95		2.86		7,5		6,68
lagrenium		3940			2400	D		4040			2860		2420			2850		2940		1530		1910		3150
languorse		194			169	D		248			186		173			143		160		109		197		181
ickel		15.1			15.3	D	1	20.2	1		15.8	T	12.6	T.		15.6	1	11.9		8.41		10.4		14.5
otamium		948			903	D		868			950	101	803			839	0.1	1600	1	581		719		745
nuibo		98.2			154	D		170			183		214			106		285		255		139	1	252
haffium	<	34.3		<	34.3	D	5	34.3		<	34.3	<	34.3		<	34.3	5	34.3	<	34.3		< 34.3		45.9
in a	<	7.43	7	<	7.43	D	8	7.43			11		7.43		<	7.43	<	7.43	<	7.43		< 7.43		15.2
audium	10.00	20.7		100	12.2	D		18.9			12.6	2	11.7			16.4	1.6	12.9		6.35		11.6		13.4
line		32.9			26.1	D		30,4			23.8		22.6			23.6		28.9		14.7		27.3		20.3
EMIVOLATILE ORGANICS	1	_		24.2																				
-methylniplithalene	<	0.032		115 1	0.1	D	<	0.032			0.032	5			<	0.032	<	0.032	<	0.032		8.4		70
comphilianc	<	0.041		<	0.041	D	5	0.041			0.041	<	ara tr		5	0.041	<	0.041	<	0.041		0,56		3
exaphilitylene	<	0,033		<	0.033	D	×	0.033			0.033		0.13		<	0.033	<	0.033	<	0.033		< 0.033		< 0.3
anto a starscore	1.2	0.13		1.0	0.12	D		0.58		<	0.041		0.81		<	0.041	<	0.041	<	0.041		0.47		< 0.4
ento[b]fluoranhene	<	0.31		<	0.31	D		0.75			0.31		1.1		<	0.31	<	0.31	<	0.31		1		< 1
enno[ghi]perylene	<	0.18		<	0.18	D		0.51			0.18		0.69		<	0.18	<	0.18	<	0.18		0.44		< 2
enzo[k]fluorantbene	<	0.13		<	0.13	D	1.00	0.34			0.13		0.43		<	E1.0	<	0.13	<	0.13		0.36		< 1
erzyl Alcohol	<	0,032			0.042	D	4	0.032			0.032	5			<	0.032	10	0,07		0.077		< 0.032		< 0.1
teynese		0.16			0.14	D	11.5	0.53			0.032		0.72		<	0.032	<	0.032	<	0.032		0.76		< 0.3
licity! Phinalste	<	0.24		1	0.24	D	<	0.24			0.24	<			<	0.24	<	0,24	<	0.24		< 0.24		< 2
luorauthene	1.	0.2		1.2	0.12	D	1.5	0.9			0.032		0.7		<	0.032	<	0.032	<	0.032		0.61		< 03
luorene	<	0.065		<	0.065	D	<	0.065			0.065	<			<	0.065	<	0.065	<	0.065		1		9
lephibalme	<	0.74		<	0.74	D	<	0.74			0.74	<			<	0.74	<	0.74	<	0.74		3.2		30
lema the one		0.42		1.5	0.15	D	15	0.95			0.032	1115	0.15		<	0.032	*	0.032	<	0.032		Z		20
heno)	<	0.052		5	0.052	D	<	0.052			0.052	<	0.052		\$	0.052	5	0.052	<	0.052		< 0.052		< 0.5
yrax	150	0.25			0.25	D	1.0	1.1			0.083		1		<	0.083	~	0.983	<	0.083		0.91		< 0.8
i-o-butyl Phthalate	<	13		<	13	D	<	1.3		κ.	1.5	<	.1.3		<	1.3		35		1.8		1,6	-	< 10
DLATTLE ORGANICS	1.			-			1.5		-	-	-			_			-							
lorobarzane	<	0.1		1	0.1	D	¢	1		<.	1	<			<	4	<	1	<	4		< 1		< .1
hloroform	<	0.24		<	0,24	D	<			¢.	.24	<			<	.24	<	.24	<	.24		< .24		< .24
hylberrene	<	0.19		<	0.19	D	<	.19		<	.19	<	19	-	<	.19	<	,19	<	.19		< .19		1,1
oluene	<	0.1		15	0.1	D	15	1		×	1	<	- 4	-	¢	4	<	1	<	.1		5 4	_	< 1
THER	-		-	T		-	T		-				1612			74777		2.51	- 1 -					
otal Organic Carbon		14600			7600	D		10800			2930		4310			15100		1350	<			7800		9750
Total Petroleum Hydrocarbona		176	_		2760	D	-	2780			1530	-	1070	-		325	_	103		13.6		425		4350

TABLE 5-3 ADL, AREE 69W SOIL OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT
DEVENS, MASSACHUSETTS

Mie III Piele Danose Mander Field III Sample Date:		900-94-11 (3D9%-04) (30%-94-11 (30-04-94	Γ	2999-94-11 BDD99-21 0399-94-11 66/05/94			6999-94-11 BXD#W-81. 6999-94-11 D6/85/94			6997-94-12 8X0997-91 9977-94-12 96/92/94			59W-54-12 8XD9W-GL 59W-56-13 969254			87W-54-13 IXU9W-8413 69W-94-13 8691/94		6997-94-13 200999-11 6997-94-13 86/81/94		6797-84-14 530999-31 6997-84-15 56-92-94			908-04-14 32940-321 830-94-14 86/92/94	
Depite		2				1	•			2			· •	- 1		- <u>-</u>		•					•	
PALMETALS	1200000	147	01:000	ap's			191/E		1000000				2407			26.6		94/t					NU 8	
Ahminun	1	8200	T	8220	D	T	7550		1	8770		-	7390	T	-	8790	1	8810	1	11500	- 1	_	9810	_
Artant		14.4		11.1	D		15.8			9,19		1000	28.6			8.21		22.1		14.2			15.1	
Barium		16		17.2	D	н.	15.2			14.8			19			19.6		16.5	1	18.8			17.2	
Baron		9.78		9.65	D					9.44			n		<	6.64		9.03		9.86			15.7	
Calcium		1036		876	Ð		971			430			981		Š	1370		1030		518			961	
Ciromian		16.1		15.5	D	1	11.7			13.5			19.6			14.4		32.6		23.8			21.9	
Cobali		4.85 1		3.85	o .		4.79	1		3.08	1.1		4.8	1	<	2.5	1.1	4.03	1	10.2	7		5.31	
Copper		22.4		6.95	D	11	8.81			6.74	- 11		4.9	- 1	61	6.01	^ L	10.4	211	16.8			10.1	
iron	1.1	17500		14600	D		13500			11200	- 1		13100			11200		14500		28500			16200	
Lest		7.37		4.03	D		5.92			8.35	1		6.1			9.7		2.75		4.97			4.35	
Magnerium		3400		3070	D		2510			2250			3370			2160		4140		4530	1.01		4480	
Management		166		143	D		147			158			174			127		127		640			187	
Nickel		15	1.	13.7	D		14.2			10.3			13.4			10.6		16.4		44.4			17.9	
Poternison		605	4.1	690	Ð		603			402			1280			414	- 1-	856		427			991	
Sodium		234		224	D	1	218			125			194			196		246		137			223	
Thalliam	8	34.3		46.8	D		44.7		<	34.3			40.8	- 11	<	34.3		58.6		52.7		<	34.3	
Tin	2	7.43	<		D	13			<	7.43		*	7.43		0	7.43		10.7	<			<	7.43	
Vacadian	1	13.1	100	11,1	D	1111	9.13		100	12.4		12.0	13.8			11.5		15	1	14.5			15.9	
Zinc		32.6		23.7	D		21.6			22.9			26.7	-		21.1		20.9		44	-		28.2	_
SEMEVOLATILE ORGANICS								-			_			-	_		-							
2-methylnaphthalene	<	0.032		0.67	D		0.32	-	<	0.032		<	0.032		<	0.032		7.2	<	0.032		<	0.032	
Accupithese	<	0.041	<		D	1	0.041		<	0.041		<	0.041		\$	0.041			<	0.041		<	0.041	
Accomplititylese	<	0.033	1 <		D	1	: 0.033		<	0.033		<	0.033			0.17		0.033	<	0.033		<	0.033	
Benzo(a)enthracons	<	0.041	<		D	1	0.041		1.2	0.12		<	0.041			0.64		0.15	1.1	0.092		<	0.041	
Benzo(b)fluorantiasse	5	0.31	<		D	14			<	0.31		<	0.31	- 1		0.98	1		<	0.31	1	<	0.31	
Bound ghi)paytene	<	0.18	1		D				<	0.18		<	0.18			0.58	1		<	0.18		<	0.18	
Benzo(k)fluoranthese	<	0.13	<	0.13	D	1			<	0.13		*	0.13			0.37	1		<			<	0.13	
Beruy! Alcobol	<	0.032	115	0.06	D	11	0.08		<	0.032		<	0.032		¢	0.032	-	100000	<	7.24		<	0,032	
Chaysone	10.0	0.071	<	and the second	D		0.067		1.0	0.11		<	0.032		1.3	0.63		0.15	110	0.13		٢.	0.032	
Diethyl Phihalate	<	0.24	15	0.24	D	1			<	0.24		5	0.24		<	0.24		3.3	<	0.24		5	0.24	
Floorinthese	100	0.048	<		D	13			1.5	0.13		<	0.032			0.26	1	a second s		0.11		<	0.032	
Fluorane	<	0.065		0.15	D	1			<	0.065		<	0.065		<	0.065		0.74	<			<	0.065	
Naphähalone	¢	0.74	1 <		D	-3	444.4		<	0.74		<	0.74		<	0.74		2.4	<			<	0.74	
Plenanthrow		0.078		0.3	D		0.23		1.	0.12		5	0.032		5	0.45		1.4	1.1	0,28		<	0.032	
Phenel	4	0.052	1	0.000	D	1			<	0.052		1	0.17		<	0.052	1		<			<	0.052	
Pyrene	<	0.083	<		D	1			1.	0.19		< <	0,083	1	<	0.4				0.22			0.16	
Di-n-buryi Pfahalate	1	20	-	21	D.	_	2.7		<	11		¢	13	-	5	1,3		4 13	1<	13		<	1.3	_
VOLATILE ORGANICS	12	-	12		D	1.2		-	14		-	12		- 1	<	-	- 1	10	-		- 1	<		
Chlorobenzene Chloroform	v v	.1.	K		D	X			~	.1 24		<	.1		<	.1 24	110	.18	<	33		2	.1	
Dividence	4	19	1		D		1997.1		4	19		<	.4		~	19		23	<			<	19	
Toluene	1	19	12		D		19		5	19		i i	19		~	1	14						11	
OTHER	1.	- Me	1		0	13	-		12		-	12	-94	-	-		- 13		-1-	4				
Total Organic Carbon	T	6850	1	2020	D	T	2450	_	1	11600		1	3040	- 1	1	19000	-	4110	1	5430	1	<	1000	
Total Petroleum Heirocarbors		865		176	D		101			23.4			1390			34.1		2270		123		19	2430	

Note: <= The concentration was less than the certified reported limit. $\mu g/g = micrograms per gram.$ D = Dtriplicate analysis. I = The low späce recovery for this lot was high.

04/23/24

TABLE 5-4 ADL, AREE 69W GROUNDWATER OFF-SITE ANALYTICAL RESULTS AOC 69W

۰.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Statu Flaid Sample Standart Flaid Di Sample Data Data	49%-44.92 MCC9%-22 GP-1 959234 991	97W-54485 96XC9W-5X 678-33 8572.94 9572.94	677 34.94 507299 62 67-1 807294 82	89W-34.67 MXC9W-7X CP-15 859274 up1.	60%-54.63 MXC9%-5X GP-4 659254 9965	69W Sales 6002499 8X 870, 5449 951894 4875	6777-54.09 MCD677-94 6777-94-09 05/24/94 98/7-	40W-34.09 MC23W-92 49W-44.59 09/19/94 39/1	59W-84-10 51509W-8X 69W-84-18 96/24/94 sell	5797-34 15 4672578-43 5797-94-10 972494 9824-
PALMETALS	NAME AND ADDRESS OF AD		1	1						
Aluminum	140000	45700	27600	150000	105000	1740 D	13400	7590	85200	4200
Antimony	116	< 60	< 60	132	< 60	< 60 D	< 60	< 60	< 60	< 60
Amonic	< 2.35	< 2.35	320	37.8	170	14.5 D	19.5	29.5	470	350
Baritan	605	122	143	374	268	40.5 D	76.2	57.7	347	53.3
Baylinen	7.22	1.6	< 1.12	5.34	5.09	< 1.12 D	< 1.12	1.39	4.41	< 1.12
Cadmian	< 6.78	< 6,78	< 6.78	< 6.78	< 6.78	< 6.79 D	< 6.78	< 6.78	11.4	< 6.78
Celcium	99700	24200	63700	41000	21100	19400 D	20000	19900	34700	21100
Chromium	347	131	83.8	513	312	< 16.8 D	17	< 16.8	192	17.6
Cobalt	114	28.3	28.7	86.5	82.3	< 25 D	< 25	< 25	83.3	< 25
Copper	207	63.9	47.1	230	211	< 18.8 D	< 18.8	< 18.8	165	< 18.8
tron	225000	58400	137000	177000	172000	869 D	11700	7630	148000	
Lead	201	10000								35100
		12.8	52.1	71.5	152		7.18	5.59	102	6.06
Magnesium	38900	12800	11000	48500	35500	1920 D	3990	3120	24900	3410
Mangpriese	6460	2850	9880	3630	3550	158 D	265	248	13000	3950
Mercury	.219	< ,1	4 1	0.295	< 0.1	< 0.1 D	< 0.1	< 0.1	0.112	< 1
Nickel	318	93.6	67.7	333	278	< 32.1 DH	< 32.1	< 32.1 H	234	< 32.1 H
Polamituri	24700	5260	9710	20300	14800	1750 D	5460	3540	14700	5240
Scientism	< 2.53	< 2.53	< 2.53	< 2.33	< 2.53	< 2.53 D	< 2.53	< 2.53	4.05	< 2.53
Sotium	280000	31800	40700	45600	45400	26900 D	28500	27600	32700	23700
Variadium	195	43.9	44.1	191	137	< 27.6 D	< 27.6	< 27.6	129	< 27.6
Zinc	483	1 132	1 129	432	1 344	1 5 11 0	34	22.6	245	20
SEMIVOLATILE ORGANICS	They are a second secon	1		11 14	11	113 11 14	1		1	
2-methylemphthalene	1000	< 13	500	× 13	400	< 13 D	< 13	< 13	1 200	1 244
		\$ 58	< 60	< 5.8	< 60	< 1.3 D < 5.6 D	< 1.3 < 5.8		700	400
Acmaphilione								4 5.8	34	< 30
Fluorene	< 90	< 92	< 90	< 9.2	< 90	< 9.2 D	< 9.2	\$ 9.2	35	< 50
Naphthalene	600	11	300	< .3	500	< .5 D	< .5	< .5	500	300
Phenanthrene	< 100	< 9.9	< 100	× 9.9	× 100	< 9.9 D	< 9.9	< 9.9	61	< 30
VOLATILE ORGANICS										
1.1.1-trichloroctione	< 1	30	20	< 1	< I	< 1 D	2.6 8	< 1	e 1	< 1
1.1-dichloroethane	< 1	< 1	10	< 1	< 1	< 1 D	< 1	× 1	s 1	< 1
1,1-dehloroethene	< 1	< 1	20	< 1	< 1	< 1 D	< 1.	< 1	< 1	< 1
Acidoos	< 8	< 8	< 80	37	< 8	< 8 D	< 8	< 8	< 8	< .
Benzene	< I	< 1	9	< 1	< 1	< 1 D	< 1	< 1	1 × 1	< 1
Carbon Tetrachluride	* 1	1 1	10	1 1	1 I	< 1 D	< 1	10 I	12 1	14 1
Chlorobenzene	< 1	2	10	1		I D	2 1		12 1	1
Chloreform	5 1	< 1		1		i D		12 1		
			10		2 12		× 1	13		
Chiloromethane	¢ 1.2	1 C C C C C C C C C C C C C C C C C C C	20	< 1.2	1.4	< 1.2 D	< 1,2	< 12	< 1.2	< 12
Ediyibeane	80	< 1	40	< 1	< 1	5 1 D	< 1	× 1	52	57
Tetrachloroethese	< 1	< 1	10	< 1	< 1	< 1 D	< 1	× 1	< I	4 1
Toluce	< I	< 1	10	< 1	< 1	< 1 D	< 1	× 1	4 1	< 1
Trichloroethylene	s 1	< 1	10	× 1	× 1	× t D	3,2	< 1	< 1	< 1
Trichlorofluoromethane	< 1	< 1	20	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Xylenca	< 20	< 1	× 20	< 2	× 2	< 2 D	× 1	× 1	8	2.4
WET CHEMISTRY										
Alkalimity							21000		97400	
Chloride						34000	36000	34000	30000	22000
Nibrite, Nibrate-non Specific	14.6	22.9	20	3100	22.3	2100 D	3800	1000	210	310
Nitrogen By Kjeldahl Method	3200	764	5000	1260	2200	5 64 D N	88	< 64 ·	N 810	1180 1
	1800		457							
Phosphorus	1000	1900	437	5200	1400	51.2 D		174	1300	330
Sulfate						19000	23000	20000	4000	3660
Total Hardness						55400 D	64900	61200	183000	65300
Total Suspended Solids			-	1		344000 D	290000	348000	1620000	174000
THER										

TABLE 54 ADL, AREE 69W GROUNDWATER OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Flaid Stample Number Phot ID: Sample Inder Lietter		89W-94-11 (D)99W-8X 69W-94-11 (0\$23.94 (\$23.94 (\$25.94			6796-34-33 6996-94-33 56/23/94 Jg/L		59W-9411 MXE4W EX 69W-9411 99/26/34 #61			6FW-94-12 MXD9W-CT 8797-94-12 9603994 gg/L		49W-84-12 MEEDW-CX 69W-94-52 05/19/94 25/1-			69W-94-13 SD(D9W-D)X 69W-94-13 96/25/94 pg/1		8799: 54:13 5022599 892 8899:54:13 595:19794 8897:			6930-54-14 MCD48W-EX 8770-54-14 6623-94 9821.		8997.94.14 6997.94.44 8997.94.44 8971.94
ALMETALS	_													-				_	-			
honman	3	2790	0	1.00	2610	1.1	447		1	11300		7290		1	537	1.5	4610			18100	1.0	8330
ntimony	<	60	D	× .	60	<	60		<	60	<	60		<	60	<	60		1 <	60	<	60
restic		38	D		37.8		62.1			25.6		36		1	60.4		400			70.4		270
larium		34.4	D	1	32		22.1		1.0	56.5		54.4			25.6		42.9			76		65.3
leryllium	<	1.12	D	<	1.12	<	1.12		<	1.12	1.1	1.96		5	1.12	1.1	1.4		<	1.12	1 <	1.12
edmium	<	6.78	D	5	6.78	<	6,78		<	6.78	1.6	6.78		1	6.78	<	6.78		<	6.78	<	6.7B
ລໄດ້ເພກ		27900	D		25200		26000			24400		18700		10.1	19200	100	20200			18900		18500
bromium	<	16.8	D	1	16.8	3	16.8		1.0	22.2		20.5		<	16.8	1 <	16.8			39.5	<	16.8
obalt	<	25	D	<	25	<	25		< 1	25	<	25		<	25	<	25			28.8	<	25
opper	<	18.8	D	1	18.8	<	18.8		<	18.8	<	18.8		<	18.8	14	18.8			27.8	<	18.9
UNI		5430	D	1	5150		2470			12100		9160		1.1	4730		28400			25200		23600
ad		4.47	D	1	4.47	<	4.47			8.92		9.65		1 <	4.47		11			23.3		15.2
Gignetium	201	3630	ñ	1.2	3440		2920		1	4680		3740		1	2040		3770			5280		2850
Annamase		3340	Ď		3140		2330			395		409			3910		2920			1850		994
lenairy	4		D	4		1	1. The second		1	395		409		<		14				.108	<	1.44.4
	2	1.	0	~	1	1	1	н	12	32.1	i i		н	12	4	12	1	н			2	.1
lickel		32,1	-		32,1		321	n	· ·			32.1		1	32.1	1	32.1			49.8	14	32.1
otasnum		3270	D	1.1	3110	~	1240		1.0	4250		4360			2430	1.2	2780		1.	3460		2230
elenium	<	2.53	D	<	2.53	*	2.53		<	2.53	<	2.53		4	2.53	<	2.53		<	2.53	1	2.53
odium		37500	D	-	33500	1.0	34900		1.00	47300		34300		1.0	16700	1.0	28900		1.0	28000	1.1	38300
anadium	<	27.6	0	<	27.6	1 <	27.6		<	27.6	<	27,6		<	27.6	<	27.6		<	27.6	1	27.6
inc	1	32.5	0	1	45.5		29.1		-	38.5		29.1	_		19	_	22.6			50.8		2.0
EMIVOLATILE ORGANICS	_		_						_					_								
methylmaphthalene	-	37	0		42		7.2		<	1.3	<	1.3		1	69	1.1	29		<	6	14	1.3
comphthese	<	5,8	D	<	5.4	<	5.8		*	5,8	<	3.8		<	5.6	<	5.8		<	30	<	5.6
huorene	•	9.2	D	<	9.2	<	9.2		<	9.2	<	9.2		<	9.2	1	9.2		<	50	1	9.2
laphthalene		8.8	D	1000	11	1.1	27		<	5	<	.5		12.0	64		47		<	2	<	5
boundhrox	4	9.9	D	1	9.9	<	0.0		e .	9.9	1 <	9.9		1	9.9	<	9.9		×	50	<	9.9
OLATILE ORGANICS	-		_																			
1.1-trichloroethane	Law -	1.5	DB	<	1	<	1	-		3.1 B		1,7			4.1 8	<	4		< N	1	<	1
1-dichloroothane		1	D	<	1	<	1		1	1	<	1		1	1	1	1		<	1	<	i.
1-dichloroethese	6	1	D	<	i i	<	T.		c .	1	<	1		×	i.	<	1		<	1	1 4	1
celose	*	R	0	<	8	<	8		1	8	<	8		4	8	1	8		1		14	8
environe		i.	D	<	1	<	1		e.	1	5	- D		10	1	<	1		<	î.	1	1
arbon Tetrachloride		ř.	D	1	1.5	1	1.1			3	1	÷.		2	1	1			1	÷.	14	1.1
hloroberume	2		0	12		1			2		1			1		1			1		12	
Moroform	2	î	õ	2	1.0	2	-		10		12	÷		2		12	1		1		1.5	
hloromethane	2	1.2	D	2	1.2	12	No.		2	1.2	12	1.2		1.	1.2	1×	12		2	1.2	12	1.2
hioremanane Itylbenzene	-	1.00	D	2	A.C.	12	ine .		2	1	1×			1	19		9.1		12		12	1.00
	100					1.2			2						19		×.1					
strachloroethene	•	1	D	<		<			6	1	<	1		5	1	1	-		1		1	1
oluone	*	1	D	5	1	<	T.		15	1	<	1		1.5	. E.	14			1	1	5	5
netheroadly lene	*	1	D	<	r	<	1			5.9	<	1		× .	1	<	3		1	1	1	T
idulored uoramethave	*	1	D	<	1	<	- P		1 4	1	<			4	4	<	- 1		<	1	<	
vices	4	2	D	<	2	1 <	2		1	2	1.4	2		5	2	<	2		1 <	2		2
ET CHEMISTRY	-			-				_	-					-				-				
kalinity		52500	0	1	56000	-				21200					43300		1000			31500		(Constru
Noride		58000	D		58000		48000			81000		41000			35000		40000		1	40000		52000
inite Nitrate-non Specific		500	D		350		1190			3100		2000			20.7		162			182		1600
itrogen By Kjeldahl Method		627	D	1	681		145	N		330		203	N		635		1100	N		423		456
aurodgeod		77	D		73.9		45.2			225		400			30.9		560			620		740
uifee		17000	D		17000		13000			18000		19000			2520		1850			17000		15000
otal Hardness		83000	D		78000		75400			78600		60600			55300		64600			67200		56700
COLUMN FOR CONCERNMENT OF COLUMN			-														2190000		1			545000
Kal Hardness Kal Suspended Solids		47000	0		36000		62000			397000		815000			10000					540000		

Notes: \leq = The concentration was less than the certified reported limit. $\mu_0 f$. = microgram per liter. D = Duplicate analysis. I = The low spike recovery for this lot was high. H = Lot out of control but data accepted due to high recoveries. V = Sample was not correctly preserved. N = The high spike recovery for this lot was low. B = Analyte found in method blank or QC sample.

TABLE 5-5 ADL, AREE 69W SURFACE WATER OFF-SITE ANALYTICAL RESULTS AOC 69W

Site ID: Field Sample Number: Sample Number: Units:	69W94-15 WXD9W-FX 06/24/94 µg/L	69W-94-16 WXD9W-GX 06/24/94 μg/L
PAL METALS	12	
Aluminum	658	< 112
Arsenic	2.92	< 2.35
Barium	17.5	13.1
Calcium	26600	23900
Iron	1470	276
Lead	7.65	< 4.47
Magnesium	4480	3750
Manganese	164	38.8
Potassium	1920	1920
Sodium	42700	40000
Zinc	24.6	< 18
VOLATILE ORGANICS		
1,1,1-trichloroethane	3.7 B	2.2 B
WET CHEMISTRY		
Alkalinity	< 10000	42300
Chloride	77000	75000
Nitrite, Nitrate-non Specific	930	1100
Nitrogen By Kjeldahl Method	353	214
Phosphorus	91.2	24.1
Sulfate	14000	15000
Total Hardness	83000	73700

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes:

 $\mu g/L = micrograms per liter$

< + Concentration is less than the certified reporting limit.

B = Analyte found in method blank or QC sample.

TABLE 5-6 ADL, AREE 69W SEDIMENT OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth: Units;	69W-94-15 DXD9W-FO 06/24/94 0		69W-94-16 DDD9W-GO 06/24/94 0	******		69W94-16 DXD9W-GO 06/24/94 0
PAL METALS	µg/g		µg∕g			pg/g
Aluminum	6280		6130	D	-	7920
Arsenic	6.46		13.2	D		20.3
Barium	15.4		11.2	D		19.8
Boron	< 6.64		12.9	D	<	6.64
Calcium	1200		2660	D		1650
Chromium	1200		14.6	D		23.6
and a second second	5.5	12	2.5			
Cobalt		<	43.2	D D		5.79
Copper	13.1		10000	D		14.2
Iron	15100		14400	1.2		16200
Lead	20.7	- 1 - I	20.7	D		23.6
Magnesium	3080		3990	D		4470
Manganese	195		259	D		203
Nickel	13.4		12.8	D		18.1
Potassium	842		581	D		979
Sodium	104		79	D		132
Vanadium	12.2	-	11.5	D		15.3
Zinc	49.5		41.6	D		48.1
SEMIVOLATILE ORGANICS				-	1.0	
Acenaphthylene	0.34		.098	D	<	0.033
Benzo[a]anthracene	1.9	1.0	0.3	D	1.0	0.16
Benzo[b]fluoranthene	2.2	<	0.31	D	<	0.31
Benzo[ghi]perylene	0.95	<	0.18	D	<	0.18
Benzo[k]fluoranthene	0.99	<	0.13	D	<	0.13
Benzyl Alcohol	< 0.032		0.073	D	<	0.032
Chrysene	2		0.36	D		0.19
Fluoranthene	2.5		0.44	D	100	0.21
Fluorene	0.55	<	0.065	D	<	0.065
Phenanthrene	3.8		0.59	D		0.13
Pyrene	3.8		0.5	D		0.27
Bis(2-ethylhexyl) Phthalate	< 0.48	<	0.48	D		1.4
OTHER						
Total Organic Carbon	2580		1320	D		1670
Total Petroleum Hydrocarbons	188		123	D		131

Notes:

µg/g = micrograms per gram.

< = Concentration is less than the certified reporting limit.

D = Duplicate analysis.

TABLE 5-7 SUMMARY OF SOIL BORINGS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

EXPLORATION ID	COMPLETION DEPTH (Feet bgs)	REFERENCE BAMPLE INTERVALS (Feet bps)	OFF-STIE LABORATORY ANALYTICAL SAMPLES COLLECTED	SOIL. TYPE (USCS)	TEFFAL VOCS BY PID (ppmv)	COMMENTS
		PRI	EVIOUS INVESTIG			
69W-94-09	15	0-2		SW	0.7	
Concernant 1		2-4	2-4	SW	0.5	
-		4-6	4-6	SW	1	and the second sec
provide and a		11-13		SM	0.5	
69W-94-10	15	0-2		SW	15.7	Asphalt to 0.25 feet hgs
		2-4	2-4	OL	279	
	1 i	4-6	4-6	SW	541	A COLUMN THE REAL PROPERTY.
State of the second sec	A descent of	11-13		SW/SP	265	Alternating med and fine sands
69W-94-11	15	0-2		SW	0.5	
	10 C - 10 C	2-4	2-4	SW	0.5	
1.1		4-6	4-6	SW	237	
		11-13		SW	42	
69W-94-12	15.3	0-2		OL	0.9	
		2-4	2-4	OL	1.4	
1.1		4-6	4-6	SW	124	
		11-13		SW	48.4	
69W-94-13	14	0-2		OL	1	There are a marked to be the
		2-4	2-4	PT/GM	33	0-3" fine to med. black sand
1. Sec. 1. 1.		4-6	4-6	SW	520	and the second second second second
		11-13		SW	234	
69W-94-14	13.5	0-2		OL	0.8	
		2-4	2-4	OL/SW	4.1	
		4-6	4-6	SW	154	
		11-13		SW	107.6	
			MEDIAL INVESTIG			
ZWM-95-15X	13	0-2		ML	0	
		2-4		SM-ML	0.8	
		4-6	4-6	ML/SW	4.1	
		6-8		SP	0	
		8-9.5		SW-GW	NR	
		9.5-11		SW-GW	NR	
		11-13		SW-GW	NR	
ZWM-95-16X	17	0-2		SM	NR	
10 million (5-7	70	SP	NR	
		7-9	7-9	SP/SW	NR	
-	at the second second	10-12		SW/SP	NR	
and the series	22	14-16		SP	NR	
ZWM-95-17X	22	0-2		ML-SM	NR	
		5-7		SP	NR	
		10-12		SP-SW SW	NR	-
		12-14	14.16	SW-SP	NR	
and the second se	A second second	14-16	14-16	SW-SP SP	NR	
ZWM-95-18X	- 14	20-22		SP	NR NR	
ZWM-90-18X	14	0-2		SW	NR	
		2-4 4-6	4-6	ML/SM	NR	
			4-0			and the second
		6-8		SW-SM	NR	Strong petroleum-like odor from augers

TABLE 5-7 SUMMARY OF SOIL BORINGS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

EXPLORATION ID	COMPLETION DEPTH (Text.hgs)	REFERENCE BAMPLE INTERVALS (Feet hgs)	OFF-SETE LABORATORY ANALYTICAL SAMPLES COLLECTED	SOIL TYPE (USCS)	TOTAL VOCS BY PID (ppmv)	COMMENTS
ZWM-96-19X	16	1-3	- Y	SP	0	Concrete floor
		3-5		SP	0	
		5-7	5-7	SP	0	and the second se
		7-9	Contra -	SP-SW	0	Fuel odor
		9-11	9-11	SW	48.2	idente entre a constante a
		11-13		SM	50	50 ppm at top, 4 ppm at bottom
		13-15		SP	0	
ZWM-96-20X	13	2-4	2-4	SW	0	Concrete floor
		4-6	4-6	SP	0	
	0.0000000000000000000000000000000000000	6-8		SP	0	
		8-10		SP	0	
ZWM-96-21X	15	0-2	The second second	SW	0	and the second se
		2-4 - 4-6	4-6	SW	0	
		6-8	4-6	SW SP	0	
		28 (P40)	0.10	SP	0	
		8-10 10-12	8-10	SW-SM	0	
		10-12 12-14		SW-SM SW	0	Refusal at 15 ft on probable boulder
ZWP-95-01X	12	0-2		ML-SW	NR	Refusar at 15 ft on probable boulder
LWI-95-01A	12	5-7		SW	NR	
	1 A. M. M. M.	10-12		SW	NR	
ZWP-95-02X	12	0-2		SM-ML	NR	
5,11,25,0221		5-7		SW	NR	
		10-12		SP/GW	NR	
ZWB-95-01X	12	0-2	0-2	ML	0	
	02.0	5-7	5.03	SM	7.6	
		7-9	7-9	SM	3.3	
		10-12		SM	7.5	
ZWB-95-02X	12	0-2	0-2	SM-ML	0	
		5-7	5-7	SW-GW	5.2	·
ZWB-96-03X	15	0-2		NR	0	
		2-4		SP-SM	0	
		4-6		SP	0	
		6-8	6-8	SP	0	
		8-10		SP	0	
		10-12	10-12	SP	0	
		12-14		SW-SM	0	

Notes:

bgs = Below ground surface

USCS = Unified Soil Classification System

ppmv = Parts Per Million by Volume

SW = Well graded sand

SM = Silty sand/sandy silt

OL = Organic silts/clays

SP = Poorly graded sand

PT = Peat

GM = Silty gravel/gravel-sand-silt mixtures

ML = Inorganic silts

GW = Well graded gravel

TABLE 5-8 MONITORING WELL COMPLETION DETAILS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

WELL IDENTIFICATION	SOIL DRILLING METHOD	BEDROCK DRILLING MEHTOD	MEDIUM SCREENED	WELL SCREEN DEPTH (Feet bgs)	WELL SCREEN ELEVATION (Feet NGVD)	COMPLETION DEPTH (Feet bgs)	CONSTRUCTION
REE 69W							
69W-94-09	HSA ¹	NA	SOIL	3.5-13.5	220.0-210.0	15	2" ID PVC
69W-94-10	HSA ¹	NA	SOIL	4.5-14.5	224.2-214.2	15	2" ID PVC
69W-94-11	HSA ¹	NA	SOIL	4.5-14.5	223.2-213.2	15	2" ID PVC
69W-94-12	HSA ¹	NA	SOIL	3-13	223.5-213.5	15.3	2" ID PVC
69W-94-13	HSA ¹	NA	SOIL	3-13	222.3-212.3	13.5	2" ID PVC
69W-94-14	HSA ¹	NA	SOIL	3-13	222.5-212.5	13.5	2" ID PVC
EMEDIAL INVESTIGATI	ON						
ZWM-95-15X	HSA ²	NA	SOIL	3-13	219.9-209.9	13.5	4" ID PVC
ZWM-95-16X	HSA ²	NA	SOIL	6.3-16.3	222.7-212.7	17	4" ID PVC
ZWM-95-17X	HSA ²	NA	SOIL	12.2-22.2	223.9-213.9	22.5	4" ID PVC
ZWM-95-18X	HSA ²	NA	SOIL	3-13	217.7-207.7	14	4" ID PVC
ZWM-96-19X	CHOP & WASH	NA	SOIL	5.8-15.8	225.7-215.7	16	2" ID PVC
ZWM-96-20X	CHOP & WASH	NA	SOIL	2.8-12.8	223.7-213.7	13	2" ID PVC
ZWM-96-21X	CHOP & WASH	NA	SOIL	4.8-14.8	226.4-216.4	15	2" ID PVC
ZWP-95-01X	HSA ²	NA	SOIL	10-12	214.4-212.4	12.5	4" ID PVC
ZWP-95-02X	HSA ²	NA	SOIL	9.5-11.5	211.2-209.2	12	4" ID PVC

Notes: NA = Not Applicable

 $HSA^1 = 4-1/4$ -inch ID hollow stem augers

 $HSA^2 = 6-1/4$ -inch ID hollow stem augers

Chop & Wash = 4-inch ID casing

ID = Inner diameter

PVC = Polyvinyl chloride

NGVD = National Geodetic Vertical Datum

TABLE 6-1 SUMMARY OF WATER LEVEL ELEVATION DATA AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

		DECEMB	ER 7, 1995	MARCE	1 26, 1996	JULY	23, 1996	JANUAR	Y 15, 1997	DECEMBER, 1997 2	
WELL ID 1	ELEV. OF REF POINT (FEET MSL)	DEPTH TO WATER (FEET)	ELEV. OF WATER (FEET MSL)	DEPTH TO WATER (FEET)	ELEV, OF WATER (FEET MSL)	DEPTH TO WATER (FEET)	ELEV. OF WATER (FEET MSL)	DEPTH TO WATER (FEET)	ELEV. OF WATER (FEET MSL)	DEPTH TO WATER (FEET)	ELEV. OF WATER (FEET MSL)
69W-94-09	228.40	5.85	222.55	4.08	224.32	4.61	223.79			7.76	220.64
69W-94-10	227.99	5.81	222.18	4.40	223.59	4.93	223.06	4.39	223,60	7.79	220.20
69W-94-11	227.33	5.22	222.11	3.82	223.51	4.36	222.97	3.82	223.51	7.13	220.20
69W-94-12	228.94	7.33	221.61	5.95	222.99	6.48	222.46	6.00	222.94	7.27	221.6
69W-94-13	227.79	6.22	221.57	3.93	223.86	5.42	222.37	4,95	222.84	8.10	219.69
69W-94-14	228.02	7.24	220.78	6.17	221.85	6.53	221.49	6.16	221.86	9.02	219.00
ZWM-95-15X	225.81	5.23	220.58	4.23	221.58	4.57	221.24	4.24	221.57	6.95	218.86
ZWM-95-16X	228.38	5.73	222.65	3.93	224.45	4.47	223.91	3.88	224.5	7.58	220.80
ZWM-95-17X	238.63	14.98	223.65	12.88	225.75	13.49	225.14	12.68	225.95	16.85	221.71
ZWM-95-18X	222.95	3.48	219.47	2.59	220.36	2.88	220.07	2.61	220.34	5.05	217.90
ZWM-96-19X	231.11							6.78	224.33	10.42	220.65
ZWM-96-20X	225.99		1					1.09	224.90	4.90	221.09
ZWM-96-21X	230.93	-					1	6.52	224.41	10.15	220.7
ZWP-95-01X	226.84	5.45	221.39	4.69	222.15	4.94	221.90	4.61	222.23		
ZWP-95-02X	223.63	4.06	219.57	3.33	220.30	3.51	220.12	3.21	220.42		1

NOTES:

1 The reference point for all the monitoring wells was PVC.

2 Water levels were obtained during the period of December 11 through December 19 (during the low-flow groundwater sampling event).

MSL = Mean Sea Level

PVC = Top of polyvinyl chloride well riser

TABLE 6-2 SUMMARY OF IN-SITU HYDRAULIC CONDUCTIVITY TESTING AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

									Bouwer	and Rice	Hy	orslev	
Weff ID	Well Diam (in)	Screen Int. (feet, bgs)	Filter Pack Int (feet, bga)	Saturated Reight (feet)	Rc (feet)	Rw (feet)	Le (feet)	Hw (feet)	Hydraulic Cond. (ff/min)	Hydraulic Cond. (cm/s)	Hydraulic Cond. (ft/min)	Hydraulic Conc (cm/s)	Ceology (USC)
69W-94-09	2	3.5 to 13.5	2.5 to 15	6.85	0.17	0.29	8.35	8.35	2.60E-03	1.32E-03	1.78E-04	9.02E-05	SP-SM
69W-94-10	2	4.5 to 14.5	3.5 to 15	8.08	0.17	0.29	8.58	8.58	4.10E-02	2.08E-02	2.96E-03	1.50E-03	SW
69W-94-11	4	4.5 to 14.5	3.5 to 15	7.41	0.29	0.46	7.91	7.91	3.10E-02	1.57E-02	3.50E-03	1.78E-03	SP-SM
69W-94-12	2	3 to 13	2.5 to 15	8.26	0.17	0.29	10.26	10.26	1.10E-02	5.59E-03	3.14E-03	1.59E-03	SW
69W-94-13	4	3 to 13	2.5 to 14	9.27	0.29	0.46	10.27	10.27	5.10E-02	2.59E-02	7.40E-03	3.76E-03	SW
69W-94-14	2	3 to 13	2.5 to 13.5	7.62	0.17	0.29	8.12	8.12	5.80E-02	2.95E-02	2.72E-03	1.38E-03	SW
ZWM-95-15X	4	3 to 13	2 to 13.5	9.14	0.29	0.46	9.64	9.64	5.80E-02	2.95E-02	6.07E-03	3.08E-03	SW-GW
ZWM-95-16X	4	6.3 to 16.3	4 to 17	9.06	0.29	0.46	9.76	9.76	1.90E-02	9.65E-03	1.95E-03	9.92E-04	SW-SP
ZWM-95-17X	4	12.2 to 22.2	7 to 22.5	8.75	0.29	0.46	9.05	9.05	3.80E-02	1.93E-02	4.08E-03	2.07E-03	SW-SP
ZWM-95-18X	4	3 to 13	2 to 14	11.65	0.17	0.46	12.65	12.65	5.40E-02	2.74E-02	4.78E-03	2.43E-03	SW-SM
ZWM-96-19X	2	5.8 to 15.8	3.9 to 16	8.6	0.11	0.17	8.8	8.8	8.44E-03	2.74E-02	1.46E-03	7.39E-04	SP-SW-SM
ZWM-96-20X	2	2.8 to 12.8	2.5 to 13	11.21	0.11	0.17	11.41	11.41	8.67E-03	2.74E-02	1.44E-03	7.31E-04	SP
ZWM-96-21X	2	4.8 to 14.8	3.0 to 15.0	8.04	0.11	0.17	8.24	8.24	8.84E-03	2.74E-02	1.42E-03	7.24E-04	SP-SW

Notes: Data analyzed using AQTESOLV (Bouwer & Rice Solution).

Data analyzed using AQTESOLV (Bouwer & Rice Solution). All tests are rising head tests. Re = Well easing radius for fully saturated filterpacks and equivalent easing radius which accounts for filterpack resaturation at n=30% for partially saturated filterpacks. Rw = Radius of borehole.

kw = kannes or borenole.Le = Saturated length of filterpack. Hw = Height of Water Column above filterpack bottom.Saturated Height is height of water column measured in well. All measurements in feet unless otherwise noted.

¥.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:	ZWB-95-01, BDZW0100 09/19/95 0		ZWB-95-01 BXZW010 09/19/95 0		ZWB-95-01X BXZW0197 09/19/95 7		ZWB-95-02X BXZW0206 09/19/95 0		ZWB-95-02/ BXZW0295 89/19/95 5		ZWR 95-26X RXZW2607 09/11/95		ZWR.95.30X RXZW3006 09/11/95		ZWR 95-35 RXZW350 09/12/95	4
Units	на/а		₽₽%g		µg∕g		₽g/g		<i>не</i> /е		µg/8		H8/8		4 6 /8	
SEMIVOLATILE ORGANICS																
1,2,3-trimethylbenzene					30	S				-						
1,3-dimethylnaphthalene									1.000							
1,4,6-trimethyl-naphthalene 1,4-dimethylnaphthalene									2.1	S	2.2	S	2.3	S		
1,4-onneutyinaphthalene					80	S										
1,5,7-dimethylnaphthalene					50	S						1				
1.8-dimethylnaphthalene					100	S					1.1					
1-ethyl-2-methylbenzene					20	S		1	S		1.1	S				
1-methyl-3-(1-methylethyl)benzene / M-cymene					20	0							0.55	s		
1-methylnaphthalene					80	s							0.56 3.4	S		
2,3,6-trimethylnaphthalene					00	2			4.1	s			3.4	9		
2,3-dihydro-5-methylindene					20	S			.4.1	3						
2,3-dimethylnaphthalene					30	S								1		
2,5-dimethylphenanthrene					50								2.3	s		
2,6,10,14-tetramethylpentadecane	3	5	5	S	300	S			21	S	8.8	s	23	S		
2,6,10,15-tetramethylheptadecane					200	-					0.0	~	23			
2,7-dimethylnaphthalene					100	s								- 1		
2.8-dimethyldibenzo[b,d]thiophene					200								1.1	s		
3-propyltoluene					50	S						1		-		
4mpanr													1.1	S		
Docosane										- 1	2.2	S	0.68	S		
Dodecane										- 1	1.1	S	04040			
Eicosane			2	S							5,5	S				
Gamma-sitosterol																
Heneicosane			2	S							3,3	S				
Heptacosane																
Heptadecane	2	S	З	S							11	S				
Hexadecane	2	S	2	S							11	S				
Hexadecanoic Acid / Palmitic Acid		1.0										1.1			0.43	S
Isopropyltoluene					30	S								- 1		
Nonacosane																
Nonadecane Octadecane											7,7	S	5.6	S		
	2	S	3	S							11	S				
Pentadecane Sulfur, Molecular			2	S					-45	1.20	11	S				
Tetradecane									7.2	S	3.3	S	1.1	S		
Tricosane								- 1			11	S				
Tridecane											.99	S				
Unk514					80	s					5.5	S				
Unk515					0V	3	10	s							07	
Unk524							10	0							0.7	S
Unk537					20	S										
Unk544					20				2	S						
Unk545					60	s			2				0.7	s		
Unk546					20	s							0.7	9		
Unk548									2	S			0.9	S		
Unk549					30	S			2	S			0.8	S		
Unk550					30	S			ĩ	s			0.6	S		

04/23/98 4:05 PM

14

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number: Sample Date: Depth:	ZWB-95-01X BDZW0109 09/19/95 0	ZWB-95-01X BXZW0100 09/19/95 0	ZWB-95-01X BXZW0107 09/19/95 7	ZWB-95-02X BXZW0209 03/19/95	ZWB-95-02X BXZW0205 09/19/95	ZWR-95-26X RXZW2607 09/11/95	ZWR-95-30X RXZW3906 89/11/95	ZWR-95-35X RXZW3504 09/12/95
	Units:	₽E/g	PR/S	prg/g	0 µg/g	\$ µg/g	\$ #\$/g	0 149/8	0 P(g/g
Unk551			FO B	152		100	Parts		S
Unk552							1 m		S
Unk553				20	S	1.1			
Unk554					S		S		S
Unk555					S		S	0.7	S
Unk556					S	2	8		S
Unk557				2.7	S				S
Unk558					S				S
Unk559					S	1	S		S
Unk560				20	S				S
Unk561				3.0					S
Unk562					S		4 1 1 2 2 1		S
Unk563					S				S
Unk564 Unk565					S	2	S		S
Unk566					S S				S S
Unk567					S	2	S 0.4		5
Unk568					s	6	A CONTRACT OF A		S
Unk569					S	0			s
Unk570					S				S
Unk571					S	and the second second			S
Unk572					S	2	S	-	~
Unk573					S		0.6	S	
Unk574					S			S	
UnkS75					S		0.7	S	
Unk576				30	S		0.8	S	
Unk577						A CONTRACT OF	0.8	S	
Unk578					S		S 2	S	
Unk579					S	9	8 3	S	
Unk580					S				
Unk581				20	S			S	
Unk582						ev.		s	
Unk583					4	10		S	10
Unk584 Unk585			2		S			S	-
Unk586				100	S 5	10	S 1	5 3	S
Unk587					5		1 1	s	
Unk588					s	2000	s 1	5	
Unk589					s	2000		s	
Unk590					S	4		S	
Unk591								S	
Unk592				30	S			s	
Unk593					S		and a second		
Unk594					S				
Unk595							2	S 9	S
Unk596				50	S			S	-
Unk597									S
Unk598					S	3			S
Unk599	numerical Commit abide seals 1 wis		1	30	S	-	1	S ,	04/03/00 4.05 PM

2

g.'projects/usaec/projects/69writab/dscsob1.xls

04/23/98 4:05 PM

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Field Sample Number: Sample Date: Depth:	ZWB-95-013 BDZW0100 09/19/95 0		ZWB-95-01X BXZW0100 09/19/95 0		ZWB-95-01X BXZW0107 09/19/95 7		WB-95-02X BXZW0200 09/19/95 0	ZWB-95-02X BXZW0205 09/19/95 \$		ZWR-95-26X RXZW2607 09/11/95		ZWR-95-30X RXZW3006 09/11/95 \$		ZWR-95-35X RXZW3504 09/12/95 0
	Units:	#8/8		µg/g		PE/E		<u>ня/я</u>	µg/g		48/g		M8/8		µg/g
Unk600						4					2	S			
Unk601 Unk602		2	S	4	S				10	S	6	S			
Unk603		300	S	500	s					1.1	0.9	S			
Unk604		500	3	3	S	30	s		5	S			1	s	
Unk605					~	50				0		1.15	3	S	
Unk607													2	~	
Unk608										- 1			1	s	
Unk609															
Unk610													1	S	
Unk611									1				0.7	S	
Unk612													1	S	
Unk613												- 11			
Unk614														- 11	
Unk615 Unk616				1.								11.			
Unk617															
Unk618															
Unk619											1	s	2	s	
Unk620											a.		-		
Unk621			- 00												
Unk622												-			
Unk623															
Unk624															
Unk625															
Unk626														- 1	
Unk627												- 11		- 11	
Unk628												- 11			
Unk629 Unk630										- 11					
Unk631															
Unk632															
Unk633															
Unk635															
Unk636															
Unk637															
Unk638															
Unk641									0						
Unk642															
Unk643															
Unk644															
Unk645															
Unk650															
Unk651															
Unk654					- 4		1111						4		

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date:	ZWB-95-01X BDZW0109 09/19/95	ZWB-95-01X BXZW0100 09/19/95	ZWB-95-01X BXZW0107 09/19/95	ZWB-95-02X BXZW0300 09/19/95	ZWB-95-02X BXZW9205 09/19/95	ZWR-95-26X RXZW2607 09/11/95	ZWR-95-30X RXZW3006 09/11/95	ZWR-95-353 RXZW3504 09/12/95	
Depth: Units:	0 #g/g	0	7	0	\$	0	\$	•	
VOLATILE ORGANICS	HK/R	PB/g	<u>1 #8/g</u>	P#/g	µg/g	µg/g	148/8	pig g	
1,1,3-trimethylcyclohexane 1,2,3-trimethylbenzene			0.19	S					
1,2,4-trimethylbenzene 1,3,5-trimethylcyclohexane 1,4,6-trimethyl-naphthalene 1,4-dihydro-1,4-methanonaphthalene 1,7-dimethylnaphthalene									
I-ethylidene-Ih-indene I-methyl-3-(1-methylethyl)benzene / M-cymene I-methylethylcyclohexane						0.33 S		0.053	s
l-methylnaphthalene 2,3,5-trimethyldecane 2,3-dimethylnaphthalene					0.41	s		0.21	S
2,3-dimethyloctane 2,6-dimethylundecane 2,7-dimethylnaphthalene					0.52	S		0.11	S
2-ethyl-1-hexanol 2-methyldecane 2-methylheptane / Isooctane 2-methylnaphthalene			0.23	S	0.31	s		0.21	S
2-Incity inspiration 26dmo 3.7-dimethylnonane 3-methyl-5-propylnonane			0.13	5	0.21	S 0.22 S			
3-methyldecane 3-methyloctane 3-methylundecane			0.097	s				0.075	S
3-propyltoluene 4-(1-methylethyl)toluene / P-cymene 4-methyldecane 6-methyldodecane					0.31	0.11 S 0.33 S 0.33 S		0.11	S
Cumene / Isopropylbenzene Decahydro-2-methylnapthalene Decane						0.22 S			
Dodecane Hendecane / Undecane Heptacosane Nonane			0.097	s	0.21	s 0.11 S			
Ometsx Octane Pentadecane			0.097	s					
Trans-1-ethyl-4-methylcyclohexane Unk011 Unk130			0.06	S					
Unk135 Unk136 Unk143 Unk147			0.3	s s					
Unk151		0.02	s 0.2	5			+		

g:\projects\usaec\projects\69writab\dscsob1.xls

04/23/98 4:05 PM

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number: Sample Date: Depth: Units:	ZWB-95-01X BDZW0109 09/19/95 0 8	ZWB-95-01X BXZW0100 09/19/95 0 µg/g	ZWB-95-01X BXZW0107 0%19/95 7 PE/E		ZWB-95-02X BXZW0109 09/19/95 0 FE/E		ZWB-95-02X BXZW0205 09/19/95 5 8		ZWR-95-26X RXZW2607 09/11/95 0 µg/g		ZWR-95-30X RXZW3906 99/11/95 9 99/2	ZWR-95-35 RXZW3504 09/12/95 0 #E/g	
Unk152 Unk158				0.1	s								0.07	F
Unk162 Unk166							- 1						0.07	9
Unk172														
Unk173 Unk175									- 1					
Unk178													0.07	S
Unk179 Unk182								0.2 0.1	S S				1.000	
Unk183								0.1		0.6	S			
Unk186				0.1	S		- 1			0.3	s		0.07	S
Unk187 Unk188			0.01 5	5		0.005	s			0.3	8			
Unk189			1000								1.1		0.06	S
Unk190 Unk191				0.06	S			0.2 0.4	S	0.2	s		0.1	s
Unk193														
Unk194								0.1	S	0.3	S		0	
Unk196 Unk200														
Unk202								0.2	S	1.00	1		0.1	S
Unk205 Unk206	1								1.1	0.2	S			
Unk215					_									

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID Field Sampie Number: Sample Date: Depth:	ZWR-95-362 RXZW3607 09/13/95 0		RXZW3704 09/13/95 0		ZWR-95-38X RXZW3803 09/14/95	ZWR-9 RXZW 09/1	3904 V95	ZWR-95-45X RXZW4594 09/15/95	ZWS-95-35X SXZW3500 09/18/95	ZWS-95-37X SXZW3790 09/18/95	ZWS-95-37 SXZW370 09/28/95	þ
Units	PE/S		ружи 2/24		3 #g/g	0 #8		4 8/344	0 #8/8	0 µg/g	0 ##/g	
SEMIVOLATILE ORGANICS					1.4.6							
1,2,3-trimethylbenzene		T										
1,3-dimethylnaphthalene 1,4,6-trimethylnaphthalene 1,4-dimethylnaphthalene 1,6,7-trimethylnaphthalene 1,7-dimethylnaphthalene			3.5 2.3	S S								
1,8-dimethylnaphthalene 1-ethyl-2-methylbenzene												
l-methyl-3-(1-methylethyl)benzene / M-cymene l-methylnaphthalene												
2,3,6-trimethylnaphthalene 2,3-dihydro-5-methylindene 2,3-dimethylnaphthalene			2.3 .81	s s				1.1				
2,5-dimethylphenanthrene			.58	s								
2,6,10,14-tetramethylpentadecane	4.3	s	.56	0		0					1.1	S
2,6,10,15-tetramethylheptadecane 2,7-dimethylnaphthalene 2,8-dimethyldibenzo[b,d]thiophene	1.1	s										3
3-propyltoluene 4mpanr												
Docosane	1.1	S										
Dodecane												
Eicosane	3.2	S							911			
Gamma-sitosterol											.77	S
Heneicosane	2.2	S										
Heptacosane	1.1.1										1.1	S
Heptadecane	7.6	S										
Hexadecane Hexadecanoic Acid / Palmitic Acid	5.4	S										
Isopropyltoluene Nonacosane											.55	s
Nonadecane	4.3	s									3.3	
Octadecane	5.4	S		1								
Pentadecane	6.5	S									.66	S
Sulfur, Molecular	.32	S	6.9	S	.79	S						
Tetradecane	5.4	S										
Tricosane	.76	S										
Tridecane	2.2	S										
Unk514												
Unk515 Unk524												
Unk537											.5	S
Unk544												
Unk545												
Unk546										0.00		
Unk548												
Unk549												
Unk550									-			

6

g/projects/ussec/projects/69writab/dacaob1.xla

04/23/98 4:05 PM

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number: Sample Date:	ZWR-95-36X RXZW3607 09/13/95		RXZW3704 09/13/95		WR-95-38X RXZW3803 09/14/95	ZWR-95-39X RXZW3964 09/13/95	ZWR-95-45X RXZW4594 05/15/95	ZWS-95-35X SXZW3500 09/18/95	ZWS-95-37X SXZW3700 09/18/95	ZWS-95-37X SXZW3799 99/28/95
	Depth: Units:	0 PR/g		0		3	θ	4	0	0	0
Unk551	emts.	HK 8		₽g/g	-	#g/g	PE/E	HR/R	PR/g	#B/g	14K/g
Unk552							() · · · · · · · · · · · · · · · · · · ·			A CONTRACTOR OF THE	
Unk553											
Unk554				0.3	S		0				
Unk555				0.3	S		2		0		
Unk556					E.				TX as a second		
Unk557		0.5	S	0.9	S						
Unk558		03	S	0,5	S			1.1.2.1			
Unk559					S				- X		
Unk560				0.6	S						
Unk561				0.5	S						
Unk562				0.3	S						
Unk563		1	S	2	S						
Unk564		0.5	S	0.7	S						
Unk565				0.3	S						
Unk566		0.4	S	0.6	S					100	
Unk567		0.5	S	0.7	S						
Unk568		0.6	S	0.7	S						
Unk569		2	S	1	S						
Unk570		0.8	S	1	S						
Unk571				0,6	S						
Unk572		1	S	0.6	S			8			
Unk573				3	S						
Unk574				0.8	S						
Unk575		0.6	S	1	S					S0.1	
Unk576		0.8	S	1	S						
Unk577		1	S	1	S						
Unk578		3	S	2	S					1	
Unk579		0.0	1.0	3	S						
Unk580		0.8	S	1	S						
Unk581		0.5	S	0.7	S						
Unk582		2	S	1	S		12				
Unk583		2	S		S						
Unk584		I	S	1	S						
Unk585 Unk586		5	S	1	S						
Unk587		0.6 0.9	S		S						
Unk588		0.9	S	6 0.5	S						
Unk589		2	s		S						
Unk590		1	S		S						
Unk591		0.8	S		S						1
Unk592		0.0	5		S						
Unk593				3	S						
Unk594		1	s		S						
Unk595		i	S		S						
Unk596		0.9	S	0.6	S					9	
Unk597		0.5			S						
Unk598		1	s		S						
Unk599		i	s	0.6	s				1		

g/projects/usaec/projects/69writab/dscsob1_xls

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Field Sample Number: Sample Date: Depth:	ZWR-95-36X RXZW3607 09/13/95 0	RXZW3704 09/13/95 0	ZWR-95-38X RXZW3863 09/14/95 3	ZWR-95-39X RXZW3904 09/13/95 0	ZWR-95-45X RXZW4504 09/15/95 4	ZW5-95-35X SXZW3500 09/18/95 0	2W5-95-37X SX2W3700 09/18/95 0	ZWS-95-37X SXZW3700 09/28/95 0
	Units	PE/g	μg/g	µg/g	PR/R	PK/E	PB/g	pg/g	PB/g
Unk600			2	S					
Unk601		3	S 0.5	S					0.9 S
Unk602			0.6	S					
Unk603 Unk604			2	S					
Unk605			3	S S					
Unk607			0.5	S					
Unk608			0.7	S					
Unk609			0.9	S					
Unk610			1	S					
Unk611			i	S					
Unk612			2	S					
Unk613			0.9	S					
Unk614			1	S					
Unk615			1	S					
Unk616			20	S					
Unk617			2	S					
Unk618			1	S					
Unk619		4	S 3	S	0.7	S 0.6	S		
Unk620			3	S					
Unk621			1	S					
Unk622			2	S					
Unk623			3	S			1.2		
Unk624			0.7	S					
Unk625 Unk626	r		2	S S					
Unk627			1	2					
Unk628			0.6	S					
Unk629			1	S					
Unk630			1	S					3 S
Unk631			0.3	S					3 5
Unk632			0.7	S					
Unk633			1	S					
Unk635				18 J.					4 S
Unk636								1.00	3 S
Unk637									1 5
Unk638									0.9 S
Unk641									2 S
Unk642									
Unk643									1 A A
Unk644									1 S
Unk645									
Unk650									1 S
Unk651									0.7 S
Unk654	the second se								0.7

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID Field Sample Number: Sampte Date Depth	ZWR-95-36X RXZW3607 09/13/95 0		RXZW3704 99/13/95 6		ZWR-95-38X RXZW3803 09/14/95 3		ZWR-95-39X RXZW3904 99/13/95 0		ZWR-95-45X RXZW4504 09/15/95		2WS-95-35X SX2W3500 09/18/95		ZWS-95-37X SXZW3700 09/18/95		ZWS-95-37X SXZW3760 09/38/95
Units	#8/8		µr/g		₽₽/g		₽₽/g		#8/8		PE/S		₽8/g		PE/E
OLATILE ORGANICS															
1,3-trimethylcyclohexane			0.12	S							10.0				
2,3-trimethylbenzene											0.032	S			
2,4-trimethylbenzene		1.01									0.043	S		-	
3,5-trimethylcyclohexane			0,23	S						1.1					
4,6-trimethyl-naphthalene									0.0087	S			0.0052	S	
4-dihydro-1,4-methanonaphthalene	1.15														
7-dimethylnaphthalene	0.054	S			0.0079	S				- 1					
ethylidene-1h-indene													0.01	S	
methyl-3-(1-methylethyl)benzene / M-cymene										1					
methylethylcyclohexane						1.1					0.022	S			
methylnaphthalene					0.068	S							0.01	S	
3,5-trimethyldecane														1.1	
3-dimethylnaphthalene					0.0056	S									
3-dimethyloctane							0.022	S							
6-dimethylundecane							0.066	S							
7-dimethylnaphthalene					0,0056	S								- 1	
ethyl-1-hexanol									0.0099	S					
methyldecane															
methylheptane / Isooctane														- 1	
methylnaphthalene					0.023	S								- 1	
ódmo			0.23	S								- 1			
7-dimethylnonane															
methyl-5-propylnonane	0.076	S													
methyldecane						- 1									
methyloctane								- 02							
methylundecane	0.11	S				- 1							1.1		
propyltoluene												1			
(1-methylethyl)toluene / P-cymene	1144							- 0							
methyldecane	0.054	S					0.011	S		- 1					
methyldodecane	0.22	S													
umene / Isopropylbenzene											0.022	S			
ecahydro-2-methylnapthalene										- 1					
ecane			0.23	S							0,097	S			
odecane	0.11	S									0.054	S			
endecane / Undecane										- 1	0.086	S			
eptacosane															
onane															
metsx													0.21	S	
stane															
mtadecane							0.088	S							
rans-1-ethyl-4-methylcyclohexane			0.12	S							22.0				
nk011											0.1	S			
nk130 nk135															
		1					0.02	S							
nk136			6.0										0.008	S	
nk143			0.3	S											
nk147 nk151		- U.													

9

g:\projects\usaec\projects\69writab\dscsob] xls

04/23/98 4:05 PM

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Pield Sample Number: Sample Date Depth:	ZWR-95-36X RXZW3607 09/13/95 0		RXZW3704 09/13/95 0		ZWR -95-38X RXZW3803 09/14/95 3		ZWR-95-39X RXZW3964 09/13/95 0		ZWR-95-45X RXZW4594 89/15/95 4	ZWS-95-35X SXZW3500 99/18/95 9		ZWS-95-37X SXZW3700 09/18/95 0		2WS-95-37X SXZW3700 09/28/95 0
	Units	µg/g		PR/S		₽8/g		µg/g		₽g/g	<u>нв/в</u>		₽₽/g		PHR/R
Jnk152				0.3	S										
Unk158											1				
Jnk162								0.02	S		12				
Unk166						0.009	S					1.5			
Jnk172									1.1		0.05	S			
Unk173								0.01	S		-				
Unk175															
Unk178				0.1	S			0.03	S		1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1. 1	1.1			
Unk179											0.04	S			
Unk182		0.08	S				- 1								
Unk183			1.17	0.2	S							- 1			
Unk186											1				
Unk187			1.1						1				0.2	S	
Unk188		0.1	S								0.1	S			
Unk189				0,1	S										
Unk190				0.2	S										
Unk191					1.2			0.03	S						
Unk193				0.2	S										
Jnk194		0.05	S												
Unk196									1.11		1		0.005	S	
Unk200		0.06	S									1		- 1	
Unk202								0.02	S					- 1 I	
Unk205											1 2				
Unk206												- 1			
Unk215		0.06	S			0.01	S							-	

1.

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Fleid Sample Number: Sample Date:	ZWS-95-39X SXZW3900 09/18/95	ZWS-95-38X SXZW3800 09/18/95		ZWS-95-45X SXZW4500 09/18/95		ZWS-95-46X SXZW4600 09/18/95	ZWS-95-47X SXZW4700 09/18/95	
Depth:	0	0		0		0	0	
Units:] SEMIVOLATILE ORGANICS	₽g/g	pg/g		#g/g		ha/a	#R/g	
,2,3-trimethylbenzene		1	T		1		1	-
1,3-dimethylnaphthalene 1,4,6-trimethylnaphthalene 1,4-dimethylnaphthalene 1,6,7-trimethylnaphthalene 1,7-dimethylnaphthalene 1,8-dimethylnaphthalene 1-ethyl-2-methylbenzene 1-methyl-3-(1-methylethyl)benzene / M-cymene 1-methylnaphthalene 2,3,6-trimethylnaphthalene 2,3-dimethylnaphthalene 2,3-dimethylnaphthalene								
2,5-dimethylphenanthrene 2,6,10,14-tetramethylpentadecane 2,6,10,15-tetramethylheptadecane 2,7-dimethylnaphthalene 2,8-dimethyldibenzo[b,d]thiophene								
3-propyltoluene fmpanr Docosane								
Dodecane								- 1
Eicosane								
Gamma-sitosterol Heneicosane Heptacosane		3	S	.76	S		57	
Heptadecane Hexadecane Hexadecanoic Acid / Palmitic Acid							.52	S
Isopropyltoluene								- 1
Nonacosane		6	s	.76	s			- 1
Nonadecane		0	0	-10	0			
Detadecane								
Pentadecane								
Sulfur, Molecular			- 1	5	S			
Fetradecane				3				
Tricosane			- 1					
Tridecane								_ I
Unk514		1						_ I
Unk515								- 1
Unk524								- 1
Unk537				÷				
Unk544								- 1
Unk545								
Unk546								
Unk548								
Unk549 Unk550								

g:\projects/usaec\projects/69writab\dscsob1 als

04/23/98 4:05 PM

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Field Sample Number: Sample Date	ZWS-95-39X SXZW3900 09/18/95	ZWS-95-38X SXZW3800 09/18/95	ZWS-95-45X SXZW4590 09/18/95	ZWS-95-46X SNZW4600 03/18/95	2WS-95-47X SXZW4700 09/18/95
	Depth: Units:	0 #g/g	0 µg/g	0 PE/E	6 µg/g	0 148/g
Unk551			1			
Unk552						
Jnk553						
Jnk554				2		
Jnk555						
Jnk556						
Jnk557						
Jnk558						
Jnk559						
Jnk560						
Jnk561						
Jnk562						
Jnk563						1 C 0
Jnk564						
Jnk565						
Jnk566						
Jnk567						
Jnk568						
Jnk569						
Jnk570						1
Jnk571						
Unk572						
Unk573						
Jnk574						
Jnk575						
Jnk576						
Jnk577						
Jnk578 Jnk579						
Jnk580						
Jnk581						
Jnk582						
Jnk583						
Jnk584						
Jnk585						
Jnk586						
Jnk587		1.1				
Jnk588						
Jnk589						
Jnk590						
Jnk591						
Jnk592						
Jnk593						
Jnk594						
Jnk595						
Jnk596						
Jnk597						
Jnk598						
Jnk599						

1.0

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number; Sample Date Depth;	2WS-95-39X 5XZW3900 09/18/95 0	2WS-95-38X SXZW3899 09/18/95 0		ZW5-96-453 SXZW4500 0%18/95 0		2WS-95-46X SXZW4600 D\$/18/95 0	ZWS-95-47X SXZW4700 09/18/95 0	
	Units	PE/g	PR/g		#8/8		#8/g	₽g/g	
Unk600 Unk601			-						
Jnk602									0.04
Jnk603									
Jnk604									
Jnk605									
Ink607									
Ink608									
Ink609								1.5	
Ink610									
Ink611									
Ink612									
Ink613				1					
nk614				- 1					
Ink615									
Ink616									
nk617									
nk618									
Ink619									
Jnk620				- 1					
Jnk621									
Jnk622	· · · · · · · · · · · · · · · · · · ·								
Jnk623									
Jnk624									
Jnk625									
Jnk626				- 1					
Jnk627			4	S					
Jnk628								1	
Jnk629									
Jnk630								3	S
Jnk631									1.1
Jnk632			3	S				.3	S
Jnk633								200	
Jnk635			5	S				0.9	S
Jnk636			3			S		5	, di
Ink637			6	S	0,8	S		1	S
Ink638									
Ink641									
Jnk642					-			1	S
Jnk643 Jnk644					1	S			
Ink645									
Jnk650			3	S				1	
Jnk651			3	s				1 3	S
Jnk654			,	0					

-

14

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Pield Sample Number: Sample Date: Depth;	ZWS-95-39X SXZW3900 \$3/18/95 0	2WS-95-38X SXZW3890 09/18/95 0	ZWS-95-45X SXZW4500 09/18/95 0	2WS-95-46X SXZW4600 09/18/95 0	2W3-96-47X SX2W4700 03/18/95
Units	PE/E	µ₽/g	μg/g	#g/g	PE/S
VOLATILE ORGANICS					
,1,3-trimethylcyclohexane					
1,2,3-trimethylbenzene			A		The second secon
1,2,4-trimethylbenzene					
,3,5-trimethylcyclohexane					
1,4,6-trimethyl-naphthalene					
1,4-dihydro-1,4-methanonaphthalene		0.0071	S		
1,7-dimethylnaphthalene					
1-ethylidene-1h-indene					
1-methyl-3-(1-methylethyl)benzene / M-cymene					
I-methylethylcyclohexane					
I-methylnaphthalene					
2,3,5-trimethyldecane					
2,3-dimethylnaphthalene					
2,3-dimethyloctane					
2,6-dimethylundecane					
2,7-dimethyInaphthalene 2-ethyl-1-hexanol					
2-methyldecane					
2-methylheptane / Isooctane					
2-methylnaphthalene					
26dmo					
3,7-dimethylnonane					
3-methyl-5-propylnonane					
3-methyldecane					
3-methyloctane					
3-methylundecane					
3-propyltoluene					
4-(1-methylethyl)toluene / P-cymene					
4-methyldecane					
6-methyldodecane					
Curnene / Isopropylbenzene					
Decahydro-2-methylnapthalene					
Decane					
Dodecane					
Hendecane / Undecane					
Heptacosane					5
Nonane			G		
Ometsx		0.24	S		
Octane					
Pentadecane					
Irans-1-ethyl-4-methylcyclohexane					
Unk011 Unk130					
Unk130					
Unk135					
Unk136					
Unk145					
Unk151					

04/23/98 4:05 PM

1.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number: Sample Date: Depth: Units:	ZWS-95-39X SXZW3900 09/18/95 0 µg/g	ZWS-95-38X SXZW3800 09/18/95 0 PE/8		2WS-95-45X SX2W4500 09/18/95 0 #8/8		ZWS-95-46X SXZW4600 09/18/95 0 #8/8		2WS-95-472 SX2W4700 09/18/95 0 #8/8	
Unk152		Pa B	PRB		1					
Unk158										
Unk162										
Unk166										
Unk172										
Unk173			1. 1. 1. 1. 1.							
Unk175			0.006	S						
Unk178			1000							
Unk179			0.007	S						
Unk182								- 1		
Unk183										
Unk186										
Unk187			0.06	S						
Unk188		0.02			0.008	S	0.006	S	0.03	S
Unk189										
Unk190				- 1						
Unk191										
Unk193										
Unk194										
Unk196										
Unk200										
Unk202				- 1						
Unk205			100							
Unk206			0,007	S						
Unk215				-				-		

TABLE 7-1 UNKNOWN COMPOUNDS DETECTED IN GROUNDWATER AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

1.1

Site ID Field Sample Number: Sample Date: Depth:	69W94-10 MXZW10X3 11/02/95 9.5		69W-94-10 MXZW10X4 02/14/96 9.5		69W-94-11 MDZW11X4 92/14/96 9.5	69W94-11 MXZW11X4 02/14/96 9.5	69W-94-12 MXZW12X3 11/02/95 8	69W-94-12 MXZW12X4 02/13/96 8
Units:	µg/L		µg/L		µg/L	µgЛ	µg/Ն	<u> друг.</u>
SEMIVOLATILE ORGANICS								
1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin			100					
1,2,3,4-tetramethylbenzene			400	S				
1,2,3,5-tetramethylbenzene / Isodurene								
1,2-dimethylnaphthalene								
1,3,5-trimethylbenzene								
1,3-dimethyl-5-(1-methylethyl) Benzene			2000		20 U			
1,3-dimethylnaphthalene		- 11	2000	S				
1,4,6-trimethyl-naphthalene			700	S				
1,4-dimethylnaphthalene	800	S						
1,5-dimethylnaphthalene	500	S						
1,6,7-trimethylnaphthalene			800	S				
1,6-dimethylindan								
1,6-dimethylnaphthalene	600	S	2000	S				
1,7-dimethylnaphthalene			2000	S				
1,8-dimethylnaphthalene							3	
1-methyl-3-(1-methylethyl)benzene / M-cymene	200							
I-methylnaphthalene	300	S	2000	S				
2,3,6-trimethylnaphthalene			500	S				
2,3-dimethylnaphthalene			C000					
2,6,10,14-tetramethylpentadecane	(00		5000	S				
2,6-dimethylnonane	600	S						
2,7-dimethylnaphthalene	***							
2-methyldecane	500	S S						
3-methyldecane	500	2						
4,7-dimethylindan / 2,3-dihydro-4,7-dimethyl-1h-indene				11				
4-(1-methylethyl)toluene / P-cymene								1
Heptadecane Hexadecane								
Indan / 2,3-dihydro-1h-indene Nonadecane			700	s				
Octadecanei Octadecanoic Acid			/00	0				
Pentadecane								
Tetradecane								
Toluene					4 SD			
Tridecane					4 51/			

TABLE 7-1 UNKNOWN COMPOUNDS DETECTED IN GROUNDWATER AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID: Field Sample Number: Sample Date: Depth:	69W-94-10 MXZW10X3 11/02/95 9.5		69W-94-10 MXZW10X4 02/14/96 9.5	69W94-11 MDZW11X4 02/14/96 9.5	69W94-11 MXZW11X4 02/14/96 9.5	69W-94-12 MXZW12X3 11/02/95 8	69W-94-12 MXZW12X4 02/13/96 8
	Units:	μg/L		μg/L	μg/L	₽g/L	μg/L	μg/l.
nk525						The second s		
ink543					S			
Jnk544					S			
Jnk546				800	S			
Jnk547				400	S			
Jnk548					S			
Jnk549			- 1	600	S			
Ink550					S			
Ink551					S	0		
Jnk552					S			
Ink553				400				
Jnk554		200	s	600	S			
Jnk555		200			S			
nk556				600	S			1
				600	2			1
Jnk557		200		400	S			
Jnk558		200	S		S			
Ink559		200	S	600	S			
Jnk 560		200	S		S			
Jnk561				700	S			
Jnk562				600	S			
Jnk563		500	S		S			1
Jnk564					S			
Jnk565					S			
Jnk566					S			
nk567		700	S	600	S			12
nk568				800	S			
Jnk569				600	S			
nk570				300	S			
nk571				2000	S			
ink572		800	s		S		1	
Ink573			-	600	S			
Jnk574				500	S			
Jnk575		800	S	2000	S			
Jnk576			~	2000	S			
nk577				600	S			
nk578				600	S			
Ink579				600	S			
Ink580		500		2000	S			
ink581			S	2000	0			
		2000	2				- ÷	
nk582		0000		500				
nk583		9000	S		S			
nk584					S			
nk585				300000	S			

TABLE 7-1 UNKNOWN COMPOUNDS DETECTED IN GROUNDWATER AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth; Units:	69W-94-10 MXZW10X3 11/02/95 9.5 μg/L		69W-94-10 MXZW10X4 02/14/96 9.5 μg/L		69W-94-11 MDZW11X4 02/14/96 9.5 μg/L		69W-94-11 MXZW11X4 02/14/96 9.5 µg/L		69W-94-1 MXZW12X 11/02/95 8 µg/L		69W-94-12 MXZW12X 02/13/96 8 μg/L	4
Unk586	իքո	-	800	S	μgır	-	4grs		pgru		μg/u	
Unk587			2000	s								
Unk588		- 1	400	S								
Unk589			800	S								
Unk590			500	S								
Unk591			2000	S		- 11						
Unk592	2000	s	600	S				- 8				
Unk592	2000	0	400	S								
Unk594			400	S								
Unk595	3000	s	2000	S		- 11						
Unk595	3000	2		S								
			600	3								
Unk597			600	S								
Unk598		- 11	600	S								
Unk599	2000		500	S								
Unk600	2000	S				- 1						
Unk601			2000	S								
Unk602			400	S		- 0				- 1		
Unk603			400	S					- C -	- 1		
Unk610			300	S								
Unk614				1	1.1							
Unk645					5	SD	and the second second				1000	
Unk649				-	50	SD	20	S			9	S
VOLATILE ORGANICS												-
1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin	40	S							20	S		
1,2,3,4-tetramethylbenzene												
1,2,3,5-tetramethylbenzene / Isodurene	20	S										
1,2,3-trimethylbenzene									20	S		
1,2,4-trimethylbenzene												
1,3,5-trimethylbenzene	20	S										
1,3-diethylbenzene												
1-methyl-3-(1-methylethyl)benzene / M-cymene		1.1										
4-(1-methylethyl)toluene / P-cymene	30	S										
C2aec		1.1					6	S			30	S
Cumene / Isopropylbenzene	60	S										
Indan / 2,3-dihydro-1h-indene	50	S	10	S					20	S		
Sec-butylbenzene / (1-methylpropyl)benzene										0.0		
Unk162									7	S		
Unk173						- CL			7	, S		
Unk185	10	S							10	S		
Unk187									20	S		
Unk188	30	S							20	S		
Unk193	20	S							30	S		
Unk197	30	S							10	S		

	Site ID: Field Sample Number: Sample Date: Depth: Units:	69W-94-10 MXZW10X3 11/02/95 9:5 μg/L	69W-94-10 MXZW10X4 02/14/96 9.5 pg/L	69W-94-11 MDZW11X4 62/14/96 9.5 μg/L	69W-94-11 MXZW11X4 02/14/96 9.5 µg/L	69W-94-12 MXZW12X3 11/02/95 8 μg/L	69W-94-12 MXZW12X4 02/13/96 8 µg/L
Unk198	Current			P5.4		10 \$	
Unk199		90 5					
Jnk200						50 S	5
Unk202							
Unk206							
Unk211			20	S			
Unk212				8			
Unk213							
Unk214							
Unk216							
Unk217							
Unk218							
Unk220							
Unk222		χ.					
Unk224							
Unk225							
Unk226							
Unk227			10	S			
Unk229			5	S		0.0	
Unk236				d			
Jnk237			5	S			

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:	69W-94-13 MXZW13X3 11/02/95 8		69W94-13 MXZW13X4 02/13/96 8		69W-94-14 MXZW14X3 11/03/95 8		69W-94-14 MXZW14X4 02/13/96 8	ZWM-95-15X MXZW15X1 11/01/95	ZWM-95-15X MXZW15X2 82/13/96 8
Units:	µg/L		µg/L		μg/L		₽g/L	µg/L	pg/L
SEMIVOLATILE ORGANICS		00000000							
1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin	10	S	10	S					
1,2,3,4-tetramethylbenzene	6	S							and the second sec
1,2,3,5-tetramethylbenzene / Isodurene	6	S	6	S					
1,2-dimethylnaphthalene	6	S							
1,3,5-trimethylbenzene			6	S					
1,3-dimethyl-5-(1-methylethyl) Benzene			7	S					
1,3-dimethylnaphthalene	20	S	6	S					
1,4,6-trimethyl-naphthalene		0							
1,4-dimethylnaphthalene	10	S		- 11					
1,5-dimethylnaphthalene	10	s		- 1					
1,6,7-trimethylnaphthalene		-							
1,6-dimethylindan			4	S		- 1			
1,6-dimethylnaphthalene			20	S					
1,7-dimethylnaphthalene									
1,8-dimethylnaphthalene			8	s					
1-methyl-3-(1-methylethyl)benzene / M-cymene			7	S S					
1-methylnaphthalene	40	S	40	S					
2,3,6-trimethylnaphthalene			194	-					
2,3-dimethylnaphthalene	6	S	7	S					
2,6,10,14-tetramethylpentadecane				- 1	10	S			
2,6-dimethylnonane						-			
2,7-dimethylnaphthalene			6	s					
2-methyldecane				~		1			
3-methyldecane									
4,7-dimethylindan / 2,3-dihydro-4,7-dimethyl-1h-indene			7	S					
4-(1-methylethyl)toluene / P-cymene			0	-					
Heptadecane				- 1					
Hexadecane				- 1					
Indan / 2,3-dihydro-1h-indene	20	S	10	s					
Nonadecane				~					
Octadecanoic Acid								5	s
Pentadecane									
Tetradecane									
Toluene									
Tridecane									

ù,

	Site ID: Field Sample Number: Sample Date: Depth:	69W-94-13 MXZW13X3 11/02/95 8		69W-94-13 MXZW13X4 02/13/96 8		69W-94-14 MXZW14X3 11/03/95 8	69W-94-14 MXZW14X4 02/13/96	ZWM-95-15X MXZW15X1 11/01/95	ZWM-95-15X MXZW15X2 02/13/96
	Units:	φ μg/L		а µg/L		е µg/L	8 #g/L	8 µg/L	8 μg/l.
Jnk525		1.9	1				C B		19
Jnk543									
Jnk544	8								
Jnk546				6	S				
Jnk547				5	S		1		
Jnk548									
Jnk549				6	S				
Jnk550				1.0					
Jnk551									
Jnk552									
Jnk553		5	S	9	S				- 10 I
Jnk554					~				
Jnk555				9	S				
Jnk556				9 5	S				
Jnk557		20	s	20	s				
Jnk558		20		20					
Jnk559				10	e				
Jnk560			- N	7	S S				
Jnk561		9	s	9	6				
Jnk562		2	0	6	e				
Jnk563		6	s	6	S S S S S				
Jnk564		0	0	7	0				
Jnk565				-L	0				2
Jnk566									
Jnk567		10							
Jnk568		8	S S						
Jnk569		0	9		- 1				
Jnk570							6	s	
Jnk571		6	S	6	S		0	3	
Jnk572		ů.	2	U	2				
Jnk572									
Jnk574									
Jnk575			- 14						
Jnk576		9	s						N. Contraction
Ink577		2	3						
Jnk578				5	s				
nk579	1			5	3				
Jnk580						6 5			
Jnk580 Jnk581						0 3	?		
Jnk581 Jnk582				9					
Jnk 583				5	S S				
				5	3				
Jnk584 Jnk585							-		

Site ID: Field Sample Number: Sample Date:	69W-94-13 MXZW13X3 11/02/95		69W-94-13 MXZW13X4 02/13/96		69W-94-14 MXZW14X3 11/03/95		69W94-14 MXZW14X4 02/13/96		ZWM-95-15X MXZW15X1 11/01/95		ZWM-95-15X MXZW15X2 82/13/96	
Depth: Units:	8 μg/L	1	8 #g/L		8 µg/L		8 #g/L		8 µg/Լ		8 #g/L	
Unk586							Barring Barring					
Unk587		- 1										
Unk588												
Unk589												
Unk590			8	S								
Unk591						- 34						
Unk592	2000	S			600	S				-		
Unk593		- 11					5	S	4.	S		
Unk594		- 1										
Unk595			8	S								
Unk596												
Unk597												
Unk598									1 ÷			
Unk599	9	S								- 20		
Unk600												
Unk601					5	S			1000	S		
Unk602												- 1
Unk603										- 11		
Unk610				- 1					1			
Unk614				- 0					6			
Unk645											1000	
Unk649		_	9	S			20	S		_	10	S
VOLATILE ORGANICS				-						-		
1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin												-
1,2,3,4-tetramethylbenzene				- 1								
1,2;3,5-tetramethylbenzene / Isodurene				- 1		- 1						
1,2,3-trimethylbenzene		- 1										
1,2,4-trimethylbenzene				- 1								
1,3,5-trimethylbenzene				- 1				- 1				
1,3-diethylbenzene				- 1						S-10		- 1
1-methyl-3-(1-methylethyl)benzene / M-cymene				- 1						8.1		
4-(1-methylethyl)toluene / P-cymene	5						-					
C2ace	5		10	S			20	S			50	S
Cumene / Isopropylbenzene												
Indan / 2,3-dihydro-1h-indene		- 11	20	S								
Sec-butylbenzene / (1-methylpropyl)benzene												
Unk162		11										
Unk173												
Unk185												
Unk187												
Unk188												
Unk193												
Unk197												-

1.9

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

1.4

	Site ID: Field Sample Number: Sample Date: Depth:	69W-94-13 MXZW13X3 11/02/95 8	69W-94-13 MXZW13X4 02/13/96 8	69W-94-14 MXZW14X3 11/03/95 8	69W-94-14 MXZW14X4 02/13/96 8	ZWM-95-15X MXZW15X1 11/01/95 8	ZWM-95-15X MXZW15X2 92/13/96 8
Unk198	Units:	µg/L	μg/L	μg/L	μg/L	μg/L	µg/L
Unk199					I AN THE REPORT OF	10.000	
Unk200							
Unk202			8	S			
Unk206			8 7	S S			
Unk211							
Unk212							
Unk213							
Unk214				-			
Unk216			8	S			1
Unk217	7						6 S
Unk218							
Unk220							
Unk222							
Unk224							-) -
Unk225			1 (mar. 1)				
Unk226			100	C			
Unk227			20 20	S	1.0	0	
Unk229			20	S S S			
Unk236			10				
Unk237			8	S			

14

4

Site ID Field Sample Number Sample Date Depth Units	ZWM-95-16X MXZW16X1 11/01/95 11.3 µg/L	ZWM-95-16X MXZW16X2 02/13/96 11.3 µg/L	ZWM-95-17X MXZW17X2 02/12/96 17.2 µg/L	ZWM-95-18X MXZW18X1 11/02/95 8 µg/L	ZWM-95-18X MXZW18X2 92/12/96 8 µg/L
SEMIVOLATILE ORGANICS					
SEMIVOLATILE ORGANICS 1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin 1,2,3,4-tetramethylbenzene / Isodurene 1,2-dimethylnaphthalene 1,3-dimethyl-5-(1-methylethyl) Benzene 1,3-dimethyl-5-(1-methylethyl) Benzene 1,3-dimethylnaphthalene 1,4-6-trimethylnaphthalene 1,4-6-trimethylnaphthalene 1,5-dimethylnaphthalene 1,6-trimethylnaphthalene 1,6-dimethylnaphthalene 1,6-dimethylnaphthalene 1,6-dimethylnaphthalene 1,7-dimethylnaphthalene 1,7-dimethylnaphthalene 1,7-dimethylnaphthalene 2,3,6-trimethylnaphthalene 2,3,6-trimethylnaphthalene 2,3-dimethylnaphthalene 2,6-dimethylnaphthalene 2,6-dimethylnaphthalene 2,7-dimethylnaphthalene 3-methyldecane 4,7-dimethylnaphthalene 1,7-dimethylnaphthalene 2,7-dimethylnaphthalene 3-methyldecane 4,7-dimethylnaphthalene 1,7-dimethylnaphthalene 2,7-dimethylnaphthalene 3-methyldecane 4,7-dimethylnaphthalene 1,7-dimethylnaphthalene 3-methyldecane 4,7-dimethylnaphthalene 1,7-dimethylnaphthalene 1,7-dimethylnaphthalene 1,7-dimethylnaphthalene 3-methyldecane 3-methyldecane 3-methyldecane 3-methyldecane 3-methyldecane 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methylaphthalene 3-methyl	5 S 4 S 6 S 8 S 8 S 8 S				
Toluene					
Tridecane	6 S				

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth: Units:	MXZW16X1 11/01/95	ZWM-95-16X MXZW16X2 02/13/96 11.3 µg/L	ZWM-95-17X MXZW17X2 02/12/96 17.2 µg/L	ZWM-95-18X MXZW18X1 11/02/95 8 µg/L	ZWM-95-18X MXZW18X2 02/12/96 8 µg/L
Unk525		s preve	- Hgyl	4810	pgu .
Unk543	5 ,				
Unk544					
Unk546					
Unk547					
Unk548					
Unk549					
Unk550					
Unk551					
Unk552					
Unk553					
Unk554					
Unk555					
Unk556					
Unk557	5 3	S			
Unk558					
Unk559					
Unk560					
Unk561	4 :	5			
Unk562					
Unk563					
Unk564					
Unk565					
Unk566				0	
Unk567	1				
Unk568	- 4 - 1	S			
Unk569	5	S S			
Unk570		8			
Unk571					
Unk572					
Unk573	1.1.5.1.1.1				
Unk574	4	S			
Unk575		2			
Unk576	4	S			
Unk577					
Unk578					
Unk579					
Unk580					
Unk581					
Unk582					
Unk583					
Unk584					
Unk585					

t.

TABLE 7-1 UNKNOWN COMPOUNDS DETECTED IN GROUNDWATER AOC 69W

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth: Units:	ZWM-95-16X MXZW16X1 11/01/95 11.3 µg/L		ZWM-95-16X MXZW16X2 02/13/96 11.3 µg/L		ZWM-95-17X MXZW17X2 02/12/96 17.3 µg/L		ZWM-95-18X MXZW18X1 11/02/95 8 µg/L	ZWM-95-18X MXZW18X2 92/12/96 8 µg/L	
Unk586	μgru	101010	hāv		aga		490	pgot	
Unk587						- 1			
Unk588									
Unk589									
Unk590	5	s							
Unk591	5								
Unk592									
Unk593									
Unk594									
						- 11			
Unk595									
Unk596									
Unk597						- 14			
Unk598						- 10			
Unk599						- 11			
Unk600									
Unk601	4	S				- 11			
Unk602						- 14			
Unk603									
Unk610									
Unk614							200 S		
Unk645		- 1	4	S		1.1			
Unk649			40	S	10	S		7	S
VOLATILE ORGANICS		-							
1,2,3,4-tetrahydronaphthalene / Tetranap / Tetralin									
1,2,3,4-tetramethylbenzene	20	S							
1,2,3,5-tetramethylbenzene / Isodurene	10	S							
1,2,3-trimethylbenzene									
1,2,4-trimethylbenzene	6	S							
1,3,5-trimethylbenzene	7	S							
1,3-diethylbenzene	6	S							
1-methyl-3-(1-methylethyl)benzene / M-cymene	20	S							
4-(1-methylethyl)toluene / P-cymene	10	S				114			
C2aee	10	S	40	S	7	S		20	S
Cumene / Isopropylbenzene						- 61			
Indan / 2,3-dihydro-1h-indene							X		
Sec-butylbenzene / (1-methylpropyl)benzene	7	S							
Unk162									
Unk173									
Unk185									
Unk187									
Unk188									
Unk193									
Unk197		-		- 1					

4/27/98 12:09 PM

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

1

	Site ID: Field Sample Number: Sample Date: Depth: Units:	ZWM-95-16X MXZW16X1 11/01/95 11.3 µg/L		ZWM-95-16X MXZW16X2 02/13/96 11.3 µg/L	ZWM-95-17X MXZW17X2 62/12/96 17.2 µg/L	ZWM-95-18X MXZW18X1 11/02/95 8 µg/L	ZWM-95-18X MXZW18X2 92/12/96 8 µg/L
Unk198	Unis	pgit				prove.	PB'P
Unk199						11	
Jnk200							
Unk202			- 1				
Unk206			- 1				
Unk211			- U				
Unk212		9	S				
Unk213		20	S				
Jnk214		9 20 20	S				
Jnk216							
Jnk217			1.1				
Jnk218		10	S				
Jnk220		10 10 30	S S				
Unk222		30			4		
Jnk224		10 7	S				
Jnk225		7	S				
Jnk226		8	S				
Jnk227			1.1				
Jnk229		10	S				
Unk236							
Unk237							

TABLE 7-1 UNKNOWN COMPOUNDS DETECTED IN SEDIMENT AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth: Units:	ZWD-95-04X DXZW0400 09/12/95 0 µg/g		ZWD-95-05X DXZW0500 09/12/95 0 µg/g		
SEMIVOLATILE ORGANICS					
Alpha-pinene			8	S	
Beta-pinene			20	S	
Gamma-sitosterol	5	S			
Unk541			8	S	
Unk618		- 12	50	S	
Unk619			20	S	
Unk620		- 1	8	S	
Unk621			30	S	
Unk622			30	S	
Unk652	3	S			
VOLATILE ORGANICS					
Unk187			.03	S	

g:\projects\usaec\projects\69writab\dsceb1.xls

TABLE 7-2 ANALYTES DETECTED IN METHOD BLANKS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

PARAMETER	COMPOUND/ELEMENT
SVOCs in Water Blanks	Diethylphthalate ¹ bis(2-ethylhexyl)phthalate ² Dioctyladipate (TIC) ¹
SVOCs in Soil Blanks	di-n-butylphthalate ¹ nonacosane (TIC) ¹ Dioctyladipate (TIC) ³ Heptacosane (TIC) ³
VOCs in Water Blanks	Acetone ¹ Methylene Chloride ¹ Chloroform ¹ Ethanol (TIC) ²
VOCs in Soil Blanks	Toluene ¹ Acetone ³ Methylene Chloride ³ Trifluoromethane ³
Pesticides in Water Blanks	Malathion ³
Pesticides in Soil Blanks	alpha-chlordane ³ gamma-chlordane ³
Water Quality Parameters	TSS ^{1,2} TDS ^{1,2} Alkalinity ^{1,2}
TPHC in Soil Blanks	TPHC (USEPA Method 9071) ¹

1 = Detected in samples from the 1995 Field Investigations

2 = Detected in samples from the 1995 Round 2 Groundwater Sampling Event

3 = Detected in samples from the 1996 Round 3 Field Investigations

TIC = Tentatively Identified Compound

C:\FDRITABL\69W\Sec7tabs234.wpf

TABLE 7-3 ANALYTES DETECTED IN RINSE BLANKS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

PARAMETER	COMPOUND/ELEMENT				
SVOCs	bis(2-ethylhexyl)phthalate ²				
VOC	1,1,1-Trichloroethane ¹ Methylene Chloride ¹ Acetone ¹				
	Chloroform ¹ Carbon Tetrachloride ¹				
INORGANICS	Mercury ¹ Lead ¹ Iron ¹ Potassium ¹ Manganese ¹				

1 =Detected in samples from the 1995 Field Investigations

2 = Detected in samples from the 1996 Field Investigations

Rinse blank data was not required for the Round 2 Groundwater Sampling Event

TABLE 7-4 ANALYTES DETECTED IN TRIP BLANKS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

PARAMETER	COMPOUND/ELEMENT
VOCs in Water Blanks	Acetone ² Methylene Chloride ^{1,2} Chloroform ¹ Tetrachloroethane ¹ Carbon Tetrachloride ¹ Hexane (TIC) ¹

1 = Detected in samples from the 1995 Field Investigations

2 = Detected in samples from the 1995 Round 2 Groundwater Sampling Event VOCs were not detected in trip blanks from the 1996 Field Investigations TIC = Tentatively Identified Compound

TABLE 7-5 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

SAMPLE POINT	FIELD SCREENING SAMPLE ID	MEDIA	DEPTH (feet below ground surface)	TPH (ppm)
GP-01	69W-94-01-2W	Groundwater	5-7	663
GP-02	69W-94-02-2W	Groundwater	46	354
GP-02(Dup)	69W-94-01-2WD	Groundwater	5-7	208
GP-03	69W-94-03-2W	Groundwater	5-7	110
GP-04	69W-94-04-2W	Groundwater	5-7	12
GP-05	69W-94-05-2W	Groundwater	5-7	126
GP-06	69W-94-06-2W	Groundwater	5-7	188
GP-07	69W-94-07-2W	Groundwater	4-6	3
GP-08	69W-94-08-2W	Groundwater	4-6	95
GP-08(Dup)	69W-94-08-2WD	Groundwater	4-6	75
GP-09	69W-94-09-2W	Groundwater	4-6	2
GP-10	69W-94-10-2W	Groundwater	4-6	25
GP-11	69W-94-11-2W	Groundwater	4-6	37
GP-11(Dup)	69W-94-11-2WD	Groundwater	4-7	2
GP-12	69W-94-12-2W	Groundwater	3-5	20
GP-13	69W-94-13-2W	Groundwater	3-5	460
GP-14	69W-94-14-2W	Groundwater	3-5	2
GP-15	69W-94-15-2W	Groundwater	3-5	438
GP-15(Dup)	69W-94-15-2WD	Groundwater	3-5	110
GP-16	69W-94-16-2W	Groundwater	2-5	7
GP-16(Dup)	69W-94-16-2WD	Groundwater	2-5	4
GP-17	69W-94-17-2W	Groundwater	5-7	1,015
GP-17(Filtered)	69W-94-17F	Groundwater	5-7	22
GP-18	69W-94-18-2W	Groundwater	5-7	19
GP-18(Filtered)	69W-94-18F	Groundwater	5-7	12
GP-19	69W-94-19-2W	Groundwater	5-7	15
GP-19(Filtered)	69W-94-19F	Groundwater	5-7	13
GP-20	69W-94-20-2W	Groundwater	5-7	650
GP-20(Filtered)	69W-94-20F	Groundwater	5-7	15
GP-21	69W-94-21-2W	Groundwater	5-7	119
GP-21(Filtered)	69W-94-21F	Groundwater	5-7	11
GP-22	69W-94-22-2W	Groundwater	5-7	1,087
GP-22(Filtered)	69W-94-22F	Groundwater	5-7	1,007
GP-23	69W-94-23-2W	Groundwater	5-7	45
GP-23(Filtered)	69W-94-23F	Groundwater	5-7	4
GP-24	69W-94-24-2W	Groundwater	5-7	10
GP-24(Filtered)	69W-94-24F	Groundwater	5-7	4
GP-25	69W-94-25-2W	Groundwater	5-7	15
GP-25(Filtered)	69W-94-25F	Groundwater	5-7	5
GP-1	69W-94-01-1	Soil	0-2	123
GP-1(Dup)	69W-94-01-1D	Soil	0-2	120
GP-1	69W-94-01-12	Soil	3-5	226
GP-2	69W-94-02-1	Soil	0-2	1,050
GP-2	69W-94-02-2	Soil	3-5	15,500
GP-3	69W-94-02-2	Soil	0-2	1,150
GP-3	69W-94-03-2	Soil	3-5	387

TABLE 7-5 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

SAMPLE	FIELD SCREENING SAMPLE ID	MEDIA	DEPTH (feet below ground surface)	TPH (ppm)
GP-4	69W-94-04-1	Soil	0-2	1,230
GP-4	69W-94-04-2	Soil	3-5	1,010
GP-5	69W-94-05-1	Soil	0-2	201
GP-5	69W-94-05-2	Soil	3-5	102
GP-6	69W-94-06-1	Soil	0-2	806
GP-6	69W-94-06-2	Soil	3-5	53
GP-7	69W-94-07-1	Soil	0-2	38
GP-7(Dup)	69W-94-07-1D	Soil	0-2	38
GP-7	69W-94-07-2	Soil	3-5	11
GP-8	69W-94-08-1	Soil	0-2	291
GP-8(Dup)	69W-94-08-1D	Soil	0-2	272
GP-8	69W-94-08-2	Soil	3-5	48
GP-9	69W-94-09-1	Soil	0-2	219
GP-9	69W-94-09-2	Soil	3-5	11
GP-10	69W-94-10-1	Soil	0-2	19
GP-10	69W-94-10-2	Soil	3-5	51
GP-11	69W-94-11-1	Soil	0-2	20
GP-11	69W-94-11-2	Soil	3-5	74
GP-12	69W-94-12-1	Soil	0-2	40
GP-12	69W-94-12-2	Soil	3-5	8
GP-13	69W-94-13-1	Soil	0-2	25
GP-13	69W-94-13-2	Soil	3-5	157
GP-14	69W-94-14-1	Soil	0-2	51
GP-14	69W-94-14-2	Soil	3-5	8
GP-15	69W-94-15-1	Soil	0-2	95
GP-15	69W-94-15-2	Soil	3-5	13
GP-16	69W-94-16-1	Soil	0-2	607
GP-16	69W-94-16-2	Soil	3-5	18
GP-17	69W-94-17-1	Soil	3-5	51
GP-17(Dup)	69W-94-17-2D	Soil	3-5	53
GP-18	69W-94-18-2	Soil	3-5	22
GP-19	69W-94-19-2	Soil	3-5	12
GP-20	69W-94-20-2	Soil	3-5	420
GP-21	69W-94-21-2	Soil	3-5	1,497
GP-22	69W-94-22-2	Soil	3-5	14,484
GP-23	69W-94-23-2	Soil	3-5	12
GP-24	69W-94-24-2	Soil	3-5	1,520
GP-25	69W-94-25-2	Soil	3-5	21
69W-94-09	69W-94-09-1	Soil (Boring)	0-2	70
69W-94-09(Dup)	69W-94-09-1D	Soil (Boring)	0-2	71
69W-94-09	69W-94-09-2	Soil (Boring)	2-4	14
69W-94-09	69W-94-09-3	Soil (Boring)	4-6	11
69W-94-09	69W-94-09-4	Soil (Boring)	11-13	12
69W-94-10	69W-94-10-1	Soil (Boring)	0-2	664
69W-94-10	69W-94-10-2	Soil (Boring)	2-4	631
69W-94-10	69W-94-10-3	Soil (Boring)	4-6	7,675

TABLE 7-5 ADL AREE 69W FIELD ANALYTICAL RESULTS AOC 69W

SAMPLE POINT	FIELD SCREENING SAMPLE ID	MEDIA	DEPTH (feet below ground surface)	TPH (ppm)
69W-94-10	69W-94-10-4	Soil (Boring)	11-13	164
69W-94-11	69W-94-11-1	Soil (Boring)	0-2	1,735
69W-94-11	69W-94-11-2	Soil (Boring)	2-4	386
69W-94-11	69W-94-11-3	Soil (Boring)	4-6	125
69W-94-11	69W-94-11-4	Soil (Boring)	11-13	55
69W-94-12	69W-94-12-1	Soil (Boring)	0-2	11
69W-94-12	69W-94-12-2	Soil (Boring)	2-4	30
69W-94-12	69W-94-12-3	Soil (Boring)	4-6	667
69W-94-12	69W-94-12-4	Soil (Boring)	11-13	40
69W-94-13	69W-94-13-1	Soil (Boring)	0-2	1,140
69W-94-13	69W-94-13-2	Soil (Boring)	2-4	19
69W-94-13	69W-94-13-3	Soil (Boring)	4-6	2,900
69W-94-13	69W-94-13-4	Soil (Boring)	11-13	71
69W-94-14	69W-94-14-1	Soil (Boring)	0-2	25
69W-94-14	69W-94-14-2	Soil (Boring)	2-4	83
69W-94-14	69W-94-14-3	Soil (Boring)	4-6	1,990
69W-94-14	69W-94-14-4	Soil (Boring)	11-13	49
HA-1	69W-94-01	Soil	0-1	14
HA-1(Dup)	69W-94-01D	Soil	0-1	53
HA-2	69W-94-02	Soil	0-1	16
HA-3	69W-94-03	Soil	0-1	10
HA-4	69W-94-04	Soil	0-1	56
HA-5	69W-94-05	Soil	0-1	131
HA-6	69W-94-06	Soil	0-1	36

3

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes:

TPH = Total petroleum hydrocarbons ppm = Parts per million by weight Dup = Duplicate sample

TABLE 7-6 ADL, AREE 69W SOIL OFF-SITE ANALYTICAL RESULTS AOC 69W

1.4

Field Sample Fignher Field ID: Sample Date: Depth:		63W-94-81 SXC9W-10 HA-S 63/36.94 S			879-94-01 BDC9W-21 GF-2 6550294			69W-54.82 BXC999.21, GP-2 95/02/94		EX	w 34-03 IC9W 30 GP-3 5502/94			69W-94-04 BXC9W-40 GE-4 05/92/94		49W-94.05 BXC9W-50 GP-13 05502/94		870-54.85 85089 810 870-54.69 86/96/94		59W-94-09 8X039W AL 68W-94-09 9690594		69W-94-10 69W-94-10 99W-94-10		6917 94-18 8X39W-CL 69W-94-18 06/06/94
Linte		46.5		1	MOR		1	2's		1.00/4	10.9			464		Hitt		1 P\$1						Helve
ALMETALS							-				2514									Para	in a second s			
Uumicum	1	9690		1-	7310	D		10600	1		7590	1		7710		10400	1	7260	1	5870	1	8120	1	8440
unionic		12.4			11.5	D		13.7			12.6	_		11.6		13.8		9.08		11.4		10.2	- 1	25.9
ar many		17.4			19.6	D		21.1			19.8			19.7		15.1		20.9		13.2		21.5		13.1
oron		10		<	6.64	D	<	6.64		<	6.64		2	6.64	<		13	6.64	<	5.64	- 1	13.6		11.8
alcium		880			1060	D	11	1060			1510			974		638	- 14	894		792		894		1570
WORDIN		25.2			15.2	D		22.7			14.5			15.6		16.1		15.1		10.6		13.7		24.5
shalt		5.49			5.24	D		5.98			4.73			4.23	<	2.5		5.47	I	3.04	1	3.8		4.16
apper -		13.2			9.1	D		9.07	- 1		7.8			7.61		8.94		10.8		7.78		7.86	1.1	6.28
10		15300			13200	D		23300			14100			12700		15700	- 11	11300		8400		11300		14200
bec		56		1	12	D		13.4			9.41			8,78		12.1		2.95		2.86		7.5		6.08
agnesitum		3940			2400	D		4040			2860			2420		2850		2940		1530	- 1	1910		31.50
avantere		194			169	D		248			186			173		143		150		109		197		181
ickel		15.1			15.3	DI		20.2	1		15.8	1		12.6	D.	15.6	I	11.9		8.41		10.4		14.5
Minerelatory		948			903	D		868	100		950	1.1		803	100	838	100	1600		581		719		745
distra		98,2		P	154	D		170			183			214		106		285		255		139		252
allium	<	34.3		<	34.3	D	<	34.3		<	34.3		<	34.3	<	34.3	×	34.3	<	34.3		< 34,3		45.9
	<	7.43	7	 	7.43	D	<	7.43			11		۰.	7.43	1 4	7.43	<	7.43	<	7.43		< 7.43		15.2
atadium		20.7			12.2	D		18.9			12.6			11.7		16.4		12.9	0.0	6.15		11.6		13.4
nc		32.9	_		261	D		30,4	-	-	23,8			22.6	- 1	25.6		28.9		14.7		27.3		20.3
EMIVOLATILE ORGANICS			_	-	_					_			-		_								_	
monyimphinalene	<	032			1	D	15	.032		<	032		<	032	1	.032	14	.032	<	032		6.4		70
emaphthene	<	041		<	041	D	5	041		<	041		<	.041	<	041	<		<			.56		5
enaphthylene	<	.033		<	033	D	<	033		5	.033			.13	<	033	<	.033	<	033	- 1	< .033		< 3
nzo[a] with scene	12.	.13		100	.12	D		.58		<	041			.81	<	.041	1	.041	<			.47		< .4
nun(b)fluoranihene	<	.31		<	.31	D		.75		<	.31			1.1	<	10.0	<		< C	100		1 L		< 3
nuo(ghi)payime	<	.18		<	.18	D		.51		<	.18			.69	<	.19	1	.18	<			.44		< 2
nzo[k]fluorathene	5	.13		<	.13	D	1.1	.34		<	.13			-43	<	.13	1	.13	<			.36		< 1
mryl Alcohol	<	032			642	D	5	032		<	032		·5	032	<	032		.07		077		< 032		< 3
Kynane	1.	.16		1.00	.14	D	1.00	.53	- 1	<	032			.72	<	032	1	/032	<			.76		< .3
cthyl Philainic	0	.24		<	.24	D	2	.24		<	.24		<	.24	<	.24	×	.24	<			< .24		< 2
uorasbox	1.	.2			-12	D		.9		<	032		0	.7	<	.032	5	.032	<	1.00		.61		< .3
lorene	<	D65		<	065	D	<	065		<	065		<	065	<		×	065	<			1		9
phthalene	<	.74		<	.74	D	<	.74		<	.74		۲.	.74	<	.74		.74	<			3.2		30
enanthrene		.42		1.	.15	D		.95		<	032		1.	.15	<	.032	9	.032	<	1000		2		20
mo)	<	.052		*	052	D	<	.052		<	.052		*	.052	<	.052	-		<			< .0.52	1	× 5
rese	-	.28		4	25	D	1.	1,1	_	2	083		2	1	<	083			<			.91	- 11	< .8
-n-butyi Flutialate	5	1.3	-	14	13	D	1¢	1.1		<	IJ		5	1.3	<	13		35	_	LS		2.6	-	< 10
LATILE ORGANICS	1.2		-	Tat			1.			-	1.0		-		- 1-		1		1.7		-	2 1		
lerobazme	<	.1		<	-4	D	<	1		<	d		<	1	1	1	5	1	<		- 1	< .1		< _1
lorsform	<	.24		5	24	D	<	24		<	.24		5	24	<		1	.24	<			< <u>.24</u>		< .24
hylbenzene	< <	.19		×	19	D	<	19	1	< <	19		V V	.19	<	19	< <	.19	<			< 19		11
luene	15		_	10		U	15	4		2	1	_	5		1<	4	1 <	4	<			5 1		< 1
THER	-	10/00		1	7600	D	1	10000			3030		1	1110	- 1	11100	-	11/0	1<	1000		7000	-	0760
tal Organic Carbon	1	18600			7600 2760	D		10800 2780			2930			4310		15100		1360	<	1000		7800		9750 4350

TABLE 7-6 ADL, AREE 69W SOIL OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Sile ID Field Sample Number Field D: Semple Date:	8X0 694	-94-11 9W-200 1-94-13 56/94		69W-94-11 BDD9W-E1 69W-94-11 96/06/94	6		69W-94-11 BXD9W EL 69W-94-11 569674			4917-94-12 BXD9W.PU 69W-84-12 94-01/94			670-94-12 9XD9W-GL 87W-94-12 95/52/94		8	09W-9413 1209W-9413 87W-9413 964234		8936-94-13 BXD996-41 4699-84-13 56-91/94		8	878-92-14 22399-93-14 299-94-14 299-94-14 299-93-98		•	979-34-14 100997-84-14 8997-94-14 9009398	
Depth: Unit:		1								*			•					*			*			•	8883 S
PALMETALS		iet.		#2'S		00-0000	141/4	0000000	£000000	nd'd	00000000	100000	per	100000 3		HER	0000	and the second sec	0096363	98.06892	Mig	09/10/10	00029992	held	979977380
Alumanom	1 .	200	-	8220	D	-1-	7550		-	8770	_	-	7390	- 1		8790		8810	- 1	-	11500		-	9820	
Arterio		4.4		11.1	D		15.8			9.19		1	28.6			8.21		22.1			14.2	_		36.1	
Barium		16		17.2	D		15.2			14.8			19			19.6		16.5			18.8			17.2	
Boron.		20		9.65	D	<	6.64			9.44			11		×	6.64		9.03			9.86			15.7	
Calcium		030		876	D		971			430			981			1370		1030			518			981	
Chronium		6.1		15.5	D		11.7			13.5			19.6			14.4		32.6			23.6			21.9	
Cobalt		1.85	4	3,85	D		4.79			3.08			4.8		<	2.5		4.03			10.2	1.1		531	
Copper		12.4	11	6.95	D	6	8.81			5.74			4.9	- 11	-	5.01	1	10.4			16.8	1		10.1	1
Ino		7500		14600	D		13500			11200			13100			11200		14500	1.1		28500			182.00	
Lend		1.37		4.03	D		5.92			8.35			6.1			9.7		2.75	1.1		4.97			4.35	
Magnesium		400		3070	D		2510			2250			3370			2160		4140			4530			4450	
Manganese		166		143	D		147			158			174			127		127			640			187	
Nickel		15		13.7	D		14.2			10.3			13.4			10.5		16.4			44.4			17.9	
Potanuim		605		690	D		603			402		1	1280			414		856			427			991	
Sodium		234		224	D		219			125			194			196		246			137	- 1		223	
Thallium		4.3		46.8	D		44.7		×	34.3			40.8		<	34.3		58.6			52.7		<	34.3	
Tin		7.43	1	7.43	D	10	7.43		è	7.43		2	7.43		2	7.43		10.7		2	7.43		è	7.43	
Variadium		3.1	12	11.1	D	1.1	9.13		~	12.4		1	13.8		2	11.5		15			14.5	- 1	~	15.9	
Zinc		2.6		23.7	p		21.6		1.1	22.9		1	26.7			21.1		20.9	_		44			28.3	
SEMIVOLATILE ORGANICS			- 1	-		- 1					-		2017					405		-		-		50.2	
2-methylhaphthalene	1 <	032	1	.67	D	T	.32	-	5	032		13	.032	1	<	.032	1	7.2	-	<	.032	1	e	.032	
Acquiphibene		041	<	.041	D	<	.041		<	.041		1	.041	- 11	<	041	<	041		<	041		<	.041	
Acouptabylene		033	<	033	D	1	.033		R	033		1	033			.17	<	033		<	033		<	.033	
Benzo(a)antimoene		041	6	.041	D	<	.041			12		-c	.041			.64		.13			092		<	041	
Banzo(b)fluorathene		31	<	.31	D	<	.31		8	.31		<	.31			.98	<	.31		<	.31		5	.31	
Benzo striperviene		18	<	.18	D	<	18		5	.18		<	.18			.58	<	.18		<	.18		<	.18	
Benzo kifhaoranthene		13	14	.13	D	<	.13		<	.13		2	.13			.37	<	.13			.13		<	.13	
Benzyl Alcohol		032	1.1	.06	D		.08		<	.032		1	.032		<	.032	<	.032		2	.032		2	.032	
Chrystone		071	<	032	D		067		19	.11		2	032			.63	100	.15			.13		<	032	
Disthyl Phihalate		.24	<	.24	D	<	.24		e.	.24		~	.24		<	.24		3.3		<	.24		<	.24	
Fluorantheor		048	5	032	D	<	032			.13		4	032			.26	<	032	1.1		.11		<	.032	
Fluorenc		065		.15	D	<	.065		×	.065		1	065		<	065		.74		<	.065		<	.065	
Nachthalenc		.74	<	.74	D	<	.74		÷.	.74		~	.74		<	.74		2.4		<	.74		<	.74	
Pheramiltrene		078		.3	D		.23			12		-c	032			.45		1.4	- N		.28	- 11	<	032	
Phenol		052	<	052	D	<	052		<	.052			.17		<	052	<	052		6	052		<	052	
Pyrme		083	<	.083	D	<	.083		1	.19		1	093			.4	<	.083	1.2		.22			.16	
Di-o-buryl Fhilialate		20		21	D		2.7		<	1.3		2	1.3		<	1.3	<	13		<	1.3		<	1.1	
VOLATILE ORGANICS	-							_	2		-	-		-					-	-		-	-		_
Chlorobenzme	<	1	1<	1	D	<	1		<	4		<	.1	1	<	1	1	.18	1		33	1	<	1	_
Chieroform		.24	<	24	D	<	.24		<	.24			A		<	24	<	.24		÷.	24			44	
Ethylbanane		19	<	19	D	<	19		<	19		1.0	.19		<	19		23		<	19	1	<	19	
Tolume	4	1	<	1	D	1 <	4		ic.	4		1	4		<	1	<	1		2	1			.11	
OTHER	2000	2	-			-					-	-	12	-	-		-		-			_			-
Total Organic Carbon	6	850	1	2020	D	1	2450			11600	- 1	1	3040	1		19000	1	4110			5430	1	<	1000	
Total Petroleum Hydrocarbors		865	1	175	D		101			23.4			1390			34.1		2270			123	_		2430	

Notes: <= The concentration was less than the certified reported limit. $\mu g/g = micrograms per gram.$ D = Duplicate analysis. I = The low spike recovery for this lot was high

TABLE 7-7 ADL, AREE 69W GROUNDWATER OFF-SITE ANALYTICAL RESULTS AOC 69W

1.5

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Ske ID: Fleid Sample Number: Fleid ID: Sample Date: Units:	69W-94-02 MXC9W-2X GP-2 05/02/94 µg/L	69W-94-05 MXC9W-5X GP-13 05/02/94 μ±L	59W-94.06 MXC9W-6X GP-1 05/02/94 με/L	69W-94.07 MXC9W-7X GP-15 05/02/94 µg/L	69W-94.08 MXC9W-8X GP-6 05/02/94 μg L	69W-94-09 MDE9W-9X 69W-94-09 09/19/94 μg/L	69W-94.09 MXD9W-9X 69W-94-09 06/24/94 μg/L	69W-94-09 MXE9W-9X 69W-94-09 09/19/94 µg/L	69W-94-10 MXD9W-AX 69W-94-10 06/24/94 µg/L	69W-94-10 MXE9W-AX 69W-94-10 09/20/94 gg/L
PALMETALS										
Aluminum	140000	45700	27600	150000	105000	1740 D	13400	7590	\$5200	4200
Antimony	116	< 60	< 60	132	< 60	< 60 D	< 60	< , 60	< 60	< 60
Artenic	< 2.35	< 2.35	320	37.8	170	14.5 D	19.5	20,6	470	350
Barkan	605	122	143	374	268	40.5 D	76.2	\$7.7	347	53.3
Berytlium	7.22	1.6	< 1.12	5,34	5.09	< 1.12 D	< 1.12	1.39	4.41	< 1.12
Cadmium	< 6.78	< 6.78	< 6.78	< 6.78	< 6.78	< 6.78 D	< 6.78	< 6.78	11.4	< 6.78
Calcium	99700	24200	63700	41000	21100	19400 D	20000	19900	34700	21100
Chroenium	347	131	83.8	513	312	< 16.8 D	17	< 16.8	192	17.6
Cobalt	114	28.3	28,7	86.5	82.3	< 25 D	< 25	< 25	83.3	< 25
Copper	207	63.8	47.1	230	211	< 18.8 D	< 18.8	< 18.8	165	< 18.8
ron	225000	58400	137000	177000	172000	868 D	11700	7630	148000	35100
Lead	201	12.8	52.1	73.5	152	< 4.47 D	7.18	5.59	102	6.06
Magnesium	38900	12800	11000	48500	35500	1920 D	3990	2142	24900	3410
Mangarese	6460	2850	9880	3630	3550	158 D	265	248	13000	3950
Mercury	.219	* 1	< 1	.295	< .1	< 1 D	< .1	< .1	.112	< .1
Nickel	318	93.6	67.7	333	278	< 321 DH	< 32.1	< 32.1 H	234	< 32.1 H
Potantium	24700	5260	9710	20300	14800	1750 D	5460	3540	14700	5240
Selenium	≤ 2.53									
Sodium	2.55		< 2.53 40700						4.05	
Variadiam		31800		46600	45400	26900 D	28500	27600	32700	23700
	195	43.9	44,1	191	137	< 27.6 D	and a state	< 27.6	129	< 27.6
Line	483	1 132	1 129	1 432	1 344	1 × 18 D	34	22.6	245	20
SEMIVOLATILE ORGANICS			-1	1	1	1	1		1	
2-methylmaphthalene	1000	< 13	600	< 1.3	400	< 1.3 D	< 1.3	< 1.3	700	400
Accouptitione	< 60	< 5.8	< 60	< 5.8	< 60	< 5.8 D	< 5.8	< 5.8	34	< 30
Fluorene	< 90	< 9.2	< 90	< 9.2	< 90	< 9.2 D	< 9.2	< 9.2	35	< 50
Vaphthalene	600	11	300	< .5	500	< .5 D	< .5	< .5	500	300
	< 100	< 9.9	< 100	< 9.9	< 100	< 9.9 D	< 9,9	< 99	64	< 50
VOLATILE ORGANICS	-									
	< 1	30	20	< 1	< 1	< 1 D	2.6 B	< 1	< 1	< 1
1,1-dichloroethane	< 1	< 1	10	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
1,1-dichloroethene	< 1	S 1	20	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Acetone	< 8	< 8	< 80	37	< 8	< 8 D	< 8	< 8	< 8	< *
Baizene	< 1	< 1	9	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Carbon Tetrachloride	< 1	< 1	10	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Chlorobenzme	< 1	< 1	10	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Chloreform	< 1	< 1	10	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Chloromethane	< 1.2	< 1.2	20	< 12	< 12	< 1.2 D	< 1.3	< 12	< 1.2	< 1.2
Ethylbenzese	80	< 1	40	< 1	< 1	< 1 D	< 1	< 1	52	57
Tetrachloroethene	< 1	< 1	10	< 1	< 1	< 1 D	< 1	< 1	< 1	< 1
Tolucne	< 1	< 1	10	< 1	< T	< 1 D	< i	< 1	1 1	1× 1
Trichloroethylene	< 1	e 1	10	1 1	i i	I I D	3.2	1	i i	la i
	< 1	< i	20	< 1	1 1	< 1 D	< 1	< 1	i i	< 1
Cylenes	< 20	6 2	< 20	e 2	12 1	S 2 D	1 2	4 1		24
WET CHEMISTRY	**									
Alkalinity		1	1	T	1	1	21000	1	97400	1
Chloride			-			34000	36000	34000	30000	22000
Nitrite, Nitrate-non Specific	14.6	22.9	20	3100	22.3	2100 D	3800	1800	210	310
	3200		5000							
Nitrogen By Kjeldahl Method		764	7.304	1260	2200	< 64 D N	88	0.000	N 810	
Phosphorum	1800	1900	457	5200	1400	51.2 D	-	174	1300	330
Sulfite						19000	23000	20000	4000	3660
Total Hardness						55400 D	64900	61200	183000	65300
Total Surperided Solids					1	344000 D	290000	348000	1620000	174000
OTHER										

Page 1

1.1.1

TABLE 7-7 ADL, AREE 69W GROUNDWATER OFF-SITE ANALYTICAL RESULTS AOC 69W

.

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

She ID; Fleid Sample Number: Fleid ID; Sample Date: Units;		69W-9411 MDD9W-BX 69W-94-11 06/23/94 μg/L			69W-94-11 MXD9W-BX 69W-94-11 06/23/94 μg/L		69W-94-11 MXE9W-BX 69W-94-11 09/20/94 μg/L			69W-94-12 MXD9W-CX 69W-94-12 06/23/94 µg/L		69W-94-12 MXE9W-CX 69W-94-12 09/19/94 µg/L			69W94-13 MXD9W-DX 69W-94-13 96/23/94 µg/L		69W-94-13 MXE9W-DX 69W-94-13 09/19/94 pg/L		1	69W-94-14 МХD9W-EX 69W-94-14 06/23/94 µg/L		69W-94-14 MXE9W-EX 69W-94-14 09/19/94 µg/L
AL METALS	_								-					-								
luminum	1.	2790	D		2610	1111	447		1.5	11300	1100	7290		100	537		4610		1000	18100		8330
timory	<	60	D	<	60	<	60		<	50	<	60		<	60	<	60		<	60	<	60
senic	154	38	D	1.	37.8		62.1			25.6	1.1.1.1.1	36			60.4		400		1.1	70.4		270
rium		34.4	D	1	32		221			56.5		54.4			25.6		42.9			76		65.3
ayilium	<	1.12	D	e	1.12	<	1.12		<	1.12		1.96		<	1.12		1.4		<	1.12	<	1.12
edminum.	<	6.78	D	<	6.78	<	6.78		<	6.78	1 <	6.78		<	6.78	<	6.78		<	6.78	<	6.78
licium	1.0	27900	D	1.5	26200		26000		100	24400	11	18700		16.5	19200		20200		100	18900		18500
hromium	<	168	D	1 <	168	1 <	16.B			22.2		20.5		<	16.8	<	16.8			39.5	1	16.8
obalt	è	25	D	<	25	<	25		<	25	<	25		1	25	1	25			28.8	12	25
	2	18.8	D	12	18.8	4	18.8		2	18.8	14	18.8		i e	18.8	<	18.8			27.8	2	18.8
opper	~		-	1		1.			1		1.			1		1						
9R		5430	D	1.	5150	1.5	2470			12100		9160		1	4730		28400			25200		23600
ad the	<	4.47	D	<	4.47	<	4.47		1	8 92		9.65		<	4.47		11			23.3		15.2
sgresium		3630	D		3440		2920		1	4680		3740			2040		3770			5280		2850
anganese	1.1	3340	D		3140	1.5	2330		1.	395		409		12	3910		2920			1850	1.0	994
forcury	<	1	D	<	.1	<	.1		<	.1	<	.1		<	.1	<	1			108	<	.1
ickel	<	32.1	D	<	32.1	<	32.1	н	<	32.1	<	32,1	H	<	32.1	<	32.1	H		49.8	<	32.1
of an a statement of the statement of th	1.0	3270	D		3110	<	1240		- C	4250	10.0	4360			2430	1.1	1780		1.0	3460	100	2230
deniam		2.53	D	<	2.53	<	2.53		<	2.53	<	2.53		<	2.53	<	2.53		<	2.53	<	2.53
odium	10.0	37500	D	1.1	35500	1.1	34900		1	47300		34300			16700		28900			28000 -		38300
madium	<	27.6	D	<	27.6	<	27.6		<	27.6	<	27.6		<	27.6	<	27.6		<	27.6	<	27.6
inc	1.1	32.5	D	100	45.5	1.1	29.1		1.5	38.5		29.1		1	19	ing. 224	22.6		1	50.8	1 N.	20
MIVOLATILE ORGANICS	-	360		-	10,0		***: C.			303	_	Sec.16						-	-			
nethy inspititualene	-	37	D	1	42	- 1	7.2		1 <	1.3	1 <	13		1	69	T	29		<	6	<	1.3
comphthese	<	5.8	D	<	58	<	5.8		2	5.8	1	5.8		<	5.8	<	5.8		<	30	1	3.8
uorene	i c	9.2	D	1	92	<	9.2		1	9.2	~	9.2		<	9.2	1	9.2			50	1	9.2
	1.2			1.		1.			1.000	5	1	5		1	64	1			è	2	2	.5
aphibulane	1.1	88	D	1.	11		27		×		1 e	9.9			9.9	<	47		i a	50	- lè	9.9
housthree	15	9,9	D	<	9.9	<	9.9		15	9.9	15	2.9		<	49	15	3.3	-	1.	30	15	9.9
OLATILE ORGANICS		1.4	DB	1.2		1.			1	3.1 B	-	1.7		-	4.1 B	Te	1		<		1 <	1
1,1-trichloroethane		1.5		1		1			1.00	3.1 B	1.2	1.1		1	4.1 0	1	1		2		2	
1-dichloroethane	*	1	D	<		1	1		<	1	*			1.00	1		1			1	1.5	
1-dichloroethese	<	1	D	<		<	4		1	1	<	1		<	4	-	1		<	1	1	
celone	<	8	D	<		<	8		~	8	1	8		<	8	<	8		<	8	<	8
attale	<	4	D	<	1	1 <	1		<	1	1	3.		<	1	<	1		<	- 4	<	1
abon Tetrachloride	<	1	D	1	1	<	1		<	1.	<	1		<	1	1	1		<	1	<	1
hlorobenzene	<	1	D	<	1	5	1		1	1	<	1		4	1	5	1		<	1	<	1
hloroform	<	1	D	<	1	<	1		<	1	<	1		<	1	<	1		<	1	<	1
loromethane	<	1.2	D	<	1.2	1 4	1.2		<	1.2	<	1.2		× .	1.2	5	1.2		<	1.2	<	1.2
hylbenzene		1	D	<	1	4	1		5	1	<	1		1.1	19		9.1		<	1		1
trachloroethene	<	1	D	<	1	1 0	1		<	1	1 <	1		×	1	<	1		<	1	<	1
lume		1	D	<	1	e	1		a.	1	<	1		*	1	1	1.		<	1	<	1
ichloroethylene	~	î	D	<	1	1	1		1	3.9	<	1		×	1	<	1		<	1		1
schlorofluoromethane	4	1	D	<		1.2	i		<	1				*	i.	1	1		<	i.		1
denes	2		D	1		2			6	2	1	2		2		14	2		<	2	i e	-
ET CHEMISTRY			0	1.		13			1.2		12	-		1.		1.*	•				1-	
calinity	-	52600	D	1	56000				-	21200	1				43300	1			-	31,500		
loride		58000	D	1	58000		48000			81000		41000			35000		40000			40000		52000
			D									2000			20.7		162		1	182		1600
trite, Nitrate-non Specific		500			350		1100	1.1		3100												
		627	D		681		145	N		330		203	N		635		1100	N		423		456
		77	D		73.9		45.2			225		400			30,9		560			620		740
osphorus			D		17000		13000		1	18000		19000			2520		1850			17000		15000
itrogen By Kjeldahl Method nosphorus ilfate		17000																				
oosphorus Ifate stal Hardness		83000	D		78000		75400			78600		60600			55300		64600		-	67200		56700
osphorus lfate														-			64600 2190000					56700 545000

Notes: < = The concentration was less than the certified reported limit. $\mu g/g = micrograms per gram.$ D = Duplicate analysis. I = The low spike recovery for this lot was high. H = Lot out of control but data accepted due to high recoveries. V = Sample was not correctly preserved. N = The high spike recovery for this low was low. B = Analyte found in method black or QC sample.

TAB7-7 XLS

4727/10

TABLE 7-8 ADL, AREE 69W SURFACE WATER OFF-SITE ANALYTICAL RESULTS AOC 69W

Site ID: Field Sample Number: Sample Number: Units:	69W-94-15 WXD9W-FX 06/24/94 μg/L	69W-94-16 WXD9W-GX 06/24/94 µg/L
PAL METALS		
Aluminum	658	< 112
Arsenic	2.92	< 2.35
Barium	17.5	13.1
Calcium	26600	23900
Iron	1470	276
Lead	7.65	< 4.47
Magnesium	4480	3750
Manganese	164	38.8
Potassium	1920	1920
Sodium	42700	40000
Zinc	24.6	< 18
VOLATILE ORGANICS		
1,1.1-trichloroethane	3.7 B	2.2 B
WET CHEMISTRY	Contraction of the second	
Alkalinity	< 10000	42300
Chloride	77000	75000
Nitrite, Nitrate-non Specific	930	1100
Nitrogen By Kjeldahl Method	353	214
Phosphorus	91.2	24.1
Sulfate	14000	15000
Total Hardness	83000	73700

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes:

.

 $\mu g/L = micrograms per liter$

< + Concentration is less than the certified reporting limit.

B = Analyte found in method blank or QC sample.

TABLE 7-9 ADL, AREE 69W SEDIMENT OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth: Units:	69W94-15 DXD9W-FO 06/24/94 0 µg/g		69W-94-16 DDD9W-GC 06/24/94 0 µg/g	90000000000000000000000000000000000000		69W-94-16 DXD9W-GO 06/24/94 0 Hg/g	
PAL METALS			200000000 Feb 7-6 7-70000000			F.0. P	19220
Aluminum	. 6280		6130	D		7920	-
Arsenic	6.46		13.2	D		20.3	
Barium	15.4		11.2	D		19.8	
Boron	< 6.64		12.9	D	<	6.64	
Calcium	1200		2660	D	1	1650	
Chromium	18.2		14.6	D		23.6	
Cobalt	5.5	<	2.5	D		5.79	
Copper	13.1		43.2	D		14.2	
Iron	15100		14400	D		16200	
Lead	20.7		20.7	D		23.6	
Magnesium	3080		3990	D		4470	
Manganese	195		259	D		203	
Nickel	13.4		12.8	D		18.1	
Potassium	842		581	D		979	
Sodium	104		79	D		132	
Vanadium	12.2		11.5	D		15.3	
Zinc	49.5		41.6	D		48.1	
SEMIVOLATILE ORGANICS							
Acenaphthylene	.34		.098	D	<	.033	
Benzo[a]anthracene	1.9		.3	D	10.0	.16	
Benzo[b]fluoranthene	2.2	<	.31	D	<	.31	
Benzo[ghi]perylene	.95	<	.18	D	<	.18	
Benzo[k]fluoranthene	.99	<	.13	D	<	.13	
Benzyl Alcohol	< .032		.073	D	<	.032	
Chrysene	2		.36	D		.19	
Fluoranthene	2.5		.44	D		.21	
Fluorene	.55	<	.065	D	<	.065	
Phenanthrene	3.8		.59	D		.13	
Pyrene	3.8		.5	D		.27	
Bis(2-ethylhexyl) Phthalate	< .48	<	.48	D		1.4	
OTHER							
Total Organic Carbon	2580		1320	D		1670	
Total Petroleum Hydrocarbons	188		123	D	1	131	_

Notes:

µg/g = micrograms per gram.

< = Concentration is less than the certified reporting limit,

D = Duplicate analysis.

1

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWB-96-03X 2 4-Sep-96 134	ZWB-96-03X 4 4-Sep-96 144	ZWB-96-03X 4(Dup) 4-Sep-96 144	ZWB-96-03X 6 4-Sep-96 141	ZWB-96-03X 6(Dup) 4-Sep-96 148	ZWB-96-03X 8 4-Sep-96 161	ZWB-96-03X 10 4-Sep-96 151
Vinyl Chloride	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
1,1-DCE	5 µg/kg/250µg/kg	270 UJ	290 UJ	290 UJ	280 UJ	300 UJ	320 UJ	300 UJ
t-1,2-DCE	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
c-1,2-DCE	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Chloroform	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
1,1,1-TCA	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Carbon Tetrachloride		270 U	290 U	290 U	280 U	300 U	320 U	300 U
Trichloroethene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Tetrachloroethene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
1,3-DCB	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
1,4-DCB	2 µg/kg/250µg/kg	270 U	290 U	290 U	.280 U	300 U	320 U	300 U
1,2-DCB	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Benzene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Toluene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Chlorobenzene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Ethylbenzene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
m/p-Xylene	4 µg/kg/500µg/kg	540 U	580 U	580 U	560 U	590 U	640 U	600 U
o-Xylene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
Naphthalene	2 µg/kg/250µg/kg	270 U	290 U	290 U	280 U	300 U	320 U	300 U
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-ir (95)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPHC (96)	50 mg/kg	54 U	58 U	58 U	56 U	59 U	64 U	60 U

Notes:

J

- U = Concentrations Less than Reporting Limit
 - = Value is Estimated

NA = Not Analyzed

E = Concentration Exceeds the Maximum Reporting Limit

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWB-96-03X 10(Dup) 4-Sep-96 149	ZWB-96-03X 12 4-Sep-96 146	ZWM-96-19X 1 27-Aug-96 131	ZWM-96-19X 1(Dup) 27-Ang-96 131	ZWM-96-19X 3 3-Sep-96 130	ZWM-96-19X 5 28-Aug-96 133	ZWM-96-19X 7 3-Sep-96 140
Vinyl Chloride	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
1,1-DCE	5 µg/kg/250µg/kg	300 UJ	290 UJ	260 U	260 U	260 U	270 U	280 U
t-1,2-DCE	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
c-1,2-DCE	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Chloroform	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
1,1,1-TCA	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Trichloroethene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Tetrachloroethene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
1,3-DCB	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
1,4-DCB	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
1,2-DCB	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Benzene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Toluene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Chlorobenzene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Ethylbenzene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
m/p-Xylene	4 µg/kg/500µg/kg	600 U	580 U	520 U	520 U	520 U	530 U	560 U
o-Xylene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	260 U	270 U	280 U
Naphthalene	2 µg/kg/250µg/kg	300 U	290 U	260 U	260 U	680	270 U	280 U
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-ir (95)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPHC (96)	50 mg/kg	NA	58 U	54 U	53	150	52 U	56

1.1

	Site ID: Depth (ft bgs): Date analyzed: Dilution:	ZWM-96-19X 9 3-Sep-96 143	ZWM-96-19X 11 10-Sep-96 154	ZWM-96-19X 13 3-Sep-96 156	ZWM-96-20X 2 3-Sep-96 154	ZWM-96-20X 4 3-Sep-96 155	ZWM-96-20X 6 3-Sep-96 158	ZWM-96-20X 8 3-Sep-96 159
ANALYTES	Reporting Limit		22.0					
Vinyl Chloride	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
1,1-DCE	5 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
t-1,2-DCE	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
c-1,2-DCE	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Chloroform	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
1,1,1-TCA	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Trichloroethene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Tetrachloroethene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
1,3-DCB	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
1,4-DCB	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
1,2-DCB	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Benzene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Toluene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Chlorobenzene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Ethylbenzene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
m/p-Xylene	4 µg/kg/500µg/kg	580	620 U	620 U	620 U	620 U	630 U	640 U
o-Xylene	2 µg/kg/250µg/kg	290 U	310 U	310 U	310 U	310 U	320 U	320 U
Naphthalene	2 µg/kg/250µg/kg	3800 E	4500	1400	310 U	310 U	320 U	320 U
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-ir (95)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPHC (96)	50 mg/kg	840	790	63	62	62 U	63. U	64 U

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWM-96-21X 0 3-Sep-96 134	ZWM-96-21X 2 4-Sep-96 134	ZWM-96-21X 4 5-Sep-96 138	ZWM-96-21X 6 5-Sep-96 136	ZWM-96-21X 8 5-Sep-96 143	ZWM-96-21X 10 4-Sep-96 143	ZWM-96-21X 12 5-Sep-96 154
Vinyl Chloride	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
1,1-DCE	5 µg/kg/250µg/kg	270 U	270 UJ	280 UJ	270 UJ	290 UJ	290 UJ	310 UJ
t-1,2-DCE	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
c-1,2-DCE	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Chloroform	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
1,1,1-TCA	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Trichloroethene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Tetrachloroethene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
1,3-DCB	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
1,4-DCB	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
1,2-DCB	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Benzene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Toluene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Chlorobenzene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Ethylbenzene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
m/p-Xylene	4 µg/kg/500µg/kg	540 U	540 U	550 U	540 U	570 U	570 U	620 U
o-Xylene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
Naphthalene	2 µg/kg/250µg/kg	270 U	270 U	280 U	270 U	290 U	290 U	310 U
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	· NA	NA
TPH-gro	100 µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-ir (95)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPHC (96)	50 mg/kg	54 U	54 U	55 U	54 U	57	57 U	62 U

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWE-95-01X 3 6-Oct-95 1.1	ZWE-95-01X 4 6-Oct-95 1.07	ZWE-95-01X 5 6-Oct-95 1.08	ZWR-96-26X 4 19-Sep-95 1.04	ZWR-95-26X 7 20-Sep-95 1.15	ZWR-95-26X 10 20-Sep-95 1.3	ZWR-95-27X 4 19-Sep-95 1.06
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.5 U	5.4 U	5.4 U	5.2 U	5.8 U	6.5 U	5.3 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2,3 U	2.6 U	2.1 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 UJ	2.3 U	2.6 U	2.1 UJ
Trichloroethene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2,3 U	2.6 U	2.1 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
Toluene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
Chlorobenzene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
Ethylbenzene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3 U	2.6 U	2.1 U
m/p-Xylene	4 µg/kg/500µg/kg	4.4 U	4.3 U	4.3 U	4.2 UJ	4.6 UJ	5.2 UJ	4.2 UJ
o-Xylene	2 µg/kg/250µg/kg	2.2 U	2.1 U	2.2 U	2.1 U	2.3	2.6 U	2.1 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	710	NA	NA
TPH-gro	100 µg/kg	55 U	110 U	110 U	NA	45000 E	130 U	NA
TPH-ir (95)	50 mg/kg	82	67	270	130	2100	65 U	53 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

~

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-27X 7 18-Sep-95 1.09	ZWR-95-27X 10 21-Sep-95 1.3	ZWR-95-28X 4 23-Sep-95 275	ZWR-95-28X 6 23-Sep-95 550	ZWR-95-28X 10 23-Sep-95 320	ZWR-95-30X 2 19-Sep-95 1.1	ZWR-95-30X 6 20-Sep-95 5.7
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.5 U	6.5 U	1400 UJ	2800 UJ	1600 UJ	5.5 U	29 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.2.U	2.6 U	550 U	1100 U	640 U	NA	11 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	11 U
Carbon Tetrachloride		2.2 UJ	2.6 U	550 U	1100 U	640 U	2.2 UJ	11 U
Trichloroethene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	11 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	11 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	11 U
Toluene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	26
Chlorobenzene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	640 U	2.2 U	31
Ethylbenzene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	2200	2.2 U	260 E
m/p-Xylene	4 µg/kg/500µg/kg	4.4 UJ	5.2 UJ	1100 UJ	2200 UJ	3200 J	4.4 UJ	300 J
o-Xylene	2 µg/kg/250µg/kg	2.2 U	2.6 U	550 U	1100 U	1200	2.2 U	6200 E
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	110 U	3400
TPH-gro	100 µg/kg	NA	130 U	600000 E	1.1 e+6E	3.5 e+6E	NA	540000 E
TPH-ir (95)	50 mg/kg	55 U	65 U	7500	4500	3100	55	7700
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

...

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-30X 10 20-Sep-95 1.15	ZWR-95-31X 4 20-Sep-95 1.05	ZWR-95-31X 7 20-Sep-95 139	ZWR-95-31X 10 20-Sep-95 1.24	ZWR-95-32X 4 19-Sep-95 1.05	ZWR-95-32X 7 20-Sep-95 1.18	ZWR-95-32X 10 20-Sep-95 1.18
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.8 U	5.3 U	700 U	6.2 U	5,3 U	5.9 U	5.9 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	NA	2.1 U	280 U	2.5 U	NA	2.4 U	2.4 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.3 U	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.4	2.1 U	280 U	2.5 U	2.1 UJ	- 2.4 U	2.4 U
Trichloroethene	2 µg/kg/250µg/kg	2.3 U	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.3 U	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.3 U	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
Toluene	2 µg/kg/250µg/kg	NA	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
Chlorobenzene	2 µg/kg/250µg/kg	2.3 U	2.3 J	280 U	2.5 UJ	2.1 U	2.4 U	2.4 UJ
Ethylbenzene	2 µg/kg/250µg/kg	2.3 U	2.1 U	280 U	2.5 U	2.1 U	2.4 U	2.4 U
m/p-Xylene	4 µg/kg/500µg/kg	4.6 UJ	4.2 UJ	560 UJ	5.0 UJ	4.2 UJ	4.7 UJ	4.7 UJ
o-Xylene	2 µg/kg/250µg/kg	2.3 U	2.1 U	550	2.5 U	2.1 U	2.4 U	2.4 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	1400	5200	120 U	NA	120 U	120 U
TPH-gro	100 µg/kg	120 U	32000 E	220000 E	1100	NA	120 U	120 U
TPH-ir (95)	50 mg/kg	58 U	3700	4900	110 J	68 J	59 U	88 J
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-33X 4 20-Sep-95 1.09	ZWR-95-33X 7 20-Sep-95 1,3	ZWR-95-33X 10 20-Sep-95 1.35	ZWR-95-34X 4 20-Sep-95 1.17	ZWR-95-34X 7 21-Sep-95 1.14	ZWR-95-34X 10 20-Sep-95 1.18	ZWR-95-35X 4 21-Sep-95 1,11
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.5 U	6.5 U	6.8 U	5.9 U	5.7 U	5.9 U	5.6 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	3.8	NA	NA	2.3 U	2.3 U	2.4 U	NA
1,1,1-TCA	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Trichloroethene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Toluene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Chlorobenzene	2 µg/kg/250µg/kg	2.2 UJ	2.6 UJ	2.7 UJ	2.3 UJ	2.3 U	2.4 U	2.2 U
Ethylbenzene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
m/p-Xylene	4 µg/kg/500µg/kg	4.4 UJ	5.2 UJ	5.4 UJ	4.7 UJ	4.6 UJ	4.7 UJ	4.4 UJ
o-Xylene	2 µg/kg/250µg/kg	2.2 U	2.6 U	2.7 U	2.3 U	2.3 U	2.4 U	2.2 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	110 U	130 U	140 U	120 U	110 U	120 U	110 U
TPH-ir (95)	50 mg/kg	66	160	45	59 U	57	59 U	55
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

. . .

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-35X 6 19-Sep-95 1.18	ZWR-95-36X 4 20-Sep-95 1.06	ZWR-95-36X 7 24-Sep-95 2.28	ZWR-95-36X 10 20-Sep-95 1,19	ZWR-95-37X 4 23-Sep-95 145	ZWR-95-37X 6 20-Sep-95 1.22	ZWR-95-37X 10 18-Sep-95 1.16
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.9 U	5.3 U	11 U	6.0 U	730 UJ	6.1 U	5.8 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	NA	290 U	NA	2.3 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
Carbon Tetrachloride		2.4 UJ	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 UJ
Trichloroethene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 UJ	2.4 U	290 U	2.4 U	2.3 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
Toluene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
Chlorobenzene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 UJ	2.4 UJ	290 U	2.4 UJ	2.3 U
Ethylbenzene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
m/p-Xylene	4 µg/kg/500µg/kg	4.7 UJ	4.2 UJ	9.1 UJ	4.8 UJ	580 UJ	4.9 UJ	4.6 UJ
o-Xylene	2 µg/kg/250µg/kg	2.4 U	2.1 U	4.6 U	2.4 U	290 U	2.4 U	2.3 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	120 U	540	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	NA	14000 E	22000 E	120 U	530000 E	120 U	NA
TPH-ir (95)	50 mg/kg	59 U	1200	1100	53	1800	61 U	58 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-38X 3 24-Sep-95 2.44	ZWR-95-38X 6 24-Sep-95 1.18	ZWR-95-38X 10 18-Sep-95 1.22	ZWR-95-39X 4 23-Sep-95 1.14	ZWR-95-39X 6 20-Sep-95 1,19	ZWR-95-39X 10 21-Sep-95 1.3	ZWR-95-40X 0 24-Sep-95 1.09
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	12 U	5.9 U	6.1 U	5.7 UJ	6.0 U	6.5 U	5.5 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	NA	2.2 U
1,1,1-TCA	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	2.6 U	2.2 U
Carbon Tetrachloride		4.9 U	2.4 U	2.4 UJ	2.3 U	2.4 U	2.6 U	2.2 U
Trichloroethene	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.5	2.3 U	2.4 U	2.6 U	2.2 U
Tetrachloroethene	2 µg/kg/250µg/kg	4.9 UJ	2.4 UJ	2.4 U	2.3 U	2.4 U	2.6 U	2.2 UJ
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	2.6 U	2.2 U
Toluene	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	2.6 U	2.2 U
Chlorobenzene	2 µg/kg/250µg/kg	4.9 UJ	2.4 UJ	2.4 U	2.3 U	2.4 UJ	2.6 U	2.2 UJ
Ethylbenzene	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	2.6 U	2.2 U
m/p-Xylene	4 µg/kg/500µg/kg	9.8 UJ	4.7 UJ	4.9 UJ	4.6 UJ	4.8 UJ	5.2 UJ	4.4 UJ
o-Xylene	2 µg/kg/250µg/kg	4.9 U	2.4 U	2.4 U	2.3 U	2.4 U	2.6 U	2.2 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	120 U	NA	NA	· NA	NA
TPH-gro	100 µg/kg	240 U	120 U	NA	110 U	120 U	130 U	110 U
TPH-ir (95)	50 mg/kg	120 U	59 U	61 U	57 U	60 U	64	55 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-40X 3 24-Sep-95 1.11	ZWR-95-40X 6 24-Sep-95 1.15	ZWR-95-40X 10 19-Sep-95 1.13	ZWR-95-41X 3 23-Sep-95 1.23	ZWR-95-41X 6 18-Sep-95 1.23	ZWR-95-41X 10 19-Sep-95 1.19	ZWR-95-42X 3 23-Sep-95 1.22
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.6 U	5.8 U	5.7 U	6.2 UJ	6.2 U	6.0 U	6.1 UJ
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.2 U	2.3 U	NA	2.5 U	2.5 U	2.4 U	2.4 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 UJ	2.5 U	2.5 UJ	2.4 UJ	2.4 U
Trichloroethene	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.2 UJ	2.3 UJ	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
Toluene	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4.U	2.4 U
Chlorobenzene	2 µg/kg/250µg/kg	2.2 UJ	2.3 UJ	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
Ethylbenzene	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
m/p-Xylene	4 µg/kg/500µg/kg	4.4 UJ	4.6 UJ	4.5 UJ	4.9 UJ	4.9 UJ	4.8 UJ	4.9 UJ
o-Xylene	2 µg/kg/250µg/kg	2.2 U	2.3 U	2.3 U	2.5 U	2.5 U	2.4 U	2.4 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	120 U	NA
TPH-gro	100 µg/kg	110 U	120 U	NA	120 U	NA	NA	120 U
TPH-ir (95)	50 mg/kg	56 U	73 J	57 J	62 U	62 U	60 U	61 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-42X 6 24-Sep-95 1.21	ZWR-95-42X 10 19-Sep-95 1.21	ZWR-95-43X 3 18-Sep-95 1.04	ZWR-95-43X 6 18-Sep-95 1.25	ZWR-95-43X 10 24-Sep-95 1.2	ZWR-95-44X 0 25-Sep-95 1.04	ZWR-95-44X 4 18-Sep-95 1.1
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	6.1 U	6.1 U	5.2 U	6.3 U	6.0 U	5.2 U	5.5 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.4 U	2.4 UJ	2.1 UJ	2.5 UJ	2.4 U	2.1 U	2.2 UJ
Trichloroethene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.4 UJ	2.4 U	2.1 U	2.5 U	2.4 UJ	2.1 U	2.2 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2 U
Toluene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.6	2.2 U
Chlorobenzene	2 µg/kg/250µg/kg	2.4 UJ	2.4 U	2.1 U	2.5 U	2.4 UJ	2.2	2.2 U
Ethylbenzene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2.U
m/p-Xylene	4 µg/kg/500µg/kg	4.8 UJ	4.8 UJ	4.2 UJ	5.0 UJ	4.8 UJ	4.2 U	4.4 UJ
o-Xylene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.1 U	2.5 U	2.4 U	2.1 U	2.2 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	120 U	100 U	NA	NA	NA	NA
TPH-gro	100 µg/kg	120 U	NA	NA	NA	120 U	100 U	NA
TPH-ir (95)	50 mg/kg	61 U	61 U	52 U	63 U	53	86	55 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

100

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-44X 6 18-Sep-95 1.18	ZWR-95-44X 10 18-Sep-95 1.22	ZWR-95-45X 0 24-Sep-95 1.31	ZWR-95-45X 1 25-Sep-95 1,17	ZWR-95-45X 4 25-Sep-95 1.22	ZWR-95-45X 6 18-Sep-95 1.19	ZWR-95-45X 10 18-Sep-95 1.19
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.9 U	6.1 U	6.6 U	5.9 U	6.1 U	6.0 U	6.0 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.3 U	2.4 U	2.4 U	2.4 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.3 U	2.4 U	2.4 U	2.4 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.4 UJ	2.4 UJ	2.6 U	2.3 U	2.4 U	- 2.4 UJ	2.4 UJ
Trichloroethene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.3 U	2.4 U	2.4 U	2.4 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.3 U	2.4 U	2.4 U	2.4 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.3 U	2.4 U	2.4 U	2.4 U
Toluene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	2.4	2.4 U	2.4 U	2.4 U
Chlorobenzene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	4.1	2.4 U	2.4 U	2.4 U
Ethylbenzene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	12	2.4 U	2.4 U	2.4 U
m/p-Xylene	4 µg/kg/500µg/kg	4.7 UJ	4.9 UJ	5.2 UJ	13	4.9 U	4.8 UJ	4.8 UJ
o-Xylene	2 µg/kg/250µg/kg	2.4 U	2.4 U	2.6 U	12	2.4 U	2.4 U	2.4 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	2400	120 U	NA	NA
TPH-gro	100 µg/kg	NA	NA	130 U	83000 E	11000 E	NA	NA
TPH-ir (95)	50 mg/kg	59 U	61 U	66 U	3400	61 U	59 J	59 J
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

1000

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-46X 0 23-Sep-95 1.04	ZWR-95-46X 4 19-Sep-95 1.12	ZWR-95-46X 6 19-Sep-95 1.18	ZWR-95-46X 10 20-Sep-95 1.14	ZWR-95-47X 0 18-Sep-95 1.04	ZWR-95-47X 4 25-Sep-95 1.15	ZWR-95-47X 6 18-Sep-95 1.18
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.2 UJ	5.6 U	5.9 U	5.7 U	5.2 U	5.8 U	5.9 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2,1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.5	2.4 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.1 U	2.2 UJ	2.4 UJ	2.3 U	2.1 UJ	2.3 U	2.4 UJ
Trichloroethene	2 µg/kg/250µg/kg	2.1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
Toluene	2 µg/kg/250µg/kg	2.1 U	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
Chlorobenzene	2 µg/kg/250µg/kg	6.7	2.2 U	2.4 U	2.3 UJ	2.1 U	2.3 U	2.4 U
Ethylbenzene	2 µg/kg/250µg/kg	5.6	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
m/p-Xylene	4 µg/kg/500µg/kg	7.2 J	4.5 UJ	4.7 UJ	4.6 UJ	4.2 UJ	4.6 U	4.7 UJ
o-Xylene	2 µg/kg/250µg/kg	11	2.2 U	2.4 U	2.3 U	2.1 U	2.3 U	2.4 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	100 U	NA	NA	110 U	NA	120 U	NA
TPH-ir (95)	50 mg/kg	460	56 U	59 U	57 U	98	58 U	59 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

1

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-47X 10 18-Sep-95 1.18	ZWR-95-48X 0 25-Sep-95 1.41	ZWR-95-48X 1 18-Sep-95 1.15	ZWR-95-48X 3 25-Sep-95 1.2	ZWR-95-48X 8 28-Sep-95 1.28	ZWR-95-49X 0 28-Sep-95 1.32	ZWR-95-49X 2 25-Sep-95 1.23
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	5.9 U	7.1 U	5.8 U	6.0 U	6.4 U	6.6 U	6.2 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.4 UJ	2.8 U	2.3 UJ	2.4 U	2.6 U	2.6 U	2.5 U
Trichloroethene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
Toluene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
Chlorobenzene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
Ethylbenzene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2.3 U	2.4 U	2.6 U	2.6 U	2.5 U
m/p-Xylene	4 µg/kg/500µg/kg	4.7 UJ	5.6 U	4.6 UJ	4.8 U	5.1 U	5.3 U	4.9 U
o-Xylene	2 µg/kg/250µg/kg	2.4 U	2.8 U	2,3 U	2.4 U	2.6 U	2.6 U	2.5 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/kg	NA	. 140 U	NA	120 U	130 U	130 U	2500 E
TPH-ir (95)	50 mg/kg	59 U	71 U	58 U	130	64 U	140	62 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-49X 4 28-Sep-95 1.25	ZWR-95-49X 10 28-Sep-95 1.21	ZWR-95-50X 4 25-Sep-95 1.1	ZWR-95-50X 6 28-Sep-95 1.12	ZWR-95-50X 10 28-Sep-95 1.22	ZWR-95-51X 4 28-Sep-95 1.03	ZWR-95-51X 10 28-Sep-95 1.02
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	6.3 U	6.1 U	5.5 U	5.6 U	6.1 U	5.2 U	5.1 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Trichloroethene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Toluene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 UJ	2.1 UJ	2.0 UJ
Chlorobenzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Ethylbenzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
m/p-Xylene	4 µg/kg/500µg/kg	5.0 U	4,8 U	4.4 U	4.5 U	4.9 U	4.1 U	4.1 U
o-Xylene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.2 U	2.2 U	2.4 U	2.1 U	2.0 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	130 U	NA	110 U	NA	NA	· NA	NA
TPH-gro	100 µg/kg	130 U	120 U	110 U	110 U	120 U	100 U	100 U
TPH-ir (95)	50 mg/kg	63 U	61 U	160	56 U	61 U	52 U	79
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-51X 18 28-Sep-95 1.25	ZWR-95-52X 4 21-Sep-95 1.19	ZWR-95-52X 6 21-Sep-95 1.19	ZWR-95-52X 10 21-Sep-95 1.22	ZWR-95-53X 3 21-Sep-95 1.22	ZWR-95-53X 7 21-Sep-95 1.23	ZWR-95-53X 10 21-Sep-95 1.3
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	6.3 U	6.0 U	6.0 U	6.1 U	6.1 U	6.2 U	6.5 U
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.5 U	NA	2.4 U	2.4 U	NA	2.5 U	NA
1,1,1-TCA	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Trichloroethene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Toluene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5.U	2.6 U
Chlorobenzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Ethylbenzene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
m/p-Xylene	4 µg/kg/500µg/kg	5.0 U	4.8 UJ	4.8 UJ	4.9 UJ	4.9 UJ	4.9 UJ	5.2 UJ
o-Xylene	2 µg/kg/250µg/kg	2.5 U	2.4 U	2.4 U	2.4 U	2.4 U	2.5 U	2.6 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	120 U	NA	NA	NA
TPH-gro	100 µg/kg	130 U	120 U	120 U	120 U	120 U	120 U	130 U
TPH-ir (95)	50 mg/kg	63 U	59	60 U	61 U	54	62 U	65 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA	NA	NA	NA

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

ANALYTES	Site ID: Depth (ft bgs): Date analyzed: Dilution: Reporting Limit	ZWR-95-54X 6 11-Oct-95 1.22	ZWR-95-54X 8 11-Oct-95 1.26	ZWR-95-55X 6 11-Oct-95 1.1	ZWR-95-55X 8 11-Oct-95 1.13
Vinyl Chloride	2 µg/kg/250µg/kg	NA	NA	NA	NA
1,1-DCE	5 µg/kg/250µg/kg	6.1 UJ	6.3 UJ	5.5 UJ	5.7 UJ
t-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA
c-1,2-DCE	2 µg/kg/250µg/kg	NA	NA	NA	NA
Chloroform	2 µg/kg/250µg/kg	2.4 U	2.5 UJ	2.2 U	2.3 U
1,1,1-TCA	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Carbon Tetrachloride	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Trichloroethene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Tetrachloroethene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
1,3-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA
1,4-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA
1,2-DCB	2 µg/kg/250µg/kg	NA	NA	NA	NA
Benzene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Toluene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Chlorobenzene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Ethylbenzene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
m/p-Xylene	4 µg/kg/500µg/kg	4.9 U	5.0 U	4.4 U	4.5 U
o-Xylene	2 µg/kg/250µg/kg	2.4 U	2.5 U	2.2 U	2.3 U
Naphthalene	2 µg/kg/250µg/kg	NA	NA	NA	NA
TPH-dro	100 mg/kg	NA	NA	NA	NA
TPH-gro	100 µg/kg	120 U	130 U	110 U	110 U
TPH-ir (95)	50 mg/kg	61 U	63 U	55 U	57 U
TPHC (96)	50 mg/kg	NA	NA	NA	NA

1.11

Site ID: Field Sample Number: Lab Sample Number: Sample Date: Depth: Units:	DEVENS BACKGROUND CONCENTRATION (44/2)	*	ZWB-95-01X BDZW0100 DV4S*435 09/19/95 0 (4g/g)			ZWB-95-01X BXZW0100 DV4S*246 09/19/95 0 (µg/g)	ZWB-95-01X BXZW0107 DV45*247 09/19/95 7 (µg/g)	2WB-95-02X BX2W0200 DV4S*248 09/19/95 0 (µg/g)	ZWB-95-02X BX2W0205 DV45*250 09/19/95 5 (ws/g)	ZWB-96-03X BXZW0306 DV4S*500 08/23/96 6 (4g/g)
PAL METALS	048/87		UAR/R	-		04870	1 04/0	1 048/10	1 049/80	04990
Aluminum	18,000	-	5690	D	-	6140	13700	7180	6140	t
Arsenic	19	100	10.6	D	11	14	28.2	11	16	
Barium	54		17	D		19	32.9	18.4	14.1	
Beryllnum	0.81	<	.5	D		.565	< .5	< .5	.568	
alcium	810	-	600	D		803	1410	478	522	
Iromium	33		14.1	D		16.4	28.8	15.7	16.3	
Cobalt	4.7		4.81	D		4.08	45	4.98	4.09	
	13.5		7.47	D		9.43	18.9	8.4	10.6	
Copper	18,000		10300	D		8070	16300	10700	10600	
ron .cad	48		7.53	D		6.91	14.9	21.7	8.57	
	48 5,500		1990	D		2150	4060	21.7 2130	2520	
Magnesium	380			D		152	367	2130	117	
Manganese Mercury	380	<	228	D	<	.05	< .05	< .05	< .05	
Vickel	142	<	14	D	<	13.3	33.3	13.7		
	14.6								161	
otassium	2,400	1.5	684	D		872	1430	483	724	
Selenium	101	<	.25	D	<	.25	< .25	< .25	< .25	
odium	131	3	305	D	3	374	1470	323	312	
anadium	32.2		10	D		10.7	20.5	12.9	12.6	
Zinc	43.9		18,8	D		20,8	48.5	26.1	21.6	
SEMIVOLATILE ORGANICS		-						1		
2-methylnaphthalene		<	.2	D	<	.2	80	< ,2	< :049	
Acenaphthene		<	.2	D	<	.2	8	< .2	< .036	
Acenaphthylene		<	.2	D	<	2	< .2	< .2	. < .033	
Anthracene		<	.2	D	<	2	< .2	< .2	< .033	
Benzo[k]fluoranthene		<	.3	D	<	3	< .3	< .3	< .066	
Juysene		<	.6	D	<	.6	< .6	< .6	< .12	
luoranthene		<	.3	D	<	3	< .3	< .3	< .068	
luorene		<	.2	D	<	2	9	< .2	.37	
Japhthalene		<	.2	D	<	2	40	< 2	< .037	
Phenanthrene		<	.2	D	<	2	10	< .2	.56	
Pyrene		<	.2	D	<	2	2	< .2	.12	
Bis(2-ethylhexyl) Phthalate		<	3	D	<	3	< 3	< 3	< .62	
OLATILE ORGANICS				-				1		
1,2,2-tetrachloroethane		<	.0024	D	<	.0024	< 0024	< .0024	.009	
,1,2,2-tetrachioroethane		~	.017	D	×	.017	< .017	< .017	< .017	
Dichloromethane		< C	017	D	×	.012	< .017	< .017	< .012 '	
Ethylbenzene		~ ~	.0012	D		.0017	.012	< 0012	< .012	
Styrene		< <	.0017	D	< <	.0017	< .0026	< 0017	< .0017	
oluene		<	0026	D	5	.0026	< 00026	< 00026	< .0026	
richlorofluoromethane		-		D				< .00078	< .00078	
vienes		<	0063 0015	D	<	014 0015	.022	< 0015	<	
			a subscription			1000	1		1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	
OTHER otal Organic Carbon		-			1		1	1	1	
GUI CIERING CRIDON					1					

Site ID: Field Sample Number: Lab Sample Number: Sample Date: Depth: Units:	DEVENS BACKGROUND CONCENTRATION (P\$/8)	ZWM-95-15X BXZW1504 DV4S*265 09/19/95 4 (ug/g)	ZWM-95-16X BXZW1607 DV45*266 09/20/95 7 (μg/g)	ZWM-95-18X BXZW1804 DV45*268 10/02/95 4 (pg/g)	ZWM-96-19X BXZW1909 DV4S*505 08/21/96 9 (µg/g)	ZWM-96-20X BXZW2004 D¥45*507 98/22/96 4 (µg/g)	ZWM-96-21X BXZW2108 DV45*509 08/26/96 8 (##/#)
PAL METALS	04000	(HAR)		0.04280			046/10
Aluminum	18,000		1			1	
Arsenic	19		· · · · · · · · · · · · · · · · · · ·				
Barium	54						
Beryllium	0.81						
Calcium	810						
Thromium	33						
Cobalt	47						
Copper	13.5						
ron	18,000						
,ead	48						
Aagnesium	5,500						
Aagnesium Manganese	380						
Aanganese Aercury	380						
Viercury Vickel	14.6					1	
otassium	2,400						
elenium	2,400						
odium	131						
anadium	32,2						
	43.9						
Linc	43,9				the second se		
SEMIVOLATTLE ORGANICS							
2-methylnaphthalene Acenaphthene Acenaphthylene Anthracene Benzo[k]fluoranthene Chrysene Fluoranthene							
Fluorene Naphthalene Phenanthrene Syrene Bis(2-ethylhexyl) Phthalate							
VOLATILE ORGANICS							10 million
1,1,2,2-tetrachloroethane Acetone Dichloromethane Ethylbenzene Styrene Toluene							
Frichlorofluoromethane Kylenes			0				
THER		02.10		1	1 0140	1 (00	1100
otal Organic Carbon		8740	671	758	2140	622 < 27.8	1180 < 27.8
otal Petroleum Hydrocarbons					1740	< 27.8	\$ 27.8

Stte D: Field Sample Number: Lab Sample Number: Sample Date: Depth:	DEVENS BACKGROUND CONCENTRATION	ZWR-95-26X RXZW2607 DV45*245 09/11/95 0	2WR-95-30X RX2W3006 DV45*249 09/11/95 0	ZWR-95-35X RXZW3504 DV4S*254 09/12/95 0	ZWR-95-36X RXZW3607 DV48*255 09/13/95 0	ZWR-95-37X RXZW3704 DV4S*256 09/13/95 0	ZWR-95-38X RXZW3803 DV45*257 09/14/95 3
Units: PAL METALS	(µg/g)	(ug/g)	(µg/g)	(µg/g)	(µg/g)	(HK/g)	(ug/g)
	18 000	2010	5760	6270	1 4440	1 12100	1 3740
Aluminum Arsenic Barium Beryllium Calcium Calcium Chromium Cobalt Copper orn Lead Magnesium Manganese Magnesium Manganese Mercury Nickel Potassium Selenium Sodium Yanadium	18,000 19 54 0.81 810 33 4.7 13.5 18,000 48 5,500 380 14.6 2,400 131 32.2 43.9	$\begin{array}{c} 2910 \\ 4.74 \\ 8.21 \\ < 5 \\ .369 \\ < 4.05 \\ 2.88 \\ 4.6 \\ 5880 \\ 1.87 \\ 1090 \\ 56.4 \\ < 05 \\ 8.57 \\ 515 \\ < 25 \\ 291 \\ 4.5 \\ 14 \end{array}$	5760 14 13 < 5 #44 19.1 3.61 7.26 7440 3.93 2550 92.5 < 05 19.2 735 < 25 336 11.7 16.9	$ \begin{array}{r} $	4440 17 12.4 5 660 11.8 3.52 12.6 10200 5.48 2230 236 <	12100 23 13,6 < 5 \$100 15,4 6,43 41;4 25000 88 5450 430 < 05 26,7 370 < 25 325 18,8 35,9	$\begin{array}{c ccccccccccccccccccccccccccccccccccc$
EMIVOLATILE ORGANICS		44					1
2-methylnaphthalene Acenaphthylene Acenaphthylene Anthracene Benzolklifluoranthene Chrysene Thuoranthene Fluorene Vaphthalene Phenanthrene Pyrene Bis(2-ethylhexyl) Phthalate		$\begin{array}{c ccccccccccccccccccccccccccccccccccc$	$\begin{array}{c c} & 9 \\ 1.5 \\ < & 033 \\ .16 \\ < & 066 \\ < & 12 \\ < & 068 \\ 2.4 \\ .18 \\ 4 \\ < & 033 \\ < & 62 \end{array}$.049 .036 .033 .066 .12 .068 .033 .033 .033 .033 .033 .033 .033 	 < 049 < 036 < 033 < 066 < 12 < 068 < 033 < 033 < 033 < 033 < 3.1 	 < .049 .2 < .033 .053 < .066 < .12 < .068 .53 < .037 .89 .096 < .62 	 < 049 < 036 < 003 < 003 < 066 < 12 < 068 < 003 < 003 < 003 < 003 < 62
i,1,2,2-tetrachloroethane Acetone Dichloromethane Eduylbenzene Styrene Foluene Frichlorofluoromethane Xylenes		< 0024 .022 < .012 < .0017 < .0026 < .00078 < .0059 < .0015	< 0024 .024 < .012 < .0017 .0034 .0044 < .0059 .0023	< 0024 < 017 < 012 < 00017 < 00026 < 00078 < 00059 < 0015	< .0024 033 < .012 < .0017 < .0026 < .00078 < .0059 < .0015	< .0024 .03 < .012 < .0017 < .0026 .0024 < .0059 < .0015	< .0024 < .017 < .012 < .0017 < .0026 < .00078 < .0059 < .0015
OTHER Fotal Organic Carbon Fotal Petroleum Hydrocarbons		902	3240	< 27.8	566	1400 ,	34,4

116

Site ID: Field Sample Number: Lab Sample Number: Sample Data: Depth: Units:	DEVENS BACKGROUND CONCENTRATION (48/g)	ZWR-95-39X RXZW3904 DV45*258 09/13/95 0 (µg/g)	ZWR-95-45X RXZW4504 DV45*264 09/15/95 4 (##/g)	ZWS-95-35X SXZW3500 DV45*251 09/18/95 0 (µ#/#)	ZWS-95-37X SXZW3700 DV4S*252 09/18/95 0	ZWS-95-37X SXZW3700 DV45*401 09/28/95 0	ZWS-95-38X SXZW3800 DV4S*253 09/18/95 0
PAL METALS	(48%)	Qi gi gi	1 (49/8)	VER	(µg/g)	(µg/g)	(µg/g)
Auminum	18,000	9570	3060	5940	4520		5930
Annimum Ansenic Barium Beryllium Calcium Chromium Cobalt Copper ren ead Manganese Manganese Mercury Nickel Potassium Selenium Selenium Sedium Vanadium Zine	18,000 19 54 0,81 810 33 4.7 13.5 18,000 48 5,500 380 14.6 2,400 131 32.2 43.9	26 26 23.1 < 5 1260 27.7 6.13 10.3 17700 6.79 3600 239 < .05 27.8 1260 < .25 3399 20.5 38.4	$ \begin{array}{r} 30660 \\ 7.32 \\ 8.14 \\ < .5 \\ 463 \\ 10.3 \\ 2.22 \\ 5.14 \\ 5460 \\ 1.91 \\ 1430 \\ 90.3 \\ < .05 \\ 8.26 \\ 460 \\ < .25 \\ 358 \\ 6.47 \\ < 8.03 \\ \end{array} $	3940 8.6 22.4 <	4520 10.9 9.97 .757 548 14.3 2.16 4.42 6590 15 1650 134 < .05 8.45 384 < .25 513 10.1 16.4	4	5930 8,94 14.3 .62 546 11.4 2.17 6.87 5680 45.7 943 99.8 < .05 6.4 230 368 358 13.6 21.1
EMIVOLATILE ORGANICS				1	1		
2-methylnaphthalene Acenaphthene Acenaphthylene Anthraceme Benzo[k]fluoranthene Drysene Fluoranthene Vaphthalene Vaphthalene Yrene Sis(2-ethylhexyl) Phthalate		< 049 < 036 < 033 < 066 < .12 < 068 < 033 < 033 < 033 < 033 < .033 < .046 < .056 < .056 < .057 < .057	 < .049 < .036 < .033 < .066 < .12 < .068 < .033 < .037 < .033 < .033 < .62 	< 1 < .7 < .7 < .7 < 1 < 2 < 1 < .7 < .7 < .7 < .7 < .7 < .7 < .7 < .7 < .7 < .10	< 049 < 036 < 033 < 033 < 066 < 12 089 < 033 < 037 068 078 < .62		 2 2 2 2 2 2 3 6 7 2 2 4 6 3
OLATILE ORGANICS			1				
, 1, 2, 2-tetrachloroethane Acetone Dichloromethane Zihylbenzene Styrene Foluene Frichlorofluoromethane Kylenes		< 0024 < 017 < 012 < .0017 < .0026 < .00078 < .0059 < .0015	< 0024 < 017 025 < 0017 < 0026 0013 < 0059 < 0015	< .0024 < .017 < .012 < .0017 < .0026 .0021 .0072 .0027	< .0024 < .017 < .012 < .0017 < .0026 .00096 < .0059 < .0015		< .0024 < .017 < .012 < .0017 < .0026 .0015 < .0059 < .0015
OTHER			-				-
otal Organic Carbon Iotal Petroleum Hydrocarbons		< 27.5	< 27.5	936	160		222

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Coits:	DEVENS BACKGROUND CONCENTRATION (µg/g)	SXZW3900 DV4S*259 09/18/95 0 (µg/g)	SXZW4200 DV4S*260 09/18/95 0 (µg/g)	SXZW4500 DV4S*261 09/18/95 0 (µg/g)	SXZW4600 DV4S*263 09/18/95 0 (\4g/g)	ZWS-95-47X SXZW4700 DV45*262 09/18/95 0 (ug/g)
PAL METALS	18,000	6150	6110	6160	5210	5930
Ansenie Barium Baryllium Zaleium Chromium Cobalt Copper fron Lead	19 54 0.81 810 33 4.7 13.5 18,000 48	15 15.6 5 771 15.4 5.36 9.25 10300 22	14.5 20.8 85 821 28.1 5.11 29.9 9390 238	7.66 14.1 < 5 470 12.1 2.51 5.59 6780 11.4	18 16.3 < .5 796 17.5 4.81 10.5 10200 31.6	8.47 20 < .5 333 17.2 3.82 7.54 8260 18.1
Magnesium Manganese Mercury Nickel Potassium Selenium Sodium Vunadium Zinc	5,500 380 14.6 2,400 131 32.2 43.9	2670 230 05 181 500 <	2670 240 0735 14.4 436 < 25 506 19.1 73.7	1360 52.4 0384 5.98 367 <	2620 204 < 05 117 601 < 25 24 12.9 27.2	2510 122 < 05 11.5 784 364 256 12.5 18.9
EMIVOLATILE ORGANICS						
2-methylnaphthalene Acenaphthylene Acenaphthylene Anthracene Benzo[k]fluoranthene Chrysene Thuoranthene Fluorene Naphthalene Phenanthrene Pyrene Bia(2-ethylhexyl) Phthalate		< 1 < .7 2 1 2 5 5 1 < .7 7 8 < 10	 < 1 < .7 < .7 < .7 < 1 < 2 < .7 < .7	 < 049 < 036 < 033 < 033 < 066 < .12 < 068 < 033 < 037 .065 .075 < .62 	 < 5 < 4 < 3 < 7 < 10 9 < 3 < 4 9 10 < 60 	< 049 < 036 < 033 < 033 < 066
1,2,2-4trachloroethane Acetone Dichloromethane Ethylbenzene Styrene Foluene Trichlorofluoromethane Xylenes		< 0024 < 017 < 012 < 0017 < 0026 < 00078 < 0059 < 0015	< 0024 069 < 012 < 0017 < 0026 0011 < 0059 < 0015	< .0024 < .017 < 012 < 0017 < .0026 < .00078 < .0059 < .0015	< .0024 < .017 < .012 < .0017 < .0026 < .00078 < .0059 < .0015	< .0024 < .017 < .012 < .0017 < .0026 .001 .0055 < .0015
OTHER						
Fotal Organic Carbon Fotal Petroleum Hydrocarbons		310	378	< 27.5	652	52.5

Notes:

 $\mu g/g = micrograms per gram$ < = Concentrations are less than the certified reporting limit<math>D = Duplicate analysisShading indicates exceedance of established Devens background concentrations

TABLE 7-12 EMI - VMI SUBSURFACE SOLL OFF-SITE ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sampie Number: Sampie Date:		ZWM-96-19X BXZW1905 8/21/96		ZWM-96-19X BXZW1909 8/21/96		ZWM-96-20X BXZW2002 08/22/96		ZWM-96-20X BXZW2004 8/22/96		ZWB-96-03X BDZW0306 8/23/96		ZWB-96-03X EXZW0306 8/23/96		ZWB-96-03X EXZW0310 08/23/96	1	ZWM-96-21X BXZW2104 8/26/96		ZWM-96-21X BXZW2108 08/25/96
Depth (Post BGS):		5		9		2		4	L	6	10	6	1	10	h.	4		8
EPU (mg/kg) ALEPBATIC	<	0.15	1	150	<	0.15	<	710	<	0.15	<	0.17	<	0.17	<	017	<	0.16
n-C9 to p-C18	<	1.5		560	<	1.5	<	1.7	<	1.5	<	1.7	<	1.7	<	1.7	<	1.6
n-C19 to n-C36	<	0.15		110	<	0.15	<	0.17	<	0.15	<	0.17	<	0.17	<		<	0.16
AROMATICS	1.00				1				1.1							0.000		1. A A A A A A A A A A A A A A A A A A A
n-C 10 to n-C 22	<	30	1	120	<	30	<	34	<	30	<	34	<	34	<	34	<	32
TARGETED PAR ANALYTES (mg/t	(8)									-								
Naphthalene	1<	05	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
2-Methy Insphthalene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Acenaphthylene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Acenaphthene	<	05	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Fluorene	<	05	1	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Phenanthrene	<	05	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Anthracene	<	05	•	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Fluoranthene	<	0 5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Pyrene	<	05	•	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Benzo(s)anthracene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.33
Chyrsepe	<	05	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Benzo(b)fluoranthene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Benzo(k)fluoranthene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Benzo(a)pyrene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Indeno(1,2,3-c,d)pyrene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Dibenzo(a,h)anthracene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
Benzo(g,h,i)perylene	<	0.5	<	0.5	<	0.51	<	0.57	<	0.51	<	0.57	<	0.57	<	0.56	<	0.53
VPB (me/kg)	<	13		4100	<	13	<	13	<	13	<	13	<	13	<	13	<	13
ALIPHATICS					1								0				1	
n-C510n-C8	<	130		270	<	130	<	130	<	130	<	130	<	130	<	130	<	130
n-C910n-C12	<	13		8300	<	13	<	13	<	13	<	13	<	13	<	13	<	13
AROMATICS																		
D-C910 D-C10	<	250		3500	<	250	<	250	<	250	<	250	<	250	<	250	<	250
TARGETED VOAs (ug/kg)					1.1		1											
Methyl tert-butyl Ether	<	250	<	250	<	250	<	250	<	250	<	250	<	250	<	250	<	250
Benzene	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50
Tohuene	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50
Edaylbenzene	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50
m/p-Xylene	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	181. C	<	50
o-Xylene	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50	<	50
Naphthalene	<	250	1	250	1	250	10	250	1	250	10	250	<	250	1	- 250	<	250

Notes: < = Less than certified reporting limits

69CSOHIT.wk1

	MCP SI/GW1	69W-HS-SSW1	69W-HS-SSW-2	69W-HS-OB-15	69W-HS-WSW	69W-HS-FL-2	69W-HS-FL-
DEPTH in feet (bgs)	Standards	7-10	7-10	6	7-10	11	11
EPH PARAMETERS	Cristians Co						
Aliphatics/Aromatics (µg/g)							
C9-C18 Aliphatics	1,000	3,600	5,400	1,800	480	10,000	560
C19-C36 Aliphatics	2,500	360	670	310	91	1,200	67
C11-C22 Aromatics	200	1,100	1,200	700	44	2,300	110
EPH	400	1,100	1,200	100	610	2,500	110
Target PAH Analytes (ug/g)					010		
2-Methylnaphthalene	4	26	42	18	<0.56	81	2.5
Acenaphthene	20	7.6	7.4	4.1	<0.56	16	1.4
Acenaphthylene	100	16	<0.064	9.6	<0.56	<0.076	<0.61
Anthracene	1,000	<0.068	< 0.064	0.24	<0.56	E2.2	<0.61
Benzo (a) anthracene	0.7	<0.068	0.099	<0.07	<0.56	0.096	<0.61
Benzo (b) fluoranthene	0.7	< 0.068	0.064	<0.07	<0.56	<0.076	<0.61
Benzo (k) fluoranthene	7	< 0.068	<0.064	<0.07	<0.56	<0.076	<0.61
Benzo (a) pyrene	0.7	<0.068	< 0.064	<0.07	<0.56	<0.076	<0.61
Benzo (g,h.i) perylene	1,000	< 0.068	<0.064	<0.07	<0.56	<0.076	<0.61
Chrysene	7	J0.039	0.076	<0.07	<0.56	0.089	<0.61
Dibenz (a,h) anthracene	0.7	<0.068	<0.064	<0.07	<0.56	<0.076	<0.61
Fluoranthene	1,000	0.24	<0.064	0.13	<0.56	E1.8	<0.61
Flourene	400	20	26	E0.68	<0.56	44	2.8
Indeno (1,2,3-cd) pyrene	0.7	<0.068	<0.064	<0.07	<0.56	<0.076	<0.61
Napthalene	-4	7.1	8.5	12	<0.56	17	<0.61
Phenanthrene	700	7	9	4.6	<0.56	14	1
Pyrene	700	E0.44	<0.064	0.18	<0.56	E.5	<0.61
VPH PARAMETERS							
Aliphatics/Aromatics (ug/g)			[
C5-C8 Aliphatics	100	<96	<120	<9.8	<11	<110	<13
C-9-C12 Aliphatics	1,000	770	670	21	14	1300	140
C9-C10 Aromatics	100	650	560	20	15	960	100
VPH					30		
Target VPH Analytes (µg/g)							
Benzene	10	<12	<14	<1.2	<1.3	<13	<1.6
Ethylbenzene	80	<12	<14	<1.2	<1.3	. <13	<1.6
m,p-Xylenes	500	<48	<58	<4.9	<5.4	<54	<6.4
Methyl-tert-butyl ether	0.3	<36	<43	<3.6	<4.0	<40	<4.8
Naphthalene	4	<24	<29	2.6	<2.7	<27	<3.2
o-Xylene	500	<24	<29	<2.5	<2.7	<27	<3.2
Toluene	90	<36	<43	<3.6	<4.0	<40	<4.8

the second second	MCP SI/GWI	69W-HS-SE-FL	69W-HS-ESW-1	5. 1. 10 M. 1	69W-HS-WSW-1 DUP	69W-UST-ESW	69W-UST-FI
DEPTH in feet (bgs)	Standards	11	7-10	7-10	7-10	7-10	12-13
EPH PARAMETERS							
Aliphatics/Aromatics (µg/g)							
C9-C18 Aliphatics	1,000	13	<3.4	69	75	11	97
C19-C36 Aliphatics	2,500	10	<4.6	25	23	8.3	23
C11-C22 Aromatics	200	<11	<9.7	26	33	12	48
EPH			1	120	130		
Target PAH Analytes (ug/g)							
2-Methylnaphthalene	4	<0.63	<0.57	< 0.58	<0.57	<0.62	<0.60
Acenaphthene	20	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Acenaphthylene	100	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Anthracene	1,000	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Benzo (a) anthracene	0.7	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Benzo (b) fluoranthene	0.7	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Benzo (k) fluoranthene	7	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Benzo (a) pyrene	0.7	<0.63	<0.57	< 0.58	<0.57	<0.62	<0.60
Benzo (g,h.i) perylene	1,000	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Chrysene	7	<0,63	<0.57	<0.58	<0.57	<0.62	<0.60
Dibenz (a,h) anthracene	0.7	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Fluoranthene	1,090	<0,63	<0.57	<0.58	<0.57	<0.62	<0.60
Flourene	400	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Indeno (1,2,3-cd) pyrene	0.7	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Napthalene	4	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Phenanthrene	709	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
Pyrene	700	<0.63	<0.57	<0.58	<0.57	<0.62	<0.60
VPH PARAMETERS							
Aliphatics/Acomatics (ag/g)							
C5-C8 Aliphatics	100	<10	<12	<14	<12	<10	<10
C-9-C12 Aliphatics	1,000	<2.7	5.5	<3.5	<3.1	<2.6	26
C9-C10 Aromatics	100	<5.3	<6.1	<6.9	<6.2	<5.3	31
VPH				< 24	<21		
Farget VPH Analytes (ug/g)							
Benzene	10	<1.3	<1.5	<1.7	<1.5	<1.	<1.2
Ethylbenzene	80	<1.3	<1.5	<1.7	<1.5	<1.3	<1.2
n,p-Xylenes	500	<5.3	<6.1	<6.9	<6.2	<5.3	<5.1
Methyl-tert-butyl ether	0.3	<4.0	<4.5	<5.1	<4.6	<3.9	<3.8
Naphthalene	4	<2.7	<3.1	<3.5	<3.1	<2.6	<2.6
o-Xylene	500	<2.7	<3.1	<3.5	<3.1	<2.6	<2.6
Toluene	90	<4.0	<4.5	<5.1	<4.6	<3.9	<3.8

		and the state of the state of the		South Life and			and the second
	MCP S1/GW1	69W-HS-FL-3	69W-HS-WSW-2	69W-HS-WSW-2-2	69W-PL-FL-5	69W-HS-FL-4	69W-PL-ESW-1
DEPTH in feet (bgs)	Standards	7-10	6-9	5-6	7-10	5-8	5-8
EPH PARAMETERS							
Aliphatics/Aromatics (µg/g)							
C9-C18 Aliphatics	1,000	. 36	2,200	14	310	16	120
C19-C36 Aliphatics	2,500	15	290	19	56	17	44
C11-C22 Aromatics	200	24	520	21	80	14	20
EPH		75	3,000	54	450	47	180
Target PAH Analytes (pg/g)							
2-Methylnaphthalene	4	<0.60	14	< 0.58	<0.58	<0.61	<0.58
Acenaphthene	20	<0.60	3.3	< 0.58	<0.58	<0.61	<0.58
Acenaphthylene	100	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Anthracene	1,000	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Benzo (a) anthracene	0.7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Benzo (b) fluoranthene	0.7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Benzo (k) fluoranthene	7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Benzo (a) pyrene	0.7	<0.60	<2.1	< 0.58	<0.58	⊲0.61	<0.58
Benzo (g,h.i) perylene	1,000	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Chrysene	7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Dibenz (a,h) anthracene	0.7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Fluoranthene	1,000	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Flourene	400	<0.60	6.8	< 0.58	<0.58	<0.61	<0.58
Indeno (1,2,3-cd) pyrene	0.7	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
Napthalene	4	<0.60	2.4	< 0.58	<0.58	<0.61	<0.58
Phenanthrene	700	<0.60	4.1	< 0.58	<0.58	<0.61	<0.58
Pyrene	700	<0.60	<2.1	< 0.58	<0.58	<0.61	<0.58
VPH PARAMETERS							
Aliphatics/Aromatics (µg/g)							
C5-C8 Aliphatics	100	<13	<46	<15	<13	<13	<14
C-9-C12 Aliphatics	1,000	5.8	200	< 3.9	130	3.8	<3.6
C9-C10 Aromatics	100	<6.5	290	<7.8	62	<6.5	<7.3
VPH		< 22	500	< 27	200	< 23	< 25
Target VPH Analytes (µg/g)							
Benzene	10	<1.6	<5.6	< 1.9	<1.5	<1.6	<1.7
Ethylbenzene	80	<1.6	<5.6	<1.9	<1.5	' <1.6	<1.7
m,p-Xylenes	500	<6.5	<23	< 7.8	<6.4	<6.5	<7.3
Methyl-tert-butyl ether	0.3	<4.8	<17	< 5.8	<4.7	<4.8	<5.4
Naphthalene	4	<3.2	<12	< 3.9	<3.2	<3.3	<3.6
o-Xylene	500	<3.2	<12	< 3.9	<3.2	<3.3	<3.6
Toluene	90	<4.8	<17	< 5.8	<4.7	<4.8	<5.4

1.0

	MCP S1/GW1	69W-HS-WSW-3	69W-PL-NSW-2	69W-HS-NSW-1	69W-UP-G3-WSW	69W-UP-G3-FL	69W-UP-G3-ESW
DEPTH in feet (bgs)	Standards	5-8	3-6	3-6	4-5	5	4-5
EPH PARAMETERS							
Aliphatics/Aromatics (µg/g)							
C9-C18 Aliphatics	1,000	860	3.3	410	13	10	<3.8
C19-C36 Aliphatics	2,500	130	8.1	68	11	19	5.4
C11-C22 Aromatics	200	220	9	120	10	27	<11
EPH		1,200	20	600	35	57	< 20
Target PAH Analytes (ug/g)							
2-Methylnaphthalene	4	<0.55	<0.53	1.9	<0.55	<0.58	<0.64
Acenaphthene	20	0.83	<0.53	0.79	<0.55	<0.58	<0.64
Acenaphthylene	100	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Anthracene	1,000	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Benzo (a) anthracene	0.7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Benzo (b) fluoranthene	0.7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Benzo (k) fluoranthene	7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Benzo (a) pyrene	0.7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Benzo (g,h.i) perylene	1,000	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Chrysene	7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Dibenz (a,h) anthracene	0.7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Fluoranthene	1,000	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Flourene	400	2.1	<0.53	1.4	<0.55	<0.58	<0.64
Indeno (1,2,3-cd) pyrene	0.7	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Napthalene	4	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
Phenanthrene	700	1.5	<0.53	0.98	<0.55	<0.58	<0.64
Pyrene	700	<0.55	<0.53	<0.61	<0.55	<0.58	<0.64
VPH PARAMETERS Aliphatics/Aromatics (42/2)							
C5-C8 Aliphatics	100	<12	<12	<13	<9.2	<9.8	<11
C-9-C12 Aliphatics	1,000	150	<2.9	7.4	<2.3	<2.5	<2.7
C9-C10 Aromatics	100	100	<5.8	<6.6	<4.7	<5.0	<5.4
VPH	100	250	< 20	<23	<16	<17	<19
Target VPH Analytes (ug/g)		230	- 40	- 43	510	-1/	~15
Benzene	10	<1.4	<1.4	<1.6	<1.1	<1.2	<1.3
Ethylbenzene	80	<1.4	<1.4	<1.6	<1.1	<1.2	<1.3
n,p-Xylenes	500	<6.0	<5.8	<6.6	<4.7	<5.0	<5.4
Methyl-tert-butyl ether	0,3	<4.4	<4.3	<4.9	<3.5	<3.7	<4.0
Naphthalene	4	<3.0	<2.9	<3.3	<2.3	<2.5	<2.7
o-Xylene	500	<3.0	<2.9	<3.3	<2.3	<2.5	<2.7
Toluene	90	<4.4	<4.3	<4.9	<3.5	3.7	<4.0

N. 1. 4040. 11	MCP SI/GW1	69W-UP-G2-WSW	69W-UP-G2-FL	69W-UP-G2-FL-2	69W-UP-G2-ESW	69W-UP-G1-WSW
DEPTH in feet (bgs)	Standards	4-5	07W-01-02-1L	5	4-5	03W-01-G1-WSV
EPH PARAMETERS	of all using s	10			+5	
Aliphatics/Aromatics (ug/g)						
C9-C18 Aliphatics	1,000	43	1,300	4.3	3.4	2,000
C19-C36 Aliphatics	2,500	28	300	6.1	<4.2	460
C11-C22 Aromatics	200	17	130	<10	<8.9	190
EPH		88	1,700	<18	<16	2,700
Farget PAH Analytes (ug/g)						
2-Methylnaphthalene	4	<0.52	<0.56	<0.59	<0.52	<0.57
Acenaphthene	20	<0.52	<0.56	<0.59	<0.52	<0.57
Acenaphthylene	100	<0.52	<0.56	<0.59	<0.52	<0.57
Anthracene	1,000	<0.52	<0.56	<0.59	<0.52	<0.57
Benzo (a) anthracene	0.7	<0.52	<0.56	<0.59	<0.52	<0.57
Benzo (b) fluoranthene	0.7	<0.52	<0.56	<0.59	<0.52	<0.57
Benzo (k) fluoranthene	7	<0.52	<0.56	<0.59	<0.52	<0.57
Benzo (a) pyrene	0,7	<0.52	<0.56	<0.59	<0.52	<0.57
Benzo (g,h.i) pervlene	1,000	<0.52	<0.56	<0.59	<0.52	<0.57
Chrysene	7	<0.52	<0.56	<0.59	<0.52	<0.57
Dibenz (a,h) anthracene	0.7	<0.52	<0.56	<0.59	<0.52	<0.57
Fluoranthene	1,000	<0.52	<0.56	<0.59	<0.52	<0.57
Flourene	400	<0.52	<0.56	<0.59	<0.52	<0.57
Indeno (1,2,3-cd) pyrene	0.7	<0.52	<0.56	<0.59	<0.52	<0.57
Napthalene	4	<0.52	<0.56	<0.59	<0.52	<0.57
Phenanthrene	700	<0.52	<0.56	<0.59	<0.52	<0.57
Рутепе	700	<0.52	<0.56	<0.59	<0.52	<0.57
VPH PARAMETERS						
Aliphatics/Aromatics (µg/g)						
C5-C8 Aliphatics	100	<9	<9.5	<13	<8.9	<13
C-9-C12 Aliphatics	1,000	<2.3	<2.4	<3.2	<2.2	180
C9-C10 Aromatics	100	<4.5	<4.8	<6.5	<4.5	120
VPH		< 16	<17	< 22	< 16	300
Farget VPH Analytes (µg/g)						
Benzene	10	<1.1	<1.2	<1.6	<1.1	<1.5
Ethylbenzene	80	<1.1	<1.2	<1.6	<1.1	<1.5
m,p-Xylenes	500	<4.5	<4.8	<6.5	<4.5	<6.4
Methyl-tert-butyl ether	0.3	<3.4	<3.6	<4.8	<3.3	<4.7
Naphthalene	4	<2.3	<2,4	<3.2	<2.2	<3.2
o-Xylene	500	<2.3	<2.4	<3.2	<2.2 .	<3.2
Toluene	90	<3.4	<3.6	<4.8	<3.3	<4.7

- 6	MCP SI/GW1	69W-UP-G1-WSW-2	OW UP CI FI	69W-UP-G1-ESW	69W-V-WSW-1	69W-V-FL-1	69W-V-ESW-1
DEPTH in feet (bgs)	Standards	3-4	09W-UP-GI-FL	4-6	5-7	09W-V-FL-1	6-8
EPH PARAMETERS	Standards	5-4	4	4-0	5-1		0-0
Aliphatics/Aromatics (µg/g)							
C9-C18 Aliphatics	1,000	720	14	120	1,200	600	1,700
C19-C18 Aliphatics	2,500	350	14	68	380	130	480
C11-C22 Aromatics	2,500	39	10	28	82	77	180
EPH	2007	1,100	36	220	1,600	800	2,300
		1,100	30	220	1,000	800	2,300
Target PAH Analytes (µg/g) 2-Methylnaphthalene	4	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
	20		<0.59		<0.70	<0.59	<0.60
Acenaphthene	100	<0.54 <0.54	<0.59	<0.54 <0.54	<0.70	<0.59	<0.60
Acenaphthylene Anthracene	1,000	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Benzo (a) anthracene	0.7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Benzo (a) anunracene Benzo (b) fluoranthene	0.7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Benzo (b) fluoranthene Benzo (k) fluoranthene	7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
	0.7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Benzo (a) pyrene			<0.59	<0.54	<0.70 <0.70	<0.59	<0.60
Benzo (g,h.i) perylene	1,000	<0.54 <0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Chrysene	0.7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Dibenz (a,h) anthracene			and the second se		<0.70	<0.59	
Fluoranthene	1,000	<0.54 <0.54	<0.59 <0.59	<0.54 <0.54	<0.70	<0.59	<0.60 <0.60
Flourene	0.7	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
Indeno (1,2,3-cd) pyrene					<0.70	<0.59	<0.60
Napthalene	4	<0.54	<0.59	<0.54		<0.59	the second se
Phenanthrene	700	<0.54	<0.59	<0.54	<0.70		<0.60
Pyrene	700	<0.54	<0.59	<0.54	<0.70	<0.59	<0.60
VPH PARAMETERS							
Allphatics/Aromatics (ug/g)	100	<12	<12	<11	<16	<13	<15
C5-C8 Aliphatics			43	<11	<16	43	36
C-9-C12 Aliphatics	1,000	3		1	<8.1	37	35
C9-C10 Aromatics VPH	100	<6 <21	24 67	<5.8 < 20	<28	80	71
		<u>~21</u>	0/	× 20	~20	80	/1
Target VPH Analytes (µg/g)		-14	15	<1.4	<2.0	<1.6	<1.8
Benzene	10	<1.4	<1.5	<1.4	<2.0	<1.6	<1.8
Ethylbenzene	80 500	<1.4	<1.5	<1.4	<2.0	<6.5	<7.4
m,p-Xylenes	0.3	<6.0	<4.6	<4.3	<6.0	<4.8	<5.5
Methyl-tert-butyl ether				A101			1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1
Naphthalene	4	<3.0	<3.1	<2.9	<4.0	<3.2	<3.7
o-Xylene	500	<3.0	3.1	<2.9 <4.3	<4.0 <6.0	<3.2 <4.8	<3.7 <5.5
Toluene	90	<4.4	<4.6	\$4.5	0.0	-4.8	\$3.5

· · · ·	MCP S1/GW1	69W-V-NSW-1	69W-PL-FL-6	69W-PL-FL-7	69W-PL-ESW-2
DEPTH in feet (bgs)	Standards	6-8	7	6	4-6
EPH PARAMETERS					
Aliphatics/Aromatics (ug/g)					
C9-C18 Aliphatics	1,000	74	310	370	5.6
C19-C36 Aliphatics	2,500	120	36	110	7.9
C11-C22 Aromatics	200	22	120	23	<11
EPH		210	460	500	<20
Target PAH Analytes (49/2)					
2-Methylnaphthalene	4	<0.58	1.2	< 0.61	< 0.65
Acenaphthene	20	<0.58	0.92	< 0.61	< 0.65
Acenaphthylene	100	<0.58	< 0.61	< 0.61	< 0.65
Anthracene	1,000	<0.58	< 0.61	< 0.61	< 0.65
Benzo (a) anthracene	0.7	<0.58	< 0.61	< 0.61	< 0.65
Benzo (b) fluoranthene	0.7	<0.58	< 0.61	< 0.61	< 0.65
Benzo (k) fluoranthene	7	<0.58	< 0.61	< 0.61	< 0.65
Benzo (a) pyrene	0.7	<0.58	< 0.61	< 0.61	< 0.65
Benzo (g,h.i) perylene	1,000	<0.58	< 0.61	< 0.61	< 0.65
Chrysene	7	<0.58	< 0.61	< 0.61	< 0.65
Dibenz (a,h) anthracene	0.7	<0.58	< 0.61	< 0.61	< 0.65
Fluoranthene	1,000	<0.58	< 0.61	< 0.61	< 0.65
Flourene	400	<0.58	1.4	< 0.61	< 0.65
Indeno (1,2,3-cd) pyrene	0.7	<0.58	< 0.61	< 0.61	< 0.65
Napthalene	4	<0.58	< 0.61	< 0.61	< 0.65
Phenanthrene	700	<0.58	< 0.61	< 0.61	< 0.65
Pyrene	700	<0.58	< 0.61	< 0.61	< 0.65
VPH PARAMETERS					
Aliphatics/Aromatics (µg/g)					
C5-C8 Aliphatics	100	<13	<13	<13	<14
C-9-C12 Aliphatics	1,000	<3.2	52	< 3.2	< 3.5
C9-C10 Aromatics	100	<6.5	33	< 6.5	< 7.0
VPH		< 23	84	<22	<24
Target VPH Analytes (µg/g)	<u> </u>				
Benzene	10	<1.6	<1.6	<1.6	<1.7
Ethylbenzene	80	<1.6	<1.6	< 1.6	. <1.7
m,p-Xylenes	500	<6.5	< 6.5	< 6.5	<7.0
Methyl-tert-butyl ether	0,3	<4.8	< 4.8	< 4.8	< 5.2
Naphthalene	4	<3.2	< 3.2	< 3.2	< 3.5
o-Xylene	500	⊲.2	< 3.2	< 3.2	< 3.5
Toluene	90	<4.8	< 4.8	< 4.8	< 5.2

TABLE 7-14 RI GROUNDWATER FIELD ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Depth (ft bgs) Date Analyzed: Dilution:	ZWR-96-35X 7 14-Sep-95 1	ZWR-95-36X 7 15-Sep-95 1'	ZWR-95-37X 7 15-Sep-95 1	ZWR-95-38X 5 14-Sep-95 1	ZWR-95-39R 5 14-Sep-95 1	ZWR-95-40X 6 15-Sep-95 1	ZWR-95-41X 6 15-Sep-95 1	ZWR-95-42X 6 15-Sep-95 1
ANALYTES	Reporting Limit							1.12	
1,1-DCE	5 μg/l	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
Chloroform	2 μg/l	2,0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
1,1,1-TCA	2 μg/l	2,0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Carbon Tetrachloride	2 µg/l	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ
Trichloroethene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Tetrachloroethene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Benzene	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Toluene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Chlorobenzene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Ethylbenzene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
m/p-Xylene	4 μg/l	4.0 U	4.0 U	7	4.0 U				
o-Xylene	2 μg/l	2.0 U	2.0 U	5.2	2.0 U				
TPH-dro	100 mg/l	NA	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 µg/l	NA	NA	NA	NA	NA	NA	NA	NA

4

Dana Tara

TABLE 7-14 RI GROUNDWATER FIELD ANALYTICAL RESULTS AOC 69W

a.,	Site ID Depth (ft bgs) Date Analyzed: Dilution:	ZWR-95-43X 5 15-Sep-95 1	ZWR-95-44X 6 15-Sep-95 1	ZWR-95-45X 6 15-Sep-95 1	ZWR-95-46X 6 15-Sep-95 1	ZWR-95-47X 6 22-Sep-95 1	ZWR-95-48X 4. 22-Sep-96 1	ZWR-96-49X 4 22-Sep-95 1	ZWR-95-50X 10 22-Sep-95 1
ANALYTES	Reporting Limit						1	1	
1,1-DCE	5 µg/l	5,0 U	5.0 U	5.0 U	5.0 U				
Chloroform	2 μg/l	2.0 U	2.0 U	2.0 U					
1,1,1-TCA	2 μg/l	2.0 U	2.0 U	2.0 U					
Carbon Tetrachloride	2 μg/l	2.0 UJ	2.0 UJ	2.0 UJ	2.0 UJ	2.0 U	2.0 U	2.0 U	2.0 U
Trichloroethene	2 µg/l	2.0 U	2.0 U	2.0 U					
Tetrachloroethene	2 µg/l	2.0 U	2.0 U	2.0 U					
Benzene	2 µg/l	2.0 U	2.0 U	2.0 U					
Toluene	2 μg/l	2.0 U	2.0 U	2.0 U					
Chlorobenzene	2 μg/l	2.0 U	2.0 U	2.0 U					
Ethylbenzene	2 μg/l	2.0 U	2.0 U	2.0 U					
m/p-Xylene	4 μg/l	4.0 U	4.0 U	7.3	4.0 U	4.0 UJ	4.0 UJ	4.0 UJ	4.0 UJ
o-Xylene	2 μg/l	2.0 U	2.0 U	2.0 U					
TPH-dro	100 mg/l	NA	NA	NA	NA	NA	NA	NA	NA
TPH-gro	100 μg/l	NA	NA	NA	NA	250	120	100 U	100 U

TABLE 7-14 RI GROUNDWATER FIELD ANALYTICAL RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Site ID Depth (ft bgs) Date Analyzed: Dilution:	ZWR-95-51X 15 22-Sep-95 1	ZWR-95-52X 5 22-Sep-95 1	ZWR-95-53X 5 22-Sep-95 1	ZWR-95-54X 9 29-Sep-95 1	ZWR-95-55X 9 29-Sep-95 1
ANALYTES	Reporting Limit		Press of the same of			
1,1-DCE	5 µg/l	5.0 U	5.0 U	5.0 U	5.0 U	5.0 U
Chloroform	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
1,1,1-TCA	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Carbon Tetrachloride	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Trichloroethene	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 Ú
Tetrachloroethene	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Benzene	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Toluene	2 µg/l	2.0 U	2.0 U	2.0 U	2.0 UJ	2.0 UJ
Chlorobenzene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
Ethylbenzene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
m/p-Xylene	4 μg/l	4.0 ÚJ	4.0 UJ	4.0 UJ	4.0 U	4.0 U
o-Xylene	2 μg/l	2.0 U	2.0 U	2.0 U	2.0 U	2.0 U
TPH-dro	100 mg/l	NA	NA	NA	NA	NA
TPH-gro	100 µg/l	100 U	100 U	100 U	100 U	100 U

NOTES:

U = Concentration is less than reporting limit

J = Concentration is estimated

NA = Not Analyzed

1.1

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Site ID: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATIONS Usits: (µg/L)	69W-94-09 MXZW09X3 DV4F*269 11/02/95 8.5 (нg/L)	69W-94.09 MXZW09X3 DV4W*269 11/02/95 8,5 (цц/L)	69W-94-09 MXZW09X3 DV4W*269 11/03/95 8.5 (µg/L)	69W-94-09 MXZW09X4 12/18/97 8.5 (µg/L)	69W-94-10 MXZW10X3 DV4F*271 11/02/95 9.5 (ag/L)	69W-94-10 MXZW10X3 DV4W*271 11/02/95 9.5 (mg/L)	65W-94-10 MXZW10X4 DV4F*272 02/14/96 9.5 (#g/L)	69W-94-10 MXZW10X4 DV4W*272 02/14/96 9.5 (ug/L)
PAL METALS Aluminum 6870 Arsenic 10.5	< 141 F < 2.54 F	448 < 2.54	448 < 2.54		141 F 160 F	< 141	< 141 F 150 F	< 141
Barium 39.6	16.3 F	17.2	17.2		14 F	15.2	16.5 F	18.4
Calcium 14700	17300 F	17400	17400		19300 F	21109	26300 F	29000
Copper 8.09	< 8.09 F	< 8.09	< 8.09		8.09 F	< 8.09	< 8.09 F	< 8.09
Iron 9100	< 38.8 F	443	443		16300 F	18600	28800 F	34200
Lead 4.25	2.17 F	1,41	1.41		1.26 F	3.36	< 1.26 F	< 1.26
Magnesium 3480	1690 F	1800	1800		1880 F	2030	2660 F	3020
Manganese 291	3.89 F	13	13		1210 F	1440	3070 F	4010
Potassium 2370	2400 F	2580	2580		4790 F	5040	4720 F	5130
Sodium 10800	27809 F	27100	27100		22100 F	23500	21300 F	23500
PESTICIDES/PCBS								
Chlordane - Garrina		< .075 T	< .075 T			< .075 T		.115 CZ
Heptachlor Epoxide		< .0245	< .0245			< .0245		.0586 C
SVOCs		(_			
2-Methyinaphthalene		< 1.7	< 1.7			500 X	0	600
Acenaphthene		< 1.7	< 1,7			< 20		< 100
Dibenzofian		< 1.7	< 1.7			< 20		< 100
Diethyl Phthalate		< 2	< 2			< 20		< 200
Fluorene		< 3.7	< 3.7			80		< 300
Naphthalene Phenanthrene		< .5	< .5			200 X 100		200
Bis(2-ethylhexyl) Phthalate		< .5 4.5	< 5 45			500	a last construct a	200 500
								1944
VOCs								
1,1,1-Trichloroethane		< 5	< 5			< .5		< 2
Acetone Chloroform		14	14 < .5			< 13 < 5	the second se	< 60 < 2
Ethylbenzene		< 5	< .5	s U		14		< 2 20
Tohiene		< 5	< 5	3 0		3		< 2
Trichloroethylene		< .5	< 3			< .5		< 2
Xylenes		< ,84	< 84			1.4		< 4
WET CHEMISTRY								
Alkalinity		36000	36000			38000		89000
Chloride		44000	44000			46000		42000
Nitrite, Nitrate-non Specific		1700	1700			1100		570
Nitrogen By Kjeldahl Method		286	286			952		200
Phosphate Sulfate		1900 18000	1900 18000			500 10000		< 13.3 < 10000
Total Dissolved Solids		112000	112000			128000		202000
VIII LABOUTOD DUBLIS		112000	212000			120000		202000
Total Hardness Total Suspended Solids		52000 1660000	52000 1660000			24000 259000	-	80800 35000
						285000 D	-	GARA -
OTHER								
Total Organic Carbon		< 185	< 185			159000	1	228000

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

			DEVENS, I	ASSACHUSEIS			and the second se	
Site ID: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATIONS Unite: (µg/L) PAL METALS	69W-94-10 MXZW10X5 DV4W*513 10/01/96 0 (µg/L)	69W-94-10 МХZW10Х6 12/11/97 9.5 (нg/L)	69W-94-11 MDZW11X4 DV4F*456 02/14/96 9.5 (4g/L)	69W-94-11 MDZW11X4 DV4W*456 0214/95 9.5 (Hg/L)	69W-94-11 MXZW11X3 DV4F*273 11/03/95 9.5 (µg/1.)	69W-94-11 MXZW11X3 DV4W*273 11/03/95 9.5 (Hg/L)	69W-94-11 MXZW33X4 DV4F*274 02/14/96 9.5 (µg/L)	69W-94-11 MXZW11X4 DV4W*274 02/14/96 9.5 (ug/L)
		1	2 141 E	< 141 D	5 141 E	1 616		2 10
Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 fron 9100 Lead 4.25 Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800			< 141 F 143 DF 7,69 F 22300 F < 8,09 F < 8,09 F < 1,26 DF 2510 F 2090 F 30600 F	< 141 D 13 D 7.44 D 1000 D < 8,09 D 599 D < 1.26 D 2430 D 4031 D 1560 D 30000 D	 141 F 143 F 143 F 143 F 143 F 143 F 143 F 140 F 	616 11 21900 < 8.09 845 < 1.26 2520 317 2110 32700	 < 141 F 9.38 F 6.18 F 20103 F < 8.09 F 484 F < 1.26 F 2240 F 405 F 2110 F 27400 F 	< 141 81.3 6.93 71000 < 8.09 807 < 1.26 2430 305 1670 79800
PESTICIDES/PCBS				- 010 DT		1		
Chlordane - Gamma Heptachlor Epoxide			-	< .075 DT < .0245 D		< .075 T < .0245		< .075 7 < .0245
SVOCs								
2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene Bis(2-ethylhexyl) Phthalate		73 4 J 39 3 J		$\begin{array}{cccccccccccccccccccccccccccccccccccc$		< 1.7 < 1.7 < 1.7 < 2 < 3.7 < 5 < 5 < 4.8		< 1.7 < 1.7 < 1.7 5.1 < 3.7 < 5 < 5 < 5 < 48
VOCs						6		
Vots Acetone Chloroform Ethylbenzene Toluene Trichloroethylene Xylenes		16		 5 D 13 D 5 D 5 D 5 D 5 D 5 D 5 D 84 D 		1.5 < 13 < 5 < 5 < 5 < 3 3.3 < 84		< 5 < 13 < 5 < 5 < 5 < 5 < 3 < 84
WETCHEMISTRY								
Alkalinity Chloride Nitrite, Nitrate-non Specific Nitrogen By Kjeldahl Method Phosphate Sulfate Total Dissolved Solids Total Hardness Total Suspended Solids	1710 36.2 139000 137000000 37000			45000 D 50000 D 2000 D < 183 D < 13.3 D 15000 D 166000 D 61200 D 5000 D		92000 44000 1300 400 30.3 14000 163000 2000 26000		44000 50000 < J0 < 183 22.7 15000 163000 66200 < 4000
Louis or Sherroot Dolling	57000					2000		4000
OTHER Fotal Organic Carbon	6070					T		
Total Organic Carbon Total Petroleum Hydrocarbons	< 170			< 175 D		< 169		2420

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Site ID:	69W-94-11	69W-94-11	DEVENS, MASSACHUSEI	A*	1	1 20001121-22	
Ster D: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATIO: Units: (pg/L) PAL METALS	MXZW11X5 DV4W*515 10/01/96	69W-94-11 MXZW11X6 12/11/97 9.5 (µg/L)	69W-94-12 MDZW12X3 DV4F4456 11/02/95 8 (pg/L)	69W-94-12 MDZW12X3 DV4W*450 11/02/95 8 (µg/L)	69W-94-12 MEXZW12X3 DV4F*275 11/0295 8 (µg/L)	69W-94-12 MXZW12X3 DV4W*275 11/02/95 8 (µg/L)	69W-94-12 MXZW12X4 DV4F*276 02/1396 8 (µg/L)
Aluninum 6870	1	1	< 141 DF	185 D	< 141 F	198	< 141 F
Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 Iron 9100 Lead 4.25 Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800			 2.54 DF 5 DF 19600 DF 8.09 DF 38.8 DF 1.74 DF 2010 DF 3.64 DF 1690 DF 20260 DF 	< 2.54 D < 5 D 18700 D < 8.09 D 188 D 1970 D 10.9 D 1750 D 265000 D	<pre>< 2.54 F < 5 F 30000 F < 8.09 F < 8.09 F 44.1 F 3.36 F 2160 F < 2.75 F 1770 F 38700 F</pre>	 2.54 5 17306 8.09 249 1.26 2040 9 1650 37566 	 < 2.54 F < 5 F 78000 F < 8.09 P < 38.8 F < 1.26 F 1870 F 4.29 F 1660 F (85800 F
PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide				< .075 TD < 0245 D		< .075 T < .0245	
SVOCs					L		
2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene Bis(2-ethylhexyl) Phthalate	-	10 U 10 U 10 U 10 U		< 1.7 D < 1.7 D < 1.7 D < 2 D < 3.7 D < 3 D < 5 D < 48 D		 1.7 1.7 1.7 2 3.7 5 5 4.8 	
VOCs							
1,1,1-Trichloroethane Acetorie Chloroform Ethylbenzene Toluene Trichloroethylene Xylenes		s U		< <u>5</u> D < <u>13</u> D < <u>5</u> D < <u>5</u> D < <u>5</u> D < <u>5</u> D < <u>84</u> D		< 5 < 13. 55 6 < 5 < 5 < 84	
WET CHEMISTRY						1	
Alkaliniity Chloride Nitrite, Nitrate-non Specific Nitrogen By Kjeldahl Method Phosphate Sulfate Total Dissolved Solids Total Hardness Total Suspended Solids	< 183 < 13.3 136000 74200000 < 4000			48000 D 42000 D 3100 D 1050 D 2000 D 16000 D 117000 D 1580000 D		41000 44000 3400 952 2200 16000 116000 60000 1520000	
OTHER Total Organic Carbon Total Petroleum Hydrocarbons	< 1000 < 172			< 180 D		< 176	

TABLE 7 - 15 RI GROUNDWATER OFF-SITE ANALYTICAL RESULTS [a] AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

	and the second se		EVENS, MASSACHUSE	10			
Site ID: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATION Units: (µg/L) PAL METALS	69W-94-12 MXZW12X4 DV4W*276 02/13/96 S 8 (µc/L)	69W-94-12 MXZW12XS 12/19/97 8 (µg/L)	69W-94-13 MXZW13X3 DV4F*277 11/62/95 8 (µg/L)	69W-94-13 MXZW13X3 DV4W*277 15/02/95 8 (µg/L)	69W-94-13 MX2W13X4 DV4F*278 02/13/96 8 (µg/L)	69W-94-13 MXZW13X4 DV4W4278 82/13/96 8 (µg/L)	69W-94-13 MXZW13X5 12/12/97 S (pg/L)
Aluminum 6870	591		< 141 F	< 141	< 141 F	1 2 10 1	
Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 tron 9100 Lead 4.25	< 2.54 6.74 39300 < 8.09 664 < 1.26		290 F 15.7 F 25960 F < 8.09 F 26100 F 2.71 F	260 16.7 36300 < 8.09 21200 < 1.26	110 F 10.5 F 32600 F < 8.09 F < 1.26 F	< 141 139 10.8 23000 < 8.09 16299 < 1.26	
Magnesium 3480	2180		2860 F	2960	2490 F	2610	
Manganese 291 Potassium 2370	22.6		2110 F	2010	1870 F	1980	
Potassium 2370 Sodium 10800	1610 21400		2740 F 29600 F	1570 29100	2080 F 18400 F	2190	
10000	Second Second			(////#TS999//////	20000 892999 600000	22469	
PESTICIDES/PCBS							
Chlordane - Gamma	< .075 T			< .075 T		< .075 T	
Heptachlor Epoxide	< .0245			< .0245	the second se	< .0245	
SVOCs						11	
2-Methylnaphthalene	< 1.7			8.7		7.6	13 J
Acenaphthene	< 1.7			2.3		2.1	
Dibenzofuran	< 1.7			2.4		2.3	
Dicthyl Phthalate	4.1			< 2		< 2	
Fluorene Naphthalene	< 3.7 < 5			< 3.7 16		< 3.7 15	3 J 10 J
Phenanthrene	< .5			2.6		1.8	2 1
Bis(2-ethylhexyl) Phthalate	< 4.8			< 4.8		< 4.8	
C. C. Designer Street S				12 28		1 Nov 1	
VOCs							
1,1,1-Trichloroethane Acetone	< .5 < 13			< 5 < 13		< .5 < 13	
Chloroform	< 5			1.1		< .5	
Ethylbenzene	< .5	5 U		< .5		3.6	3 1
Toluene	1.1			< .5		< .5	
Trichloroethylene	< .5			< .5		< .5	
Xylenes	< .84			< .84		< .84	
WET CHEMISTRY	1 1					41	
Alkalinity	27000			104000		72000	
Chloride	43000			26300		35000	
Nitrite, Nitrate-non Specific	3000 257			1400		111	
Nitrogen By Kjeldahl Method Phosphate	16.6			714 111		352 15.8	
Sulfate	14000			< 10000		< 10000	
Total Dissolved Solids	149000			165000		165000	
Total Hardness Total Suspended Solids	58000 16000			10000 74000		72800	
Total suspended Sollas	10000			/4000		18000	
					1		
OTHER	1			1	1	1 1	
Total Organic Carbon Total Detrolaum Hudroschene	< 175			523		< 181	
Total Petroleum Hydrocarbons	\$ 1/3			345		101	

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

		в	EVENS, MASSACHUSE	15	and the second second	A second contract of	and the second s
Site ID: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATION: Units: (jug/L) PAL METALS	69W-94-14 MXZW14X3 DV4F*279 11/03/95 S 8 (µg/L)	69W-94-14 MXZW14X3 DV4W*279 11/03/95 8 (µg/L)	69W-94-16 MXZ/W14X4 DV4F*280 02/13/96 8 (µg/L)	69W-94-14 MXZW14X4 DV4W*280 02/13/96 8 (µg/L)	69W-94-14 MXZW14X5 12/12/97 8 (µg/L)	ZWM-95-15X MXZW15X1 DV4F*281 11/01/95 8 (µg/L)	2WM-95-15X MX2W15X1 DV4W*281 11/01/95 8 (µg/L)
Aluminum 6870	< 141 F	< 141	< 141 F	< 141		. < 141 F	< 141
Arsenic 10,5 Barium 39,6 Calcium 14700 Copper 8,09 Iron 9100 Lead 4,25 Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800	11 F 16 F 19000 F 2000 F 2000 F 2130 F 2270 F 37000 F	<pre></pre>	 10.3 F 10.6 F 12500 F 8.09 F 3720 F 1.26 F 2040 F 325 F 1560 F 23500 F 	 141 10.3 11.3 17800 8.09 3840 2.26 2090 340 1880 23700 		3.73 F 14.4 F 18000 F < 8.09 1030 F < 1.26 1780 F 301 F 2200 F 43506 F	<pre></pre>
PESTICIDES/PCBS Chlordane - Gamma		< .075 T		< .075 T		r	< .075 T
Heptachlor Epoxide		< .0245		< .0245			< .0245
SVOCs	-						
2-Methylnaphthalene Aceraphthene Dibenzofuran Dichtyl Phthalate Fluorene Naphthalene Phenanthrene Bis(2-ethylhexyl) Phthalate		< 1.7 < 1.7 < 1.7 < 2 < 3.7 < 5 < 5 < 4.8		< 1.7 < 1.7 < 1.7 < 2 < 3.7 < 5 < .5 < 4.8	11 UI 11 UI 11 UI 11 UI		< 1.7 < 1.7 < 1.7 < 2 < 3.7 < 5 < 5 < 4.8
VOCs 1,1,1-Trichloroethane	I					1	
1,1,1-Inchloroethane Acetone Chloroform Ethylbenzene Toluene Trichloroethylene Xylenes		< .5 < 13 < .5 < .5 < .5 < .5 < .5 < .5 < .84		< 5 < 13 < 5 < 5 .61 < 5 < 84	s U		< .5 < 13 < .5 < .5 < .5 < .5 < .5 < .84
WET CHEMISTRY	1						
Alkalinity Chloride Nitride, Nitrate-non Specific Nitrogen By Kjeldahl Method Phosphate Sulfate Total Dissolved Solids Total Hardness Total Suspended Solids		27000 26300 1700 848 1700 < 10000 129000 58000 970000		44000 28500 450 210 < 13.3 28000 158000 52800 8000			43000 55000 2000 324 143 30000 158000 42000 113000
OTHER. Total Organic Carbon							
Total Organic Carbon Total Petroleum Hydrocarbons		1960		< 175			281

1.5

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Sample Date: BACKGROUND Depth: Depth: CONCENTRATIONS (ugeL) PAL METALS (ugeL) Aluminum 6870 Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 Iron 9100 Lead 4.25 Magnesium 3480 Marganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Clolordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Enaphthere Acenaphthere Diberty/Phthalate Puorene Naphthalene Acenaphthere Bis(2 - thylfncyf) Phthalate	02/13/96	MXZW15X2 DV4W*282 02/13/96	MXZW15X3	MXZW16X1 DV4F*283 11/01/95	MXZW16X1 DV4W*283 11/01/95	ZWM-95-16X MXZW16X2 DV4F*284 02/13/96	ZWM-95-16X MXZW16X2 DV4W*284 62/13/96
Units: (µg/L) PAL METALS Aluminum Aluminum 6870 Assenie 10.5 Barium 39.6 Calcitum 14700 Copper 8.09 Itom 9100 Lead 4.25 Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofinan Diethyl Phthalate Fluorene Naphthalene Phenanthrene Naphthalene	8	8	8	11.3	11.3	11.3	11.3
Aluminum 6870 < Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 < Iron 9100 Lead 4.25 < Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Na	(µg/L)	(µg/L)	(H8/L)	(µg/L)	(µg/L)	()4g/L)	(pg/L)
Arsenic 10.5 Barium 39.6 Calcium 14700 Copper 8.09 < Iron 9100 Lead 4.25 < Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofinan Diethyl Phthalate Fluorene Naphthalene Fluorene Naphthalene Acenaphthene Fluorene Naphthalene Fluorene Naphthalene Acenaphthene Fluorene Naphthalene Fluorene Naphthalene Fluorene Naphthalene Fluorene Naphthalene Fluorene Naphthalene Acenaphthene Fluorene Naphthalene Fluorene Naphthalene Fluorene Naphthalene Fluorene State Sta							
Barium 39.6 Calcium 14700 Copper 8.09 Iron 9100 Lead 4.25 < Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofinan Diethyl Phthalate Fluorene Naphthalene Phenanthrene	141 F <	141		< 141 F	< 141	< 141 F	< 141
Calcium 14700 Copper 8.09 Copper 8.09 Copper 8.09 Copper 8.09 Copper 9100 Lead 4.25 Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Phenanthre	5.44 F	5.65		11.2 F	11.8	< 2.54 F	< 2.54
Copper 8.09 < Iron 9100 Lead 4.25 < Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Florome Naphthalene Naphthalene Phenanthrene	10 F	10.6		9.81 F	10.2	9.91 F	10.8
Iron 9100 Lead 4.25 < Magnesium 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Distryl Phthalate Fluorene Naphthalene Phenanthrene	16700 F	17200		21800 F	21000	14600 F	1.5300
Lead 4.25 < Magneseinn 3480 Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofinan Diethyl Phthalate Fluorene Naphthalene Phenanthrene Phenanthr	8.09 F < 1080 F	8.09 1300		< 8.09 F 1180 F	< 8.09 1290	< 8.09 F 391 F	< 8.09 326
Magnesium 3480 Marganese 291 Potassium 2370 Sodium 10800 ESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	1,26 F <	1.26		< 1.26 F	< 1.26	391 F < 1.26 F	< 1.26
Manganese 291 Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Naphthalene Phenanthrene Phena	1610 F	1690		2200 F	2230	1480 F	1580
Potassium 2370 Sodium 10800 PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene PhenathrenePhenathrene Phenanthrene Phenanthr	256 F	285		992 F	1010	151 F	145
PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	1860 F	2000		3109 F	3050	2430 F	2860
Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	32400 F	33600		26400 F	26709	20600 F	21700
Chlordane - Gamma Heptachlor Epoxide SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	and a state of the				()))))))))))))))	0.000	souther the state of the state
Heptachlor Epoxide SVOCs 2-Methylnaphthalene Accnaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	<	.075 T			< .075 T		< .075 T
SVOCs 2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	× ×	.0245			< .0245		< .075 1
2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	4	102-15			.0240	I	10040
Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene							
Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene	<	1.7			< 1.7		< 1,7
Diethyl Phthalate Fluorene Naphthalene Phenanthrene	<	1.7		(< 1,7		< 1,7
Fluorene Naphthalene Phenanthrene	<	1.7			< 1.7		< 1.7
Naphthalene Phenanthrene	<	2			< 2	1 1	< 2 < 3.7
Phenanthrene	< <	3.7			< 3.7 < 5		< 3.7 < .5
	×	5			< 5		< 5
	× ×	4.8	a company da la		< 4.8		4.5
VOCs							
1,1,1-Trichloroethane	< <	.5		1	< .5 < 13		< 5
Acetone Chloroform	0	13			< .5		< 13 < 5
Ethylbenzene	<	5	5 U		< 5		< 5
Tohuene		1.2	2 0	1 D	< 5	1 1	9
Trichloroethylene	<	.5			1.2	1 1	< .5
Xylenes	<	.84			< .84		< .84
WET CHEMISTRY							h
Alkalinity	1	40000			50000	1	24000
Chloride		48000			46000		37000
Nitrite, Nitrate-non Specific		2300	1		47		2100
Nitrogen By Kjeldahl Method		257			219		210
Phosphate	<	13.3			147		< 13,3
Sulfate		18000			10000		14000
Total Dissolved Solids		167000			120000		142000
Total Hardness		52000			< 1000		43600
Total Farmerss Total Suspended Solids		8000			214000		7000
CITIED							
OTHER Total Organic Carbon Total Petroleum Hydrocarbons	<	177	1		1340		< 175

[a] The data for 1995 and 1996 are for AEC methods; non-detects reflect the Certified Reporting Limit (CRL). The data for 1997 are for SW-846 methods; non-detects reflect the Practical Quantitation Limit (PQL).

11111

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Site ID: Field Sample Number: Lab Sample Number: DEVENS Sample Date: BACKGROUND Depth: CONCENTRATIONS Units: (ug/L) PAL METALS	ZWM-95-16X MXZW16X3 DV4W*514 10/01/96 8 (µg/L)	ZWM-95-16X MXZW16X4 12/18/97 11.3 (pg/L)	ZWM-95.17X MXZW17X1 DV4F*285 11/03/95 17.2 (µg/L)	ZWM-95-17X MXZW17X1 DV4W*225 11/03/95 17.2 (µg/L)	ZWM-95-17X MXZW17X2 DV4F*286 92/12/96 17,2 (µg/L)	ZWM-95-17X MXZW17X2 DV4W*286 0211296 17.2 (µg/L)	ZWM-95-17X MXZW17X3 12/16/97 17.2 (jig/L)
PAL METALS Aluminum 6870 Arsenic 10.5 Baium 39.6 Calcium 14700 Copper 8.09 Iron 9100 Lead 4.25 Magnesium 3480 Marganese 291 Potassium 2370 Sodium 10800			<pre>< 141 F < 2.54 F 4.94 F < 8.09 F 88.5 F < 1.26 F 3240 F 8.26 F 2070 F 166000 F</pre>	< 2.54 7.14 23800 < 8.09 853 2.39 3290 26.9 1790 23800	<pre>< 141 F < 2.54 F < 5 F 72200 F < 38.8 F < 1.26 F 2680 F < 2.75 F 1610 F 32300 F</pre>	<pre>< 141 < 2.54 < 5 22700 < 8.09 < 38.8 < 1.26 2720 < 2.75 1480 24900</pre>	
PESTICIDES/PCBS Chlordane - Gamma Heptachlor Epoxide				< .075 T < .0245		< .075 T < .0245	
SVOCs Z-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthalene Phenanthrene Bis(2-ethylhexyl) Phthalate				 1.7 1.7 1.7 2 3.7 5 5 4.9 		< 1.7 < 1.7 < 1.7 < 2 < 3.7 < 5 < 5 < 5 < 5 < 5 < 65	
VOCs 1,1,1-Trichloroethane Acetone Chloroform Ethylbenzene Toluene Trichloroethylene Xylenes		5 U		 < .5 < .13 < .5 < .5 .52 < .5 < .84 		< .5 < 13 < .5 < .5 < .5 < .5 < .5 < .5 < .5 < .5	s U
WET CHEMISTRY Alkalinity Chloride Nitrite, Nitrate-non Specific Nitrogen By Kjeldahl Method Phosphate Sulfate Total Dissolved Solids Total Hardness Total Suspended Solids	< 183 < 13.3 165000 56800000 < 4000			43000 50000 1500 248 99 17000 130000 130000 136000 D 74000 120000 116000 D		18000 66000 3000 229 < 13.3 17000 209000 68800 10000	
OTHER. Total Organic Carbon Total Petroleum Hydrocarbons	< 1000 < 172			< 167		< 181	

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Site ID: Field Sample Number: Lab Sample Number: DEVE Sample Date: BACKGF Depth: CONCENTI Units: (µg/l	OUND 11/02/95 ATIONS 8	2WM-95-18X MX2W18X1 DV4W*287 11/02/95 8 (µg/L)	ZWM-95-18X MXZW18X2 DV4F*288 02/12/96 8 (4@/L)	ZWM-95-18X MXZW18X2 DV4W*288 02/12/96 8 (µg/L)	ZWM-95-18X MXZW18X3 12/16/97 8 (Hg/L)	ZWM-96-19X MDZW19X1 DV4W*304 09/30/96 0 (µg/L)	ZWM.96-19X MXZW19X1 DV4W*510 09/30/96 0 (µg/L)
PAL METALS			1			1	1
Aluminum 687 Arsenic 10.		 < 141 < 2.54 	< 141 F < 2.54 F	< 141 < 2.54		11 million (11)	
Banum 39.			6.46 F	7.02			
Calcium 1470			35500 F	1.02			
Copper 8.0			< 8.09 F	< 8.09			
Iron 910			< 38.8 F	< 38.8			
Lead 4.2			< 1.26 F	< 1.26			
Magnesium 348			1730 F	1750			
Manganese 291			5.97 F	6.47			
Potassium 237			1590 F	1730			
Sodium 1080			37000 F	37260			
1000				(month and a second			
PESTICIDES/PCBS							
Chlordane - Gamma		< 075 T		< ,075 T			
Heptachlor Epoxide		< .0245		< .0245			
			1				
SVOCs 2-Methylnaphthalene	1	< 1.7	1	< 17			1
Acenaphthene		< 1.7		< 1.7 < 1.7			
Dibenzofuran		< 17		< 1.7			
Diethyl Phthalate		< 2		5			
Fluorene		< 3.7		< 3.7			
Naphthalene		< 5	1	< .5			
Phenanthrene		< 5		< .5			
Bis(2-ethylhexyl) Phthalate		< 4.8		< 4.8			
VOCs							
1,1,1-Trichloroethane		< 5	1	< .5			
Acetone		< 13		20			
Chloroform		< .5		< 5	4 4		
Ethylbenzene		< .5		< .5	5 U		
Toluene		< .5		.64			
Trichloroethylene		< .5	- C - C - C - C - C - C - C - C - C - C	< 5 < 84			
Xylenes		.54	-	.04			
WET CHEMISTRY		1	1				1000 C
Alkalinity		44000	(c)	20000			
Chloride		50000		59000		and the second second	
Nitrite, Nitrate-non Specific		2300		3000		, 250 D	
Nitrogen By Kjeldahl Method		210		219		< 183 D	
Phosphate		55.1		< 13.3		18.6 D	19.8
Sulfate		18000		20000		10000	10000
Total Dissolved Solids		109000		183000		177000 D	175000
Total Hardness		46000		46800		85000000 D	82400000
Total Hardness Total Suspended Solids		23000		< 4000		8000 D	
rotar suspended sonds		23000		4000		6000 D	4000
OTHER.		-	1			1 1000 5	1
Total Organic Carbon		1.00		100		< .1000 D	
Total Petroleum Hydrocarbons		< 172		< 175		< 170 D	< 174

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

and the second sec			DD T LI IDI I	LASSACHOSE15			
Units:	DEVENS BACKGROUND CONCENTRATIONS (Htt/L)	ZWM-96-19X MXZW19X2 12/18/97 0 (µg/L)	ZWM-96-19X MDZW19X2 12/18/97 0 . (pg/L)	ZWM-96-20X MXZW20X1 DV4W*511 10/01/96 0 (µg/L)	ZWM-96-20X MXZW20X2 12/16/97 0 (µg/L)	ZWM-96-21X MXZ/¥21X1 DV4W*512 09/30/96 0 (µg/L)	ZWM-96-21X MXZW21X2 12/18/97 0 (µg/L)
PAL METALS	6870				1		
Arsenic Barium Calcium Copper Iron	10.5 39.6 14700 8.09 9100						
ead	4.25						
Magnesium	3480 291				1 1		
Manganese Potassium	2370						
Sodium	10800						
PESTICIDES/PCBS							
Chlordane - Gamma							1
Heptachlor Epoxide				the second se			
SVOCs							
2-Methylnaphthalene Acenaphthene Dibenzofuran Diethyl Phthalate Fluorene Naphthaleno Phenanthrene Bis(2-ethylhexyl) Phthalate							
VOCs							
1,1,1-Trichloroethane Acetone Chloroform Ethylbenzene Toluene Trichloroethylene Xylenes		5 U	5 U		5. U		s U
WET CHEMISTRY							
Alkalinity Chloride Nitrite, Nitrate-non Specific Nitrogen By Kjeldahl Method Phosphate Sulfate Total Dissolved Solids Total Hardness Total Suspended Solids			:	< 183 < 13.3 218000 89200000 < 4000		< 183 < 13.3 170000 73200000 < 4000	
OTHER.							
Fotal Organic Carbon				< 1000	1	< 1000 -	
Total Petroleum Hydrocarbons				< 167		< 178	

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Sample Key:

			Flagging
			Codes (Lower
Manager Dealing	a	(Upper case letters)	case letters or
Measurement Boolean	Concentration	letters)	<u>#)</u>
ND	300	ī.	adf

Measurement Boolean

<= Concentration was less than the certified reporting limit ND = Not detectable above the indicated value

EQ = Equal to the certified reporting limit

Data Qualifiers

? = Control chart for corresponding lot not yet reviewed by AEC Chemist. This qualifier is automatically set when a lot file

- has been uploaded to the database, but a corresponding control chart has not been approved.
- 1 = The low spike recovery for this lot was high
- M = The high spike recovery for this lot was high
- J = The low spike recovery for this lot was low
- K = Missed holding time for extraction or preparation
- L = Missed analysis holding time
- N = The high spike recovery for this lot was low
- O = Low spike recoveries excessively different
- R = Data is rejected and is not useable

Flagging Codes

1 = Result was less than the certified reporting limit but greater than the criteria of detection (COD) for 1990 QA Plan methods

2 = Ending calibration not within acceptable limits

3 = Internal standard not within acceptable limits

- 7 = Low spike recovery not within control limits
- 8 = Analyte recovery outside certified range but within acceptable limits. This code is used when analyte concentrations
- exceeded the certified range by <15 % and the laboratory felt a dilution was not warranted
- a = Analyte found in trip blank as well as the sample
- b = Analyte found in method blank or QC sample as well as the sample.

c = Analysis was confirmed by a different column or technique.

- d = Duplicate analysis
- f = Sample was filtered prior to analysis

g = Analyte found in that day's rinsate blank as well as the sample

h = Lot out of control but data accepted due to high recoveries

i = Interences in the sample caused the quantitation and/or identification to be suspect

j = Value is estimated

k = Reported results affected by interferences or high background. An elevated quantitation limit is reported

- 1 = Out of control. Data rejected due to low recoveries
- m= High duplicate spike not within control limits

n = Tentatively-identified compound (TIC) by GC/MC with a match greater than 70 %

- p = Value is less than the method reporting limit but greater than the instrument detection limit
- q = Confirmatory analysis was performed, however sample interferences prevented confirmation
- r = Non-target analyzed for but not detected by GC/MS. Laboratory is not certified for this analyze by the given method

Analyte was not performance demonstrated or validated

- s = Non-target compound analyzed for and detected by GC/MS. Laboratory is not certified for this anlayte by the given method.
- Analyte was not performance demonstrated or validated
- t = Non-target compound analyzed for and not detected (non-GC/MS method).
 u = Analysis is unconfirmed. Confirmatory analysis was run but did not verify
 - original result
- v = Sample was not correctly preserved (i.e. > 4 degrees C or improperly preserved)
- z = Non-target analyzed for and detected by non-GC/MS method

TABLE 7 - 16 EPH – VPH GROUNDWATER OFF-SITE LABORATORY RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:	69W94-09 MXZW0904 12/18/97 8.5	69W MXZW 10/02 9.5	10X5 2/96	69W94-10 MXZW10X6 12/11/97 9.5		69W94-11 MXZW11X5 10/01/96 11	69W94-11 MXZW11X6 12/11/97 11	69W94-12 MXZW12X5 12/19/97 8
EPH (µg/L)			740		<	2.5		
ALIPHATICS					1			
n-C9 to n-C 18	37 UJ		590	32 J	<	25	75 J	39 UJ
n-C 19 to n-C 36	49 U	<	2.5	41 U	<	2.5	38 U	52 U
AROMATICS								
n-C 11 to n-C 22	100 U		710	480	<	500	84	110 U
TARGETED PAH ANALYTES (µg/L)								
2-Methylnaphthalene	6.2 UJ		89	81 J	<	10	4.8 UJ	6.5 UJ
Acenaphthene	6.2 U		15	5.7	<	10	4.8 U	6.5 U
Acenaphthylene	6.2 U	<	10	6.9	<	10	4.8 U	6.5 U
Anthracene		<	10		<	10	000	
Benzo(a)anthracene		<	10		<	10		
Benzo(a)pyrene		<	10		<	10		
Benzo(b)fluoranthene		<	10		<	10		
Benzo(g,h,i)perylene		<	10		<	10	-	
Benzo(k)fluoranthene		<	10		<	10		
Chyrsene		<	10		<	10		
Dibenzo(a,h)anthracene		<	10		<	10		
Fluoranthene	6.2 U	<	10	5.1 U	<	10	8.1	6.5 U
Fluorene		<	10		<	10		
Indeno(1,2,3-c,d)pyrene		<	10		<	10		
Naphthalene	6.2 UJ	The second second	45	37 J	<	10	4.8 UJ	6.5 UJ
Phenanthrene	00.00	<	10		<	10		. eff. (e.e
Pyrene		<	10		<	10		
VPH (µg/L)			830		<	0.25		
ALIPHATICS								
n-C5ton-C8			17		<	2.5		
n-C9ton-C12	65 U		550	120	<	0.25	65 U	65 U
AROMATICS								
n-C9ton-C10	20 U	1	790	430	<	5	20 U	20 U
TARGETED VOCs (µg/L)								
Benzene		<	5		<	5		
Ethylbenzene	5 U	15	35	15	<	5	5 U	5 U
Methyl tert-butly Ether		<	25		<	25		
Naphthalene	10 U		94	100	<	10	10 U	10 U
Toluene		<	5		<	5	100 Ca.	
m/p-Xylene		<	5		<	5		
o-Xylene		<	5		<	5		

Notes:

J = Estimated value, below quantitation limit. U = Compound was not detected above the method detection limit shown.

1

TABLE 7 - 16 EPH – VPH GROUNDWATER OFF-SITE LABORATORY RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:	69W94-13 MXZW13X5 12/12/97 8	69W94-14 MXZW14X5 12/12/97 8	ZWM-95-15X MXZW15X3 12/12/97 8		ZWM-95-16X MXZW16X3 10/01/96 16	ZWM-95-16X MXZW16X4 12/18/97 16	ZWM-95-17X MXZW17X3 12/16/97 17.2
EPH (µg/L)				<	2.5		
ALIPHATICS				-			104 July 104
n-C9 to n-C18	43 J	72 J	84 J	<	25	37 UJ	35 UJ
n-C 19 to n-C 36	41 U	44 U	44 U	<	2.5	49 U	47 U
AROMATICS							
n-C 11 to n-C 22	210	93 U	93 U	<	500	100 U	100 U
TARGETED PAH ANALYTES (µg/L)			_				
2-Methylnaphthalene	19 J	5.5 UJ	5.5 UJ	<	10	6.1 UJ	5.9 UJ
Acenaphthene	5.2 U	5.5 U	5.5 U	<	10	6.1 U	5.9 U
Acenaphthylene	5.2 U	5.5 U	5.5 U	<	10	6.1 U	5.9 U
Anthracene	100.00			<	10		2.5 0
Benzo(a)anthracene				<	10		
Benzo(a)pyrene				<	10		
Benzo(b)fluoranthene				<	10		
Benzo(g,h,i)perylene				<	10		
Benzo(k)fluoranthene				<	10		
Chyrsene				2	10		
Dibenzo(a,h)anthracene				2	10		
Fluoranthene	5.2 U	8.3	5.5 U	1.5		6.1 U	CO.11
Fluorene	5.2 0	0.5	5.5 0	<	10	0.1 U	5.9 U
				<	10		
Indeno(1,2,3-c,d)pyrene				<	10		
Naphthalene	8.5 J	5.5 UJ	5.5 UJ	1.5	10	6.1 UJ	5.9 UJ
Phenanthrene		and the second se		<	10		
Pyrene				<	10		
VPH (µg/L)				-	7		
ALIPHATICS				-			
n-C5 to n-C8				<	2.5		
n-C 9 to n-C 12	140	65 U	65 U	<	0.25	65 U	65 U
AROMATICS							
n-C9 to n-C 10	330	20 U	20 U		7	75	20 U
TARGETED VOCs (µg/L)				1			
Benzene				<	5		
Ethylbenzene	5 U	5 U	5 U	<	5	5 U	5 U
Methyl tert-butly Ether				<	25		
Naphthalene	26	10 U	10 U	<	10	10 U	10 U
Toluene	and an			<	5		201.0
m/p-Xylene				<	5		
o-Xylene				<	5		

J = Estimated value, below quantitation limit. U = Compound was not detected above the method detection limit shown.

2

TABLE 7 - 16 EPH – VPH GROUNDWATER OFF-SITE LABORATORY RESULTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:	ZWM-95-18X MXZW18X3 12/16/97 8		ZWM-96-19X MDZW19X1 09/30/96 19		ZWM-96-19X MXZW19X1 09/30/96 19	ZWM-96-19X MXZW19X2 12/18/97 19	ZWM-96-19X MDZW19X2 12/18/97 19	ZWM-96-20X MXZW20X2 12/16/97 0
EPH (µg/L)		<	2.5	<	2.5			
ALIPHATICS				1				
n-C9 to n-C 18	37 UJ	<	25	<	25	39 UJ	33 UJ	38 UJ
n-C 19 to n-C 36	49 U	<	2.5		2.5	53 U	44 U	51 U
AROMATICS		1		1				
n-C 11 to n-C 22	100 U	<	500	<	500	110 U	93 U	110 U
TARGETED PAH ANALYTES (µg/L)				-				
2-Methylnaphthalene	6.2 UJ	<	10	<	10	6.6 UJ	5.5 UJ	6.3 UJ
Acenaphthene	6.2 U	<	10	<	10	6.6 U	5.5 U	6.3 U
Acenaphthylene	6.2 U	<	10	<	10	6.6 U	5.5 U	6.3 U
Anthracene		<	10	<	10		1000	1.11
Benzo(a)anthracene		<	10	<	10			
Benzo(a)pyrene		<	10	<	10			
Benzo(b)fluoranthene		<	10	<	10			
Benzo(g,h,i)perylene		<	10	<	10			
Benzo(k)fluoranthene		<	10	<	10		1	
Chyrsene		<	10	<	10			
Dibenzo(a,h)anthracene		<	10	<	10			
Fluoranthene	6.2 U	<	10	<	10	6.6 U	5.5 U	6.3 U
Fluorene	100 1	<	10	<	10			
Indeno(1,2,3-c,d)pyrene		<	10	<	10			
Naphthalene	6.2 UJ	<	10	<	10	6.6 UJ	5.5 UJ	6.3 UJ
Phenanthrene		<	10	<	10			
Pyrene		<	10		10			
VPH (µg/L)		-	42		47			
ALIPHATICS		-					-	
n-C5 to n-C8		<	2.5	<	2.5			
n-C9 to n-C 12	65 U		28		34	65 U	65 U	65 U
AROMATICS				-				45 0
n-C9 to n-C 10	20 U	-	41	-	45	20 U	20 U	20 U
TARGETED VOCs (µg/L)		-		-				
Benzene		<	5	<	5			
Ethylbenzene	5 U	<	5	<	5	5 U	5 U	5 U
Methyl tert-butly Ether		<	25	<	25			
Naphthalene	10 U	<	10	<	10	10 U	10 U	10 U
Toluene		<	5		5			
m/p-Xylene		<	5	<	5		100	
o-Xylene		<	5		5			

Notes:

J = Estimated value, below quantitation limit. U = Compound was not detected above the method detection limit shown.

TABLE 7 – 16EPH – VPH GROUNDWATER OFF–SITE LABORATORY RESULTSAOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Site ID: Field Sample Number: Sample Date: Depth:		ZWM-96-21X MXZW21X1 10/01/96 21	ZWM-96-21X MXZW21X2 12/18/97 21
EPH (µg/L)	<	2.5	
ALIPHATICS			
n-C9 to n-C 18	<	25	38 UJ
n-C 19 to n-C 36	<	2.5	51 U
AROMATICS			
n-C 11 to n-C 22	<	500	110 U
TARGETED PAH ANALYTES (µg/L)			
2-Methylnaphthalene	<	10	6.3 UJ
Acenaphthene	<	10	6.3 U
Acenaphthylene	<	10	6.3 U
Anthracene	<	10	214.17
Benzo(a)anthracene	<	10	
Benzo(a)pyrene	<	10	
Benzo(b)fluoranthene	<	10	
Benzo(g,h,i)perylene	<	10	
Benzo(k)fluoranthene	<	10	
Chyrsene	<	10	
Dibenzo(a,h)anthracene	<	10	
Fluoranthene	<	10	6.3 U
Fluorene	<	10	1.00
Indeno(1,2,3-c,d)pyrene	<	10	
Naphthalene	<	10	6.3 UJ
Phenanthrene	<	10	
Pyrene	<	10	
VPH (µg/L)	<	0.25	
ALIPHATICS			
n-C5 to n-C8	<	2.5	
n-C9 to n-C 12	<	0.25	65 U
AROMATICS			
n-C9 to n-C 10	<	5	20 U
TARGETED VOCs (µg/L)			
Benzene	<	5	
Ethylbenzene	<	5	5 U
Methyl tert-butly Ether	<	25	
Naphthalene	<	10	10 U
Toluene	<	5	
m/p-Xylene	<	5	
o-Xylene	<	5	

Notes:

J = Estimated value, below quantitation limit. U = Compound was not detected above the

method detection limit shown.

15

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETS

Site ID: Field Sample Number; Lab Sample Number; Sample Date; Depth; Units;	ZWD-95-81X DDZW0108 DV45*400 89/11/95 0 9/2/g		ZWD-95-01X DXZW0100 DV45*289 09/11/95 0 pg/g		ZWD-95-01X DXZW0102 DV45*390 09/11/95 2 pg/g		ZWD-95-02X DXZW0200 DV45*391 09/11/95 0 168/2		2WD-95-02X DXZW0202 DV45*392 09/11/95 2 µ8/8		ZWD:95-03X DXZW0300 DV45*393 09/11/95 2 JE/E		ZWD-95-03X DXZW0302 DY45*394 09/11/95 0 μg/g		ZWD-95-047 DXZW0400 DV4S*395 09/12/95 0 µg/g	1	ZWD-95 DXZW0 DV45* 09/12/ 0	500 196		ZWD-95-86X DXZW0600 DV4S*397 09/12/95 0 µg/g
PAL METALS Aluminum	6010	-	5240	-	5600	-	4120	1	6270	1	4840	-	4480	1	3390		3120		-	
Arsenio	8.42	D	9.95	1	21		13		8.64	1	6.39	1	5.46		9.43		6.43		100	2930 14
Barium	18.6	D	14.7		15.5		11.4	11	22.9	1.	10.7		10.3		9.52		12.9			7.13
Calcutto	392	D	957	1 3	1050		736	1.1	1500		570		527		674		588	12222		536
Chromemo	26.7	D	35.5	112	15.4		16.1		12.3	1	11.5		11:2		10.5		10.5	precioner		13.9
Cobalt	417	D	3.56		\$.87		3.85		3.15		2.82		2.23		2.55		2.02			6.93
Copper	12.8	D	251		8.43		10.9		8.79	P	10.4		6.56		7.92		165			23.4
tron	15400	D	12400		12800		10900	1	5980	<u> </u>	8030		7010		7020		7150			10200
Lead	35.9	D	58		712		21		29		15		11.4		29	- 1	33			30
Magnesium	3100	D	2810		2630		2630		1740		2110		2160		1750		1670			1580
Manggemean	230	D	172		475		161	1	81.4		93.9		70.7		93.6		96.6			186
Nicket	173	D	15.9		18.1		18.1	1	9.97	1	11.7		10.5		9.95		8.72			9.55
Polassium	783	D	525		639		415	1.	388		426		425		371	10	486		1	364
Sodern	314	D	3DGE		322		259	1.1	462	1.8	307		307	1 3	309	8	312			260
Vanadium	15.2	D	12.5		10.7		10.4		8.95		9.16		7.91		7,4		8.28			8.49
Cinst .	1997年1月1日日日	D	69.5		29 B		39.6	1.	40		27.9		22.8		41.4	_	39 7			31
PESTICIDES/PCBS		-		-		-		-		-		-		-		_			1	
4,4'-ddd	0.15	C	0.11 C 0.0137 C	1	0.00808 C		0.0174 (<	0.00826	11-	0.12 0		0.027 C	1	0.027	С	2.1	C		0.066 0
4,4'-dde	0.0172	C			0.00765	<	0,00765	<	0.00765	<	0.00765	<	0.00765		0.0122	C	0.08	C	1	0.0152 (
4,4'-ddt	0.14	C	0.094 C		0.00707		0.021 (<	0.00707	1.5	0.025 C		0.00707		0.044	C	0.4	C	1.1	0.046 (
Chlordane - Alpha <			< 0.005 T		0.005 T	1.	0.005	1 <					0.005 T		0.005	T	0,012			0.005 1
Chlordane - Gamma			< 0.005 T	<	0.005 T	1000	0.005	r <	0,005 T	<			0.005 T		0.005	T	0.02	C2		0.005 1
Dieldrin	0.00629		< 0.00629	<	0,00629	<	0.00629	<	1.105.25.445	<	0.00629	<	0.00629	<	0.00629		0,06	C		0.00629
Endosulfan li 🤟	0.00663	12.2	< 0.00663	<	0.00663	<	0.00663	<		<	0.00663	<	0.00663	<	0.00663		0,05	C		0.00663
Peb 1260 <	0.0804	D	< 0.0804	<	0 0804	<	0.0804	<	0.0804	<	0.0804	<	0 0804	<	0.0804	-	0.51	C	~	0.0804
SEMIVOLATILE ORGANICS	-	-				-		1		<u>.</u>		-		-	-	_			-	
Anthracene <	0.2		< 0.2	<	0.2	<	0.2	1	0.7	<	0.2	<	0.2	<	0.3	- 1	< 0.8		<	0.2
Benzo[a]anthracene <	0.8		< 0.8	<	0.8	<	0.8	10	2	<	0.8		1	<	2		< 4		<	0.8
Benzo[a]pyrene <	1		< 1	<	1	<	1		3	<	1	5	1	<	2		< 6		<	1
Benzo(b)fluoranthene <		-	< 1	<	1	<	1		3	<	1	5	1	<	2		< 5		<	1
Benzo[k]fluoranthene <	0.3	~	< 0.3	<	0.3	<	0.3	<	0.3	15	0.3	<	0.3	1.0	0.8		< 2		1.1	0.4
Chrysene	0.9	D	1	<	0.6	<	0.6	1.1	4	<	0.6	10	2	<	1		< 3		<	0.6
Fluoranthene	1	D	2		1	<	0.3		6		1		3		2		< 2			L
Phenanthrene	0.6	DD	0.8		0.7	14	0.2		3		0.5		2		0.6		< 0.8			0.9
Рутепе	1	0	2	1.1	63	5	0.2		6		1	1	3		4	- 1	1		1	1
TPH BY GC		-	-	_		-		-		-	- Company of a				-	-			_	
Diesel	935	D	23,7		8.98	<	8	1	49.2	<		-	185		19.6	1	52.2			12.6
Tphgas <	8	D	< 8	<	8	<	8	<	g	<	8	<	g		21.5	1	16.5			19.4
VOLATILE ORGANICS		-		1		-		-		1	and here and	-		1		-			<u> </u>	-
Dichloromethane	0.014	D	< 0,012	<	0.012	<	0.012	<		<	0.012	<	0.012	<	0.012	1	< 0.012	y	<	0.012
Trichlorofluoromethane	0.011	D	0.011		0.014		0.0096		0.0087		0.0094		0.0089		0.008		< 0.005	9		0.0082
OTHER				-		-		1		1		-		I						
Total Organic Carbon	7420	D	12400	1	1560		2400	T	59200	T	5430	1	5170	1	3430		5950		1	3100
Total Petroleum Hydrocarbons	360	D	896		97		132	10.0	131		190		66.8		368	- C - 1	1230			287

Notes: U = Concentrations Less than Reporting Limit J = Value is Estimated E = Concentration Exceeds the Maximum Reporting Limit NA = Nat Analyzed

Shading indicated exceedance of established Devens background concentrations

TA 2 7-18 RI AIR OFF-SITE ANALYTICAL RESULTS AOC 69W

		1	DEVENS, MA	SSACHUSET	TS			
Site ID: Canister: Location: COMPOUND (µg/m ³)	ZWA-97-01X A230 Classroom N of new boiler room	ZWA-97-02X A209 Kitchen/ Cafeteria	ZWA-97-03X A210 Ist Classr'm NE Wing	ZWA-97-04X B239 -03X Duplicate	ZWA-97-05X 70 Crawl space near old boiler room	ZWA-97-06X B237 Crawl space under kitchen	ZWA-97-07X B246 SW classr'm (indoar backgrnd)	Inhalation Toxicity Value (ug/m3)
1,2,4 Trimethylbenzene	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 R	< 4.4 J	< 4.4 J	< 4.4 J	180
1,3,5 Trimethylbenzene	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 R	< 4.4 J	< 4.4 J	< 4.4 J	180
2-Methylheptane	8	< 4.4	19	6.3 R	< 4.4	< 4.4	7.2	200
3-Methylheptane	< 4.4	< 4.4	8.7	< 4.4 R	< 4.4	< 4.4	8.9	200
Decane	< 4.4	< 4.4 J	< 4.4 J	< 4.4 R	< 4.4 J	< 4.4	< 4.4 J	2000
Dodecane	< 22 J	< 22	< 22	< 22 R	< 22	< 22 J	< 22	2000
Ethylbenzene	7.9	470	27	7.6 R	< 4.4	5.2	9.9	1000
Methyl tert-butly Ether	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 R	< 4.4 J	< 4.4 J	< 4.4 J	3000
Nonane	< 4.4	< 4.4	7.2	< 4.4 R	< 4.4	< 4.4	5	2000
Octane	< 4.4	< 4.4	21	8.4 R	< 4.4	5.5 J	9.1	200
Toluene	260	72	1000	350 R	< 13	< 150	< 36	400
Xylene	30.4	< 8.8	92	28.1 JR	< 8.8	< 18.3 J	34.8	310
Acetone	470	< 200	< 82	< 54 R	< 38	< 81	< 30	-
Tetrachloroethylene	< 4.4	< 4.4	< 4.4	< 4.4 R	< 4.4	< 4.4	< 4.4	-
Site ID; Canister: Location;	ZWA-97-08X A211 Classe'm S of new	ZWA-97-09X 92 Boiler room well head	ZWA-97-10X 62 NW classroom	ZWA-97-11X A207 Front parking lot downwind	ZWA-97-12X A221 Near wells upwind	ZWA-97-13X B233 Playground upwind	Inhalation Toxicity Value (ug/m3)	
COMPOUND (ug/m ³)	baller room	sample		(bkgrnd)	(bkgrnd)	(backgrnd)	(
1,2,4 Trimethylbenzene	< 4.4 J	< 22	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 J	180	
1,3,5 Trimethylbenzene	< 4.4 J	< 22	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 J	180	
2-Methylheptane	< 4.4	< NA	5.2	< 4.4	< 4.4	< 4.4	200	
3-Methylheptane	< 4.4	< NA	< 4.4	< 4.4	< 4.4	< 4.4	200	
Decane	< 4.4	< NA	< 4.4	< 4.4	< 4.4	< 4.4	2000	
Dodecane	< 22 J	< NA	< 22 J	< 22 J	< 22 J	< 22 J	2000	
Ethylbenzene	< 4.3 J	< 22	< 2.8 J	< 3.2 J	< 4.4	< 4.4	1000	
Methyl tert-butly Ether	< 4.4 J	< NA	< 4.4 J	< 4.4 J	< 4.4 J	< 4.4 J	3000	
Nonane	< 4.4	< NA	< 4.4	< 4.4	< 4.4	< 4.4	2000	
Octane	< 4.4	< NA	< 4.4	< 4.4	< 4.4	< 4.4	200	
Toluene	70	< 22	82	63	< 38	< 19	400	-
Xylene	17.1 J	< 44	8	< 8.2	< 8.8	< 8.8	310	
Acetone	< 54	< 210 J	< 52	< 440	< 27	< 31		
Tetrachloroethylene	< 4.4	360	< 4.4	< 4.4	< 4.4	< 4.4		

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Notes: <= Undetected at specified detection limit

B - Analyte detected in quality control blank

J - Estimated value, below detection limit

NA - Not analyzed

TABLE 8-1 CHEMICAL AND PHYSICAL PROPERTIES OF COMPOUNDS DETECTED¹ AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

CONTAMINANT	MOLECULAR WEIGHT	DENSITY (g/cm ³)	WATER SOLUBILITY (mg/L)	VAPOR PRESSURE (mm Hg)	HENRY'S CONSTANT, H _c (atm m ³ mol ⁻¹)	ORGANIC CARBON PARTITION COEFFICIENT K _m (mL/g)
VOLATILE ORGANIC COM	POUNDS					
1,2-DICHLOROETHANE	98.98	1.28	8.52E+03	6.40E+01	9.78E-04	1.40E+01
BENZENE	78.12	1.87	1.75E+05	9.52E+01	5.59E-03	8.30E+01
CARBON TETRACHLORIDE	153.82	1.58	7.57E+02	9.00E+01	2.41E-02	4.39E+02
ETHYLBENZENE	106.16	0.87	1.52E+02	7.00E+00	6.43E-03	1.10E+03
TOLUENE	92.13	0.86	5.35E+02	2.81E+01	6.37E-03	3.00E+02
XYLENES ²	106.16	0.88	1.75E+02	6.60E+00	5.10E-03	8.30E+02
SEMIVOLATILE ORGANIC	COMPOUNDS)		
1,2-DICHLOROBENZENE	1.4701	1.3	1.00E+02	1.00E+00	1.70E-03	1.70E+03
2-METHYLNAPHTHALENE	142.2	1.02	2.54E+01	NA	NA	8.50E+03
o-CRESOL	108.14	1.03	2.50E+04	2.43E-01	1.50E-06	NA
P-CRESOL	108.14	1.01	NA	1.14E-01	NA	NA
NAPHTHALENE	128.2	0.98	3.17E+01	2.30E-01	1.15E-03	1.30E+03
PHENANTHRENE	178.2	1.03	1.00E+00	6.80E-04	1.59E-04	1.40E+04
PYRENE	202.3	1.27	1.30E-01	2.50E-06	5.04E-06	3.80E+04

Notes:

 Primary Source: USEPA, 1990. "Basics of Pump-and-Treat Ground-Water Remediation Technology." Robert S. Kerr, Environmental Research Laboratory, Ada, OK. EPA/600/8-90/003.

2. Data presented is for o-xylene.

Aubrey, D.C., 1993. "Encyclopedia of Chemical Technology," Fourth Edition. John Wiley and Sons, New York, NY.

Verschueren, Karel, 1983. "Handbook of Environmental Data on Organic Chemicals." Van Nostrand Reinhold Company, Inc.

Howard, Philip, 1990. "Handbook of Environmental Fate and Exposure Data for Organic Chemicals; Volume II, Solvents." Lewis Publishers, Inc.

atm $m^3 \text{ mol}^{-1} = \text{cubic meter at an atmospher of pressure per mole}$ $g/cm^3 = \text{gram per cubic centimeter}$ $K_{oc} = \text{organic carbon partition coefficient}$ mg/L = miligram per liter mL/g = milliliters per gram mm Hg = pressure as millimeters of mercuryNA = Not available

TABLE 8-2 MOBILITIES OF INORGANIC ELEMENTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

		Environment	
Relative Mobility	Oxidizing (pH 5 to 8)	Oxidizing (pH <4)	Reducing
Highly mobile	Anions, B, Mo	Anions, B	Anions
Moderately mobile	Ca, Na, Mg, Sb, As, V, Zn, Be	Ca, Na, Mg, Pb, Zn, Cu, Hg, Co, Ni, V, As, Mn, Cd, Sb	Ca, Na, Mg, Ba, Mn, As*
Slightly mobile	K, Ba, Mn, Pb, Cu, Cd, Ni	K, Ba, Cr	K, Fe ²⁺
Immobile	Fe, Al, Cr, Hg	Fe, Al	Fe ³⁺ , Al, Cu, Zn, Pb, Cr, V, Ni, As*, Sb, Cd, Hg, Ba

*As⁺³ mobile under reducing conditions in absence of sulfides. As⁺⁵ immobile under reducing conditions.

As = Arsenic

Al = Aluminum

- B = Boron
- Ba = Barium
- Be = Beryllium
- Ca = Calcium
- Cd = Cadmium
- Co = Cobalt
- Cr = Chromium Cu = Copper
- Fe = Iron
- 16 100
- Hg = Mercury K = Potassium
- Mg = Magnesium
- Mn = Manganese
- Mo = Molybdenum
- Nd = Sodium
- Ni = Nickel
- Pb = Lead
- Sb = Antimony
- V = Vanadium

Z = Zinc

Source:

Rose, A.W., H.E. Hawkes, and J.S. Webb, 1979. Geochemistry in Mineral Exploration. Academic Press.

TABLE 9-1 SUMMARY OF HUMAN HEALTH RISK ASSESSMENT SAMPLE LOCATIONS AOC 69W

MEDIA	EXPLORATION	SAMPLE LOCATION	DEPTH (feet bgs)
Surface Soils	Remedial Investigation	ZWS-95-35X	0
		ZWS-95-39X	0
		ZWS-95-42X	0
		ZWS-95-45X	0
		ZWS-95-46X	0
		ZWS-95-47X	0
Subsurface Soil1	Remedial Investigation	69W-HS-ESW	10
	and Removal Action	69W-HS-FL-3	- 10
		69W-HS-FL-4	8
		69W-HS-NSW	6
		69W-HS-OB-1	6
		69W-HS-SSW	10
		69W-HS-SSW1	10
		69W-HS-WSW	10
		69W-HS-WSW	10
		69W-HS-WSW	8
		69W-HS-WSW	9
		69W-PL-ESW	8
		69W-PL-FL-5	10
		69W-PL-NSW	6
		69W-UP-G1-E	6
		69W-UP-G1-F	4
		69W-UP-G1-W	4
		69W-UP-G2-E	5
		69W-UP-G2-F	5
		69W-UP-G2-W	5
		69W-UP-G3-E	5
		69W-UP-G3-F	5
		69W-UP-G3-W	5
		69W-UST-ESW	10
		69W-V-ESW-1	8
		69W-V-E3W-1	7
		69W-V-NSW-1	8
		69W-V-WSW-1	7
		ZWB-96-03X	6
		ZWB-96-21X	4
		ZWB-96-21X	8
		ZWR-95-26X	0
		ZWR-95-45X	4
ediment	Downgradient	ZWD-95-02X	0
		ZWD-95-03X	0
		ZWD-95-06X	0
Groundwater2	Remedial Investigation		ROUND
	Rounds 1 (October 1995),	69W-94-09	1,2,4
	2 & 3 (February 1996 & October 1996)	69W-94-10	1,2,3,4
		69W-94-11	1,2,3,4
		69W-94-11 (dup)	1,2
		69W-94-12	1,2,4
		69W-94-13	1,2,4
		69W-94-14	1,2,4
		ZWM-95-15X	1,2,4
		ZWM-95-16X	1,2,3,4
		ZWM-95-18X	1,2,4
		ZWM-96-19X	2,3,4
		ZWM-96-19X (dup)	2,3,4
		ZWM-96-20X	2,3,4
		ZWM-96-21X	2,3,4
idoor Air	Supplemental Air Sampling	ZWA-97-01X	-
	copponent An Damping	ZWA-97-10X	
		ZWA-97-02X ZWA-97-08X	÷ -

TABLE 9-1 SUMMARY OF HUMAN HEALTH RISK ASSESSMENT SAMPLE LOCATIONS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	EXPLORATION	SAMPLE LOCATION	DEPTH (feet bgs)
10 A			(leet ugs)

1 Surface soil is defined as soil between 0 and 1 foot bgs. Subsurface soil is defined as soil between 1 and 10 feet bgs 2 Unfiltered data from each sample location were used to develop data sets bgs = below ground surface

DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN

AOC 69W

	Ranga	Frequency			Concentratio	n					
	of	of	Miniman	Maximum	Arithmetic	85%	Back	Region III			
	801.*	Dateotion	Detected	Detected	Mean	UCL	ground*	RBC**	ARARa	CPC?	Rictae
SURFACE SOIL (0 -	1 feet bgs)* (mg/	kg)									
PAL METALS											
Aluminum	NA	6 /	6 5210	6160	5916.667	NC	18000	7800	NA	No	Less than RBC ¹ , Background ²
Arsenia	NA	6 /	6 7.66	18	12.0383	NC	19	0.43	NA	Yes	Exceeds RBC ² , Beckground ²
Barium	NA	6 /	6 14.1	22.4	18.2	NC	54	550	NA	No	Less than RBC ¹ , Background ²
Seryflium	0.50-0.50	1 1	6 0.85	0.85	0.35	NC	0.81	0.15	NLA	Yes	Exceeds RBC ⁻²
Calcium	NA	6 /	6 333	908	683,1667	NC	810	NA	NA	No	Essential Nutrient ⁴
Chromium	NA	6 /	6 12.1	28.1	18.0167	NC	33	39	NA	No	Less than RBC ¹ , Background ²
Cobalt	NA	6 /	6 2.51	5.36	4.1283	NC	4.7	470	NA	No	Less then RBC ¹
Copper	NA	6 /	6 5.59	29.9	11.7867	NC	13.5	310	NA	No	Less than RBC ¹
Iron	NA	6 /	6 6780	10300	8818.333	NC	18000	2300	NA	Yas	Exceeds RBC ² , Background ²
Lead	NA	5 /	6 11.4	238	71.1	NC	61.1	NA	400 s	No	Less than ARAR ⁶
Magnesium	NA	6 /	6 1360	2670	2405	NC	5500	NA	NA	No	Essential Nutrient ⁴ , Background ²
Manganosa	NA	8 /	6 52.4	240	167.4	NC	380	180	NA	Yeo	Exceeds RBC ² , Background ⁴
Mercury	0.050-0.050	2 /	6 0.0755	0.0784	0.0423	NC	NA	2.3	NA	No	Less than RBC ¹
Nickel	NA	6 /	6 5.98	18.1	13.3133	NC	14.6	160	NA	No	Less than RBC ¹
Potassium	NA	6 /	6 367	993	630.1667	NC	2400	NA	NA	No	Background ² , Essential Nutrient ⁴
Selenium	0.25-0.25	1.7	6 0.364	0.364	0.1648	NC	ND	39	NA	No	Less than RBC ¹
Sodium	NA	6 /	6 241	506	347.5	NC	131	NA	NA	No	Essential Nutrient ⁴
Vanadium	NA	6 /	6 10.6	19.1	14.0667	NC	32.3	55	NA	No	Less than RBC ¹ , Background ²
Zinc	NA	6 /	6 18.9	71.7	32.4833	NC	43.9	2300	NA	No	Less than RBC'
PAL SEMIVOLATILE ORGANI	ICS										
Acenaphthylene	0.033-3	1 /	6 2	2	0.7055	NC		310 h	NA	No	Less than RBC ¹
Anthracene	0.033-3	11	6 1	1	0,5388	NC		2300	NA	No	Less then RBC ¹
Benzo[k]fluoranthene	0.066-7		6 2	2	1.0943	NC		8.8	NA	No	Less then RBC ¹
Chrysene	0.12-10	21	6 0.17	5		NC		88	NA	No	Less then RBC ¹
Fluoranthene	0.068-1	4 /	6 0.19	9	3.2873	NC	-	310	NA	No	Less than RBC ¹
Fluorene	0.033-3	17	6 1	1	0,5388	NC	1	310	NA	No	Less than RBC ¹
Phenanthrene	0.20-0.70	5 /	6 0.065	. 9	3.0925	NC		310 h	NA	No	Less than RBC ¹
Ругеле	0.20-0.70	5 /	6 0.075	10	3.7742	NC		230	NA	No	Less than RBC ¹
PAL VOLATILE ORGANICS				*							
Acetone	0.017-0.017	1 /	6 0.069	0.069	0.0186	NC		780	NA	No	Less than RBC'
Toluene	0.00078-0.0007		6 0.001	0.0021	0.0009	NC			NA	No	Less than RBC ¹
Trichlorofluoromethane	0.0059-0.0059		6 0.0055	0.0072		NC			NA	No	Less than RBC ¹
Xylenes	0.0015-0.0015		6 0.0027	0.0027		NC		16000	NA	No	Less than RBC ¹

DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN

AOC 69W

	Range	Frequency			Concentratio	n					
	of	of	Minimum	Maximum	Arithmetic	95%	Beck	Region III			
	SQL	Detection	Detected	Detected	Mean	UCL	ground*	RBC**	ARARa	CPG?	Notee
SURFACE SOIL (0 - 1 f	eet bgs)* (m	g/kg) - CONT	NUED								
OTHER				-	TANDO CONTRACTOR						
Fotal Patroleum Hydrocarbona	28-28	516	52.5	936	390.375	NC	•	NA	NA	Yes	No standard available
SUBSURFACE SOIL (1	 10 feet bgs 	i) ^b (mg/kg)									
PAL METALS											
Aluminum	NA	2/ 2	2910	3060	2985	NC	18000	7800	NA	No	Lees than RBC ¹ , Background ²
Arsenia	NA	21 2	man and the second	7.32	5.03	NC	19	0,43	NA	Yes	Exceeds RBC ³ , Background ²
Berium	NA	2/2	Accession for the second	8.21	8.175	NC	54	550	NA	No	Less than RBC ¹ , Background ²
Calcium	NA	21 3	369	463	416	NC	810	NA	NA	No	Essential Nutrient ⁴ , Background ²
Chromium	4.1-4.1	1/3	10.3	10.3	6.1625	NC	33	39	- NA	No	Less than RBC ¹ , Background ²
Cobalt	NA	2/ 1	2.22	2.98	2.55	NC	4.7	470	NA	No	Less than RBC ¹ , Background ²
Copper	NA	2/ 3	4.6	5.14	4.87	NC	13.5	310	NA	No	Less than RBC ¹ , Background ²
lion	NA	2 / 2	5460	5880	5870	NG	18000	2300	NA	Yes	Exceeds RSC ⁰ , Background ²
Lead	NA	2 / 3	1.87	1.91	1.89	NC	48	NA	400 6	No	Lese than ARAR ⁵ , Background ²
Magnesium	NA	2 / 1	1090	1430	1260	NC	5500	NA	NA	No	Essential Nutrient ⁴ , Background ²
Manganese	NA	2/ 3	56,4	90.3	73.35	NC	380	180	NA	No	Less than RBC ¹ , Background ²
Nickel	NA	2 / 3	8.26	8.57	8.415	NC	14.6	160	NA	No	Less than RBC ¹ , Beckground ²
Potaesium	NA	2 / 2	460	515	487.5	NC	2400	NA	NA	No	Essential Nutrient ⁴ , Background ²
Sodium	NA	2/ 3	299	398	348.5	NC	131	NA	NA	No	Essential Nutrient ⁴ ,
Vanadium	NA	2 / 3	4.5	6.47	5.485	NC	32.3	55	NA	No	Less than RBC ¹ , Background ²
Zine	8.0-8.0	1/:	2 14	14	9.0075	NC	43.9	2300	NA	No	Less than RBC ¹ , Background ²
PAL SEMIVOLATILE ORGANICS											
2-Methylnaphthalene	0.51-0.7	4 / 30	1.9	42	3.1797	2.858		310 h	NA	No	Less than RBC ¹
Aconaphthone	0.51-0.7	5 / 30	0.79	7.6	0.9312	1	-	470	NA	No	Less than RBC ¹
Acenaphthylene	0.06-0.7	2 / 30	9.6	16	1.1142	0.98	-	2300	NA	No	Less than RBC ¹
Benzo(a)anthracene	0.07-0.7	1 / 30	0.1	0.1	0.2655	0.35		0.88	NA	No	Less than RBC ¹
Benzo (b) fluoranthene	0.07-0.7	1 / 30	0.06	0.06	0.2642	0.354		0.88	NA	No	Less then RBC ¹
Chrysene	0.51-0.7	3 / 30	0.08	80.0	0.2652	0.347		88	NA	No	Less than RBC'
Fluoranthene	0.06-0.7	2 / 30	0.13	0.24	0.2732	0.333		310	NA	No	Less than RBC ¹
Fluorene	0.51-0.7	5 / 30	0.68	26	1.9132	1.584		310	NA	No	Less than RBC ¹
Naphthalene	0.51-0.7	3 / 30	7.1	12	1.1798	1.15	*	310	NA	No	Less than RBC ¹
Phenanthrene	0.51-7	3 / 3) 1.5	9	0.8707	0.932		310 h	NA	No	Less than RBC'
Pyrene	0.06-0.7	2 / 30	0.18	0.18	0.2815	0.34		230	NA	No	Less then RBC ¹

DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN

AOC 69W

	Benge	Frequency				Concentration	1					
	af	ot	,	- Minimum	Maximum	Arithmetic	95%	Back-	Region III			
	SQL	Detection		Detected	Detected	Mean	UCL	ground®	RBC**	ARAR	CPC?	Notae
PAL VOLATILE ORGANICS												
Acetone	0.017-0.017	1 /	2	0.022	0.022	0.0153	NC	2	780	NA	No	Less than RBC ¹
Dichloromethane	0.012-0.012	1/	2	0.025	0.025	0.0155	NC					
Toluene	0.0008-0.0008	1/	2	0.0013	0.0013	0.0008	NC		1600	NA	No	Less then RBC ¹
SUBSURFACE SOIL (1 - 10 feet bgs)	^b (mg/kg)	- CO	NTINU	D							
OTHER												
fotal Patroleum Hydrocarbons	28-28	2 /	5	57.5	902	27.8	NC		NA	NA	Yes	No standard available
Extrectable Petroleum Hydrocart	one (EPH)	and the second	(Alexande			00 2 00000						a la serie de la construcción de la
C11-C22 Aromatics	8.9-34	24 1	30	9	1,200	138	268		NA	NA	¥66	No standard available ²
C19-C36 Aliphatics	0.15-4.6		30	5.4	670	119	1,998		NA	NA	Yes	No standard available ⁷
C9-C18 Aliphetics	1.5-3.8	28 1		\$.3	5,400	588	18,583		NA	NA	Yee	No standard available ²
Volatila Petrolaum Hydrocarbon												
C-9-C12 Aliphatios	0.01-670	12 /	30	3.8	770	82.9	1.261		NA	NA	Yes	No standard available ⁷
C9-C10 Aromatics	0.25-500	8 /	30	15	650	42.7	119		NA	NA	Yes	No standard available
GROUNDWATER * (r	igit, chilitie											
Aluminum	0,141-0,141	4 /	10	0.39	0.448	0.2	NC	6.87	3.7	0.06 g	Yes	Exceeds ARAR . Background *
Araunio	0.0025-0.0025	0 /	10	0.0052	0.19	0.04	NG	0.0105	0.000045	0.05 f	Yes	Exceeds RBC 3, Exceeds ABAR*
Barium	NA	10 /	10	0.0046	0.017	0.01	NC	0.0396	0.26	21	No	Less than RBC ¹ , Less than ARAR ⁵ , Background ²
Calcium	NA	10 /	10	15.5	25	20	NC	14.7	NA	NA	No	Essential Nutrient ⁴
Copper	NA	1 /	10	0.01	0.01	0.004	NC		1.5	1.3	No	Less than RBC ¹ , Less than ARAR ⁵ , Background ²
rati	0.0388.0.0388	9 /	10	0.44	26	5.2	NC	9,1	1.1	0.3 g	Yes	Exceeds RBC ", Exceeds ARAR"
Load	0.001 - 0.001	4 /	10	0.001	0.002	0.001	NC		NA	0.015	No	Loss than ARAR ^{5,} Background ²
									NA	NA	No	Essential Nutrient ⁴ , Background ²
Cr.es	NA	10 /	10	1.7	3.02	2.2	NC	3.48	INA	1965		
Magnesium Mangarese	NA	10 / 10 /		1.7 0.013	3.02	2.2 0.66	NC	0.291	0.084	0.05 g	Yes	Exceeds RBC . Exceeds ARAR
Magnesium	NA		10	Warman	NI65310	000000000000000000000000000000000000000	ababbb abblickede		Concernments of The		Yes No	Exceeds ABC ³ , Exceeds ARAR ⁶ Essential Nutrient ⁴
Magnesium Manganese Potassium		10 /	10 10	0.013	27	0.86	NC	0.291	0.084	0.05 g		rest of the second s
Magnesium Mangarees Potassium Sodium	NA NA	10 / 10 /	10 10	0.013 1.6	2.7 5.1	0.66 2.3	NC NC	0.291 2.37	0,084 NA	0.05 g NA	No	Essential Nutrient ⁴ Essential Nutrient ⁴
Magnesium Mangaress Potassium Sodium PAL SEMIVOLATILE ORGANICS	NA NA	10 / 10 /	10 10 10	0.013 1.6	2.7 5.1	0.66 2.3	NC NC	0.291 2.37	0,084 NA	0.05 g NA	No	Essential Nutrient 4
Magnesium Manganese	NA NA	10 / 10 / 10 /	10 10 10	0.013 1.6 23.5	2.7 5.1 38	0.86 2.3 29	NC NC NC	0.291 2.37	0,084 NA NA	0.05 g NA NA	No No	Essential Nutrient ⁴ Essential Nutrient ⁴ Exceeds RBC ⁴ Less than RBC ¹
Magnesium Mangaress Potassium Sodium PAL SEMIVOLATILE ORGANICS 2 Methylnephthalene (i) Aceruphthane (j)	NA NA 9.6017-0.0017	10 / 10 / 10 / 2 /	10 10 10 10 13	0.013 1.6 23.5 0.008	2.7 5.1 38 0.6	0.86 2.3 29 0.06	NC NC NC	0.291 2.37	0.084 NA NA	0.05 g NA NA	No No Yes	Essential Nutrient ⁴ Essential Nutrient ⁴ Exceeds RBC ²
Magnesium Mangaress Potassium Sodium PAL SEMIVOLATILE ORGANICS 2 Methylnephthalene (i)	NA NA 0.0017-0.0017 D.0017-0.05	10 / 10 / 10 / 2 / 1 /	10 10 10 10 13 10	0.013 1.6 23.5 0.008 0.01	2.7 5.1 38 0.5 0.01	0.86 2.3 29 0.05 0.004	NC NC NC NC NC	0.291 2.37	0.084 NA NA 0.15 h 0.22	0.05 g NA NA NA	No No Yas No	Essential Nutrient ⁴ Essential Nutrient ⁴ Exceeds RBC ⁴ Less than RBC ¹
Magnesium Mangaress Potassium Sodium PAL SEMIVOLATILE ORGANICS 2 Methylnephtholene (i) Aceraphthere (i) Bis(2-ethylnexylptholete (i)	NA NA 0.0017-0.0017 0.0017-0.05 0.0046-0.0048	10 / 10 / 10 / 2 / 1 / 4 }	10 10 10 10 10 13 10	0.013 1.6 23.5 0.008 0.01 0.0034	2.7 5.1 38 0.6 0.01 0.5	0.86 2.3 29 0.05 0.004 0.055	NC NC NC NC NC NC	0.291 2.37	0,084 NA NA 0.15 h 0.22 0,0048	0.05 g NA NA NA NA 0.006 f	No No Yas No Yes	Essential Nutrient ⁴ Essential Nutrient ⁴ Exceeds RBC ⁶ Less than RBC ¹ Exceeds ARAR ⁶

TABLE 9-2 DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Renge	Frequency			Concentration	n						
	of	of	Minimum	Maximum	Arithmetia	95%	Back-	Region III				
	SOL	Detection	Detected	Detected	Moon	UCL	ground*	RBC**	ARARe	CPC?	Notae	
luorene (j)	0,01-0.011	2 / 8	0.003	0.007	0.005	NC		0.15	NA	No	Less than RBC ¹	
lapititalene (i)	0.0005-0.0005	2 / 10	0.015	0.2	0.021	NC	· · · · ·	0.15	NA	Үен	Excende RBC ³	
Phonanthrone (i)	0.0005-0.0005	2 / 10	0.002	0.15	0.015	NC		0.15 h	NA	No	Less than RBC ¹	
PAL VOLATILE ORGANICS												
1,1,1-Trichlorosthans (k)	0.0005-0.0013	1 / 10	0.0015	0.002	0.00035	NC		0.079	NA	No	Less than RBC ¹	
Acetone (k)	0.013-0.036	2 / 10	0.013	0.014	0.009	NC	-	0.37	NA	No	Less then RBC ¹	
Chloratarm (k)	0.0005-0.0013	2 / 10	0.00055	0.00055	0.00034	NC	•	0.00015	NA	Yes	Exceeds RBC ²	
thylbenzene (I)	0.005-0.005	1 / 13	0.026	0.026	0.0047	NC	÷	0.13	0.7 f	No	Less than RBC ¹ , Less than ARAR ⁵	
'oluene (k)	0.0005-0.0005	7 / 10	0.00045	0.0019	0.0007	NC		0.075	1 f	No	Less than RBC ¹ , Less than ARAR ⁵	
inichlaroethylene (k)	0.0005-0.0013	2 / 10	0.0033	0.0033	0.0005	NC	•	0.0016	NA	Yes	Excessis RBC ³	
(yienes (k)	0.00084-0.0008	1 / 10	0.0014	0.0014	0.00055	NC		1.2	NA	No	Less than RBC ¹	
GROUNDWATER * (mg/L) - UNFILTE	RED - CONT	INUED									
OTHER												
Extractable Petroleum Hydroca	rbone (EPH)											
C9-C18 Aliphetics ()	0.09-0.3	3 7 13	0.21	0.5	0.16	NC		NA	NA	Y85	No standard available	
C11-C22 Arometics [j]	0.03-0.04	S / 13	0.048	0.3	0.053	NC		NA	NA	Yes	No standard evailable ⁷	
Volstila Petroleum Hydrocarbo	na (VPH)											
C5 C8 Aliphatics (I)	0.0025-0.075	1/8	0.047	0.047	0.02	NA		NA	NA	Yes	No standard available ³	
C9-C12 Aliphetics ()	0.032-0.065	4 / 13	0.032	0.34	0.061	NC	~	NA	NA	Yee	No standard available	
C9-C10 Aromatics (I)	0.012-0.02	4 / 13	0.014	0,61	0.082	NC	-	NA	NA	Yes	No standard evailable ²	
DOWNGRADIENT SI	EDIMENT ^d (mg/l	(g)										
PAL METALS												
Aluminum	NA	3/3	2930	4840	3843	NC	18000	7800	NA	No	Less than RBC ¹ , Background ²	
Arsonic	NA	3/3	5.46	14.0	10.8	NC	19	0.43	NA	Yeo	Exceeds RBC ² , Background ²	
Barium	NA	3/3	7.13	11.4	9.5	NC	54	550	NA	No	Less than RBC ¹ , Background ²	
Calcium	NA	3/3	10.3	736	427	NC	810	NA	NA	No	Essential Nutrient ⁴ , Background ²	
Chromium	NA	3/3	11.2	16.1		NC	33	39	NA	No	Less than RBC ¹ , Background ²	
Cobalt	NA	3/3	2.23	6,9	4.3	NC	4.7	470	NA	No	Less than RBC'	
Copper	NA	3/3	6.56	23.4		NC	13.5	310	NA	No	Loss than ABC ¹	
ran	NA	3/3	7010	10900	cococi, sentionant	NC	18000	2300	NA	Yes	Excession RBC ² , Background ²	
.ead	NA	3/3	11.4	30.0		NC	48	NA	400 e	No	Less than ARAR ⁵ , Background ²	
Magnesium	NA	3/3	1580	2630		NC	5500	NA	NA	No	Essential Nutrient ⁴ , Background ²	
Manganese	NA	3/3	70.7	186		NC	380	180	NA	Yes	Exceeds RBC ² , Background ²	

DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN

AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Contraction of the second	Ranga	Frequency			Concentration	ı					
	af	ot	Minimum	Maximum	Arithmetic	95%	Beck-	Region III			
	SQL	Detection	Detected	Detected	Mean	UCL	ground®	RBC**	ARARa	CPC?	Notes
Nickel	NA	3/3	9.55	, 18,1	12.7	NC	14.6	160	NA	No	Less than RBC ¹
Potsesium	NA	3/3	364	426	402	NC	2400	NA	NA	No	Essential Nutrient ⁴ , Background ²
Sodium	NA	3/3	259	307	275	NC	234	NA	NA	No	Essential Nutrient ⁴
Vanadium	NA	3/3	7.91	10.4	8,9	NC	32.3	55	NA	No	Less then RBC ¹ , Background ²
Zinc	NA	3/3	22.8	39.6	31.4	NC	43.9	2300	NA	No	Less than RBC ¹
PAL SEMIVOLATILE ORGANICS											
Benzo[k]fluoranthene	0.30-0.30	1/3	0.4	0.40	0.23	NC		8.8	NA	No	Less than RBC ¹
Chrysene	0.60-0.60	1 / 3	2	2	0.86	NC		88	NA	No	Less then RBC ¹
Fluoranthene	0.30-0.30	2/3	1	3	1.04	NC	~	310	NA	No	Less than RBC ¹
Phenanthrene	0.20-0.20	2 / 3	0.9	2	1	NC		· 310 h	NA	No	Less than RBC ¹
Ругеле	0.20-0.20	2/3	1	3	1.4	NC	-	230	NA	No	Less than RBC ¹
PAL VOLATILE ORGANICS											
Trichlorofluoromethans	NA	3/3	0.0082	0,0096	0.0091	NC	-	2300	NA	No	Less than RBC ¹
PESTICIDES/PCBS								-00			
4,4-DDD	NA	3/3	0.0174	0.12	0.068	NC	-	2.7	NA	No	Less than RBC ¹
4,4-DDE	0.0077-0.0077	1/3	0.015	0.015	0.0076	NC		1.9	NA	No	Less than RBC ¹
4,4-DDT	NA	2 / 3	0.02	0.046	0.024	NC		1.9	NA	Na	Less than RBC ¹
OTHER											
Total Petroleum Hydronarbone	NA	3/3	65,8	290	162	NC		NA	NA	Yas	No standard availabla ³
INDOOR AIR " (ug/m	1 ³)										
VOLATILE ORGANICS											
2-Methylheptane	4.4	2/5	5.2	19	7.3	NC	(m)	200	NA	No	Less than RBC ¹
Ethylbenzene	NA	5 / 5	2.8	470	102	NC		100	NA	Yes	
Nonane	4.4	1 / 5	7.2	7.2	3.2	NC		200	NA	No	Less than RBC ¹
Octane	4.4	1 / 5	21	21	5.9	NC	0.0		NA	Yes	
Toluene	NA	5/5	70	1000	297	NC		42	NA	Yes	
Acetone	NA	5/5	52	470	172	NC		37	NA	Yes	and the second sec
Xylene	8.8	4 / 5	8	92	30.4	NC		730	NA	No	Less than RBC ¹
2-Methylheptane	4.4	1 / 5	8.7	8.7	3.5	NC	(n)	200	NA	No	Less than RBC ¹

NOTES:

a Samples included in data set are listed on Table 9-1

b Samples included in data set are listed on Table 9-1

Chemicals selected as CPCs are shaded.

DATA SUMMARY AND SELECTIONOF CHEMICALS OF POTENTIAL CONCERN

AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	Range	Frequency			Concentratio	n					
	af	of	Minimum	Maximum	Arithmetic	95%	Beck-	Region III			
	SQLa	Detection	Detected	Detected	Moan	UCL	ground®	RBC**	ARARe	CPC?	Notes
c Samples included in data set are listed on Table 9-	1								RBC - Risk-based co	inciintration .	
d Samples included in data set are listed on Table 9-	1								mg - milligrame		
USEPA soil lead screening level (OSWER Directive	9355.4-12, 19946	6							kg - kliograme		
MCL (USEPA, 1996b)									L - liter		
g Secondary MCL (USEPA, 1996b)									ARARs - Applicable	or Relevant and Appropriat	e Requiremente
Value for naphthalene used as surrogate									MCL - Maximum Co	ontaminant Level	
Data for SVOC analysis									CPC - chemical of p	otential concern	
Deta for EPH analysis									bgs - below ground	surface	
k Data for VOC analysis									SQL - Sample Quant	titation Limit	
Data for VPH analysis									not applicable fo	r organica	
m Samples included in data set are listed on Table 9									NC - 95 percent UC	L not calculated for data a	ets with less then 10 samples or groundwater
n Value is RfC for the C9-C12 aliphatic fraction pub	iehed by MADEP (18	97); adjusted to represen	t a value of 10%	of the RfC.							
o Value is the RIC for the C5-C8 aliphatic fraction p	ublished by MADEP	(1997); adjusted to repres	ent a value of 10	% of the RIC.							
Background: Maximum concentration in Fort Deven	e background listed;								NA - No value avait	abla	
85 percent UCL of Fort Devens background groundv	vater. See Appendix	F for development of ba	ekground.						UCL - upper confide	ance limit	
*Region III RBCe (USEPA, 1997a): Residential RBC	for soil used for sec	liment and surface and ev	beurlece soil evel	artion; tap water	RBC used						
for groundwater evaluation. Ambient Air RBCs used	for indoor air svalua	tion. RBCs based on car	cinogenic effecte a	re associated with	a 1x10 ⁴ cancer n	ik level;					
RBCe based on noncercinogenic effecte are pesociat	ed with an adjusted i	HQ of 0.1.									
Less than RBC ¹ - Maximum detected concentration (ese than risk-based o	oncentration								•	
Background ² - Sample concentrations detected are a	t or below backgroun	nd concentrations.									
Exceeds RBC ³ - Maximum detected concentration e	xceeds risk-based co	ncentration									
Essential Nutrient ⁴ - Analyte is an essential human n	utdent (magnesium,	celcium, poteseium, eodi	um) and is not con	eldered a CPC.							
Less than ARAR ⁵ - Maximum detected concentration	is less than concent	ration shown in ARARs o	olumn.								
Exceeds ARAR ⁸ - Maximum detected concentration	e greater than conce	ntration shown in ARARs	column.								
No standard available ⁷ - No standards available for c	omparison, analyte is	considered a CPC.									

6

POTENTIALLY EXPOSED POPULATION	EXPOSURE MEDIUM AND POINT	EXPOSURE ROUTE	EVALUATED	REASON FOR SELECTION OR EXCLUSION
CURRENT LAND USE				
ite Maintenance Worker	Surface soil in grassy area	Incidental ingestion	Yes	Maintaining the grassy area may expose current and future site maintenance workers to contaminated surface soil via ingestion and dermal contact
		Dermal contact	Yes	
		Inhalation of particulates	Ýes	Maintaining the grassy area may expose current and future site maintenance workers to particulates via inhalation.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in surface soil.
	Groundwater	Incidental ingestion, dermal contact, and inhalation of VOCs	No	Maintaining the grassy area would not expose the site maintenance workers to groundwater.
	Subsurface soil beneath grassy area	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	No excavation presently occurring
	Surface water	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	Maintaining the grassy area would not expose the site maintenance workers to surface water.
	Sediment	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	Maintaining the grassy area would not expose the site maintenance workers to sediment
School Occupants Children/Faculty)	Groundwater	Incidental ingestion of and dermal contact with groundwater	No	The groundwater is not considered a conductive aquifer and is very shallow Currently the school is not occupied
		Inhalation of VOCs from groundwater	No	The school is not currently open. No students or faculty are present.
	Surface soil in grassy area	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	Na	The school is not currently open. No students or faculty are present.
	Subsurface soil beneath grassy area	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	The school is not currently open. No students or faculty are present
	Surface water	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	The school is not currently open. No students or faculty are present
	Sediment	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	The school is not currently open. No students or faculty are present.

POTENTIALLY EXPOSED POPULATION	EXPOSURE MEDRIM AND POINT	EXPOSURE ROUTE	EVALUATED?	REASON FOR SELECTION OR EXCLUSION
uild Trespasser	Surface water (as groundwater discharge)	Incidental ingestion	Yes	Children may incidentally ingest surface water while wading in the wetland area when surface water is present.
		Dermal contact	Yes	Wading may expose children to surface water in the wetland area.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in shallow groundwater.
	Surface soil in grassy area	Incidental ingestion	Yes	Currently the school is not open. However, child trespassers might play in the unpaved (grassy) areas and may be exposed to contaminated surface soil via ingestion and dermal contact.
		Dermal contact	Yes	
		Inhalation of particulates	Yes	Playing could genearte dust in the grassy area.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in surface soil.
	Subsurface soil beneath grassy area	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	No excavation presently occurring
	Sediment	Incidental ingestion	Yes	Children may incidentally ingest and contact sediment while wading in the wetland area during the spring months.
		Dermal contact	Yes	
		Inhalation of particulates and VOCs	No	VOCs were not selected as CPCs in sediment; no dust from sediment

11.00

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

POTENTIALLY EXPOSED POPULATION	EXPOSURE MEDIUM AND POINT	EXPOSURE ROUTE	EVALUATED?	REASON FOR SELECTION OR EXCLUSION
UTURE LAND USE	1			
itility/Construction Worker	Surface soil and subsurface soil beneath grassy area	Incidental ingestion	Yes	Performing repairs to utility lines or during excavation activities may expose utility workers to contaminated surface and subsurface soil via ingestion and dermal contact.
		Dermal contact	Yes	
		Inhalation of particulates.	Yes	Excavation work is likely to generate dust; therefore, this exposure pathway may be significant.
		Inhalation of VOCs	Yes	Performing repairs to utility lines or during excavation activities may expose utility workers to VOCs in surface and subsurface soil via inhalation.
	Groundwater	Incidental ingestion and dermal contact	No	Workers will be wearing protective clothing, which will mitigate exposure via these routes.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in shallow groundwater
	Surface water	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	Excavation at the grassy area would not expose excavation workers to surface water.
	Sediment	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	Excavation at the grassy area would not expose excavation workers to surface water.
ichool Occupants Children/Faculty)	Groundwater	Incidental ingestion of and dermal contact	No	The groundwater is not considered a conductive aquifer and is very shallow. The planned reuse of AOC 69W is to reopen the school, which will rely on existing public water supplies in the future.
	Air in school	Inhalation of VOCs migrating from soil and/or shallow groundwater beneath the school to indoor air	Yes	Children and faculty may be exposed to VOCs which have a potentially complete migration pathway from soil and/or shallow groundwater to air within the rooms of the school
	Surface water (as groundwater discharge)	Incidental ingestion	Yes	Children may incidentally ingest surface water while wading in the wetland area when surface water is present
	1	Dermal contact	Yes	Wading may expose children to surface water in the wetland area.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in shallow groundwater.

.

POTENTIALLY EXPOSED POPULATION	EXPOSURE MEDRIM AND POINT	EXPOSURE ROUTE	EVALUATED?	REASON FOR SELECTION OR EXCLUSION
	Surface soil in grassy area	Incidental ingestion	Yes	Students might play in the unpaved (grassy) areas and may be exposed to contaminated surface soil via ingestion and dermal contact.
		Dermal contact	Yes	
		Inhalation of particulates	Yes	Playing could genearte dust in the grassy area.
		Inhalation of VOCs	No	VOCs were not selected as CPCs in surface soil.
	Subsurface soil beneath grassy area	Incidental ingestion, dermal contact, and inhalation of VOCs and particulates	No	Excavations would not be left open or unfilled
	Sediment	Incidental ingestion	Yes	Students may incidentally ingest and contact sediment while wading in the wetland area during the spring months.
		Dermal contact	Yes	
		Inhalation of particulates and VOCs	No	VOCs were not selected as CPCs in sediment; no dust from sediment
l public	Groundwater	Potable use (ingestion, dermal contact, inhalation of VOCs during showering)	Yes	Although the groundwater is not anticipated to be used as a source of potable water, potable water use is evaluated.

TABLE 9-4 EXPOSURE PARAMETERS AOC 69W

	CURRENT	LAND USE		FUTURE	LAND USE			
PARAMETER	SITE MAINTENANCE WORKER	CHILD TRESPASSER (&18 YEARS)	PUPIL (6-18 YEARS)	ENCAVATION WORKER	ADULT RESIDENT (POTABLE WATER)	CHILD RESIDENT (POTABLE WATER)	UNITS	SOURCE 1
Sediment/Soil Ingestion Rate								
RME	100	100	100	480	Not Applicable	Not Applicable	mg/day	USEPA, 1994
Central Tendency	50	50 ,	50	200 7	Not Applicable	Not Applicable	mg/day	USEPA, 1994
Surface Water Ingestion Rate								
RME	Not Applicable	0.05	0.05	Not Applicable	Not Applicable	Not Applicable	L/hour	USEPA, 1986
Central Tendency	The second s	0.025	0.025	Not Applicable	Not Applicable	Not Applicable	L/hour	Assumption
Groundwater Ingestion Rate								
RME	Not Applicable	Not Applicable	Not Applicable	Not Applicable	2	t	L/day	USEPA, 1994
Central Tendency	Not Applicable	Not Applicable	Not Applicable	Not Applicable	1.4	0.7	L/day	Assumption
Surface Area Exposed ²	5,200	5,053	5,053	5,200	5,200	Not Applicable	cm2/day	USEPA, 1989a / 1992
Fraction Ingested From Site	100%	100%	100%	100%	100%	100%		Assumption
Relative Absorption Factor	100%	100%	100%	100%	100%	100%		USEPA, 1995
Inhalation Rate 5								
RME		2.3	2.3	4.8	Not Applicable	Not Applicable	m3/hour	USEPA, 1989
Central Tendency	Not Applicable	2	2	1.5	Not Applicable	Not Applicable	m3/hour	USEPA, 1996
Soil Exposure Time	8	8	8	8	Not Applicable	Not Applicable	hours/day	Assumption
Soil Exposure Frequency 3	64	96	140	90	Not Applicable	Not Applicable	days/year	Assumption
Surface Water/Sediment Exposure Time	Not Applicable	2.6	2.6	Not Applicable	Not Applicable	Not Applicable	hours/day	USEPA, 1989a
Surface Water/Sediment Exposure Frequency *	Not Applicable	24	24	Not Applicable	Not Applicable	Not Applicable	days/year	Assumption
Groundwater Exposure Frequency	Not Applicable	Not Applicable	Not Applicable	Not Applicable	350	350 .	days/year	USEPA, 1994
Exposure Duration								
RME	25	13	13	0.35	30	6	years	USEPA, 1994, Assumption
Central Tendency	9	9	9	-	9	2	years	USEPA, 1994, Amumption
Body Weight	70	45	45	70	70	15	kg	USEPA, 1989a, USEPA, 1994
Averaging Time								
Cancer	70	70	70	70	70	-70	years	USEPA, 1989b
Noncancer	equal ED	equals ED	equals ED	equals ED	equals ED	equals ED	years	USEPA, 1989b
Particulate Emission Factor	1.32+09	Not Applicable	Not Applicable	1,32+09	Not Applicable	Not Applicable	m3/kg	USEPA, 1996
Soil-to-Air Volatilization Factor	No volatile CPCs	No volatile CPCs	No volatile CPCe	No volatile CPCs	Not Applicable	Not Applicable		
Permeability Constant	Not Applicable	Chemical-specific	Chemical-specific	Not Applicable	Not Applicable	Not Applicable	cm/hour	USEPA_ 1992

TABLE 9-4 EXPOSURE PARAMETERS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

CURRENT LAND USE	FUTURE LAND USE
SITE MAINTENANCE CHILD TRESPAN	
PARAMETER WORKER (6-18 YEARS)	S) (6-18 YEARS) WORKER (POTABLE WATER) (POTABLE WATER) UNITS SOURCE *

2

÷

Notes:

1 - Exposure variables with source listed as "assumption" are site specific; the remainder are default values. RME and central tendency parameters listed separately only when they differ from each other.

2 - body area = hands, arms, lowerlegs, and feet (trespasser & pupil); head, hands, arms (maintenance worker / utility worker).

3 - 2 days per week for 32 weeks for site maintenance worker (Apr-Nov); 3 days per week for 32 weeks for child trespasser (Apr-Nov);

5 days per week for 18 weeks for construction worker, 5 days per week for 22 weeks for pupil (Sept-Nov, Apr-mid Jun) plus 3 days per week for 10 weeks (mid Jun - Aug). 4 - 2 days per week for 12 weeks.

5 - Inhalation rate for excavation worker based on the heavy activity rate listed in the Exposure Factors Handbook (USEPA, 1989c)

6 - ED for residents used for pupil and trespasser

7 - Assumption, value based on RME child.

cm2 = square centimeters

mg - milligrams

m3 - cubic meters

kg - kilograms

CPC - Chemical of Potential Concern

NA - not applicable

USEPA, 1986 - Superfund Exposure Assessment Manual

USEPA, 1989a - Exposure Factors Handbook

USEPA, 1989b - Risk Assessment Guidance for Superfund, Part A

USEPA, 1992 - Dermal Exposure Assessment

USEPA, 1994 - USEPA Region I Risk Updates, Number 2

USEPA, 1995 - USEPA Region I Risk Updates, Number 3

USEPA, 1996 - USEPA Soil Screening Level Guidance

TABLE 9-5 PERMEABILITY COEFFICIENTS FOR CPCs IN SURFACE WATER AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Compound	Exposure Time (ET) (hr/day)	log K _{ow} 1	MW ¹	Kp ² (cm/hr)	T ² (hr)	t* ² (hr)	B ²	Kp _{event} (cm/day)
ORGANICS ³								
TPHC'	2.6	5.18E+00	202	5.3E-01	1.5E+00	7.2E+00	1.5E+01	2.9E + 00
bis(2-Ethylhexyl)phthalate	2.6	101010 1 000	202	3.3E-02	2.1E+01	1.0E+02	1.3E+01	6.7E-0
Chloroform	2.6			1.3E-01	4.7E-01	1.1E+00	9.3E-03	4.6E-0
2-Methylnaphthalene	2.6	4.11E+00	142.2	2.1E-01	6.4E-01	4.5E+00	1.3E+00	7.7E-0
Trichloroethene	2.6			2.3E-01	5.5E-01	1.3E+00	2.6E-02	8.6E-0
n-Hexane**	2.6	3.9E+00	86.17	3.3E-01	2.9E-01	1.3E+01	7.9E-01	8.1E-0
Naphthalene	2.6			6.9E-02	5.3E-01	2.2E+00	2.0E-01	2.8E-0
INOR GANICS 4						200		
Aluminum	2.6		1	1.0E-03	NA	NA	NA	2.6E-0
Arsenic	2.6			1.0E-03	NA	NA	NA	2.6E-0
Iron	2.6			1.0E-03	NA	NA	NA	2.6E-0
Manganese	2.6			1.0E-03	NA	NA	NA	2.6E-0

Notes:

¹ Needed only for compounds not included on USEPA, 1992c, Table 5-8, these values obtained from USEPA, 1993.

- ² Values from USEPA, 1992c, Table 5-8 or calculated as follows:
- 1. $\log Kp = -2.72 + 0.71 \times \log K_{ow} 0.0061 \times MW$

2.
$$B = K_{ow} / 10^{\circ}$$

3. $r = l_{sc}^{2^{W}} / (6 \ge (l_{sc} \ge 10^{(-2.72 - 0.0061 \ge MW)}))$

where $l_{sc} = 10 \ \mu m = 0.001 \ cm$ If B ≤ 0.1 , then t* = 2.4 x r 4. If $0.1 \le B \le 1.17$, then $t^* = (8.4 + 6 \log B) \times r$ If $B \ge 1.17$, then $t^* = 6x(b - (b^2 - c^2)^{0.5}) xr$ where $b = (2/\pi) x (1 + B)^2 - c$ and c = (1 + 3B)/3

³ For organics; Kp from USEPA, 1992c, Table 5-8, estimated values

 Kp_{event} calculated as follows: If ET < t* , then;

 $\label{eq:kp_event} \begin{array}{l} \mathrm{Kp}_{\mathrm{event}}\left(\mathrm{cm/event}\right) = 2\,\mathrm{x}\,\mathrm{Kp}\,\mathrm{x}\left[\left(6\,\mathrm{x}\,\mathrm{T}\,\mathrm{x}\,\mathrm{ET}\right)/\,\pi\right]^{0.5} \\ \mathrm{If}\,\mathrm{ET} > t^{*}\,,\,\mathrm{then}; \end{array}$

 Kp_{event} (cm/event) = Kp x [(ET/1 + B) + 2 x T x ((1 + 3 B)/1 + B)]

⁴ For inorganics, Kp from USEPA, 1992c, Table 5-3 or default value of 1E-03 cm/hr

Kpevent calculated as follows:

 Kp_{event} (cm/day) = Kp x ET

* – Pyrene used as a surrogate to calculate the K_p for TPHC. Also, this calculated K_p is used as a surrogate for all three Extractable Petroleum Hydrocarbon fractions (C9–C18 and C19–C36 aliphatic and C10–C22 aromatics) and the Volatile Petroleum Hydrocarbon (VPH) C9-C12 aliphatics and C9-C10 aromatics fractions. - n-Hexane's calculated Kn is used as a surrogate for the VPH C5-C8 aliphatics fraction.

Acronyms: hr = hour cm = centimeter Kp = Permeability Coefficient Kow = Octanol/water partitioning coefficient MW = molecular weight r = lag time t* = time to reach steady state B = partitioning factor NA = not applicable TPHC = Total Petroleum Hydrocarbons 5.3E - 01 = 0.53

TABLE 9-6 ORAL DOSE/RESPONSE DATA FOR CARCINOGENIC EFFECTS AOC 69W

Compound	Weight of Evidence	Oral Slope Factor (mg/kg/day) ⁻¹	Test Species	Study Type	Tumor Type	Source
2-Methylnaphthalene	ND					IRIS
Aluminum	Not Listed					HEAST
Arsenic	A	1.5E+00	Human	Oral-DW	Skin	IRIS
Beryllium	B2	4.3E+00 (w)	Rat	Oral-DW	Total	IRIS
Bis (2 - ethylhexyl) phthalate (BEHP)	B2	1.4E-02	Mouse	Oral-diet	Liver	IRIS
Chloroform	82	6.1E-03	Rat	Oral-drinking water	Kidney	IRIS
Trichloroethene	82	1.1E-02 (w)	Rat	Oral-diet	Lung/Liver	HEAST
Iron	Not Listed					
Manganese	D		1			IRIS
Naphthalene	D					IRIS
Extractable Petroleum Hydrocarbons						
C9-C18 Aliphatics	Not Listed					
C19-C36 Aliphatics	Not Listed					
C10-C22 Aromatics	Not Listed					-
Total Petroleum Hydrocarbons	Not Listed					2
Volatile Petroleum Hydrocarbons						
C5-C8 Aliphatics	Not Listed					1
C9-C12 Aliphatics	Not Listed	100 mm				
C9-C10 Aromatics	Not Listed					
ND - Not Determined W - Withdrawn from IRIS DW - Drinking water mg - milligram kg - kilogram IRIS - Integrated Risk Information System HEAST - Health Effects Assessment Summary Tables 1.5E+00 = 1.5 Sources: IRIS as of 4/98 HEAST, 1997		=-Specific): inogen (61 - limted evidenc: f carcinogenicity in animals w ins) inogen				
	E - Evidence of lack of ca					

TABLE 9–7 INHALATION DOSE/RESPONSE INFORMATION FOR CARCINOGENIC EFFECTS AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

Compound	Weight of Evidence	Inhalation Slope Factor* (mg/kg/day) ⁻¹	Unit Risk (µg/m?) ⁻¹	Test Species	Study Type	Tumor Type	Source
2-Methylnaphthalene	ND				1.002		IRIS
Aluminum	ND						HEAST
Arsenic	A	1.5E+01 +	4.3E-03	Human	Inhalation	Lung	IRIS
Beryllium	B2	8.4E+00	2.4E-03	Human	Inhalation	Lung	IRIS
Bis(2-ethylhexyl)phthalate (BEHP)	B2		ND	1.1.1			IRIS
Chloroform	B2	8.1E-02	2.3E-05	Mouse	Oral-gavage	Liver	IRIS
Trichloroethene	B2	6.0E-03 (N)	2.0E-06 (N)	NA	NA	NA	NCEA
Iron	Not Listed						
Manganese	D			1			IRIS
Naphthalene	D				G		IRIS
Extractable Petroleum Hydrocarbons						+	
C9-C18 Aliphatics	Not Listed				1		-
C19-C36 Aliphatics	Not Listed						
C10-C22 Aromatics	Not Listed					A CONTRACTOR OF CONTRACTOR OFO	
Total Petroleum Hydrocarbons	Not Listed			1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-1-			
Volatile Petroleum Hydrocarbons				1977 - C			
C5-C8 Aliphatics	Not Listed						
C9-C12 Aliphatics	Not Listed						
C9-C10 Aromatics	Not Listed					12	

÷

* - Source of slope factor is HEAST, 1997 unless otherwise noted.

B - Probable human carcinogen (B1 - limited evidence of cancer in humans;

B2 - sufficient evidence of carcinogenicity in animals with inadequate or lack

+ - Source of slope factor is IRIS

A - Human carchogen

Weight of Evidence (Route-Specific):

of evidence in humans)

C - Possible human carcinogen

D – Not classifiable as to human carcinogenicity
 E – Evidence of lack of carcinogenicity to humans

ND - Not Determined

mg - milligram

kg - kilogram

µg - microgram

IRIS - Integrated Risk Information System

HEAST - Health Effects Assessment Summary Tables

1.5E+01 = 15

Sources: IRIS as of 4/98

HEAST, 1997

(N) = This value has been provided in response to a specific request

1.4

CAR-INH.wk1

TABLE 9-8 ORAL DOSE/RESPONSE INFORMATION FOR NONCARCINOGENIC EFFECTS AOC 69W

Compound	CHRONIC ORAL RID (mg/kg-day)	SUBCHRONIC ORAL RfD ¹ (mg/kg-day)	STUDY TYPE	CONFIDENCE LEVEL	CRITICAL EFFECT	TEST ANIMAL	UNCERTAINT FACTOR	Y	
2-Methylnaphthalene	4E-02 *	4E-02 *	1					I have been a	
Aluminum	1E+00	1E+00 #					1	NCEA, 1994a	
Arsenic	3E-04	3E-04	Oral-DW	Medium	Keratosis and hyperpigmentation	Human	3 H	IRIS	
Beryllium	2E-03	2E-03	Oral-DW	Low	No effects observed	Ret	100 H.A	IRIS	
Bis(2-ethylhexyl)phthalate (BEHP)	2E-02	2E-02 #	Oral-diet	Medium	Increased liver weight	Guinea Pig	1,000 H,A,S	IRIS	
Chloroform	1E-02	1E-02	Oral-capsule	Medium	Fatty cyst formation in liver	Dog	1,000 H,A, S	IRIS	
Trichloroethene	6.0E-03 (N)	ND	NA	NA	NA	NA	NA	NCEA	
Iron	3E-01	3E-01 #						NCEA, 1994c	
Manganese (soil) USEPA Region 1	7.1E-02 ++	7.1E-02 #						Region 1	
Manganese (drinking water) USEPA Region 1	2.4E-02 +++	2.4E-02 #						Region 1	
Naphthalene	4E-02 W	4E-02	1					NCEA, 1994	
Extractable Petroleum Hydrocarbons									
C9-C18 Aliphatics	6E-01	6E-01 #						MADEP, 1997	
C19-C36 Aliphatics	6E+00	6E+00 #						MADEP, 1997	
C10-C22 Aromatics	3E-02	3E-01	1					MADEP, 1997	
Total Petroleum Hydrocarbons	3E-02	3E-01 **							
Volatile Petroleum Hydrocarbons								Low a descent	
C5-C8 Aliphatics	6E-02	6E-01	1					MADEP, 1997	
C9-C12 Aliphatics	6E-01	6E-01 #						MADEP, 1997	
C9-C10 Aromatics	3E-02	3E-01						MADEP, 1997	
ND - No data available	1 - Source for all subchronic	RIDs is HEAST, 1997	-	Uncertainty factors:	H - variation in human sensitivity				
W - RID withdrawn from IRIS/HEAST	"RfD for naphthalene is used	es surrogate for PAHs			A - enimal to human extrapolation				
mg – milligram	without assigned RID				S - extrapolation from subchronic to chronic NOAEL				
kg – kilogram	(N) = Value was provided in	response to a special reques	đ.		L - extrapolation from LOAEL to NOAEL				
DW - Drinking Water	** Value for C10-C22 used	as conservative surrogate			N - NOEL not attained				
IRIS - Integrated Risk Information System	41E-02 = 0.041				D - Lack of supporting data				
HEAST - Health Effects Assessment Summary Tables					M - addition modifying factor				
USEPA - United States Environmental Protection Agency									
NCEA - National Center for Environmental Assessment									
MADEP - Messechusetts Department of Environmental Prote	ction								
SOURCES:									
MADEP, 1994: Interim Final Petroleum Report, Development	of Health-besed Alternative to	the Total Petroleum Hydroc	arbon Parameter						
IRIS as of 4/98					++ RID for manganese in food divided by 2 to accou	nt for dietary exposure			
NCEA, 1994a, b, c, 1995					+++ RfD for manganese in food divided by 2 to acco	sunt for dietary exposure,			
HEAST, 1997					and by 3 as a modifying factor				
Region 1 Risk Update, 1996					# Subchronic RID not evallable. Chronic RID used as	a surrogata			

INHALATION DOSE/RESPONSE DATA FOR NONCARCINOGENIC EFFECTS

AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

COMPOUND	CHRONIC INHALATION RIC (mg/m ³)	SUBCHRONIC ¹ INHALATION RfC (mg/m ³)	STUDY TYPE	CONFIDENCE LEVEL		TEST	UNCERTAINTY	SOURCE
2-Methylnaphthalene	Not listed							
Acetone	ND	ND						MADEP
Aluminum	ND	ND					4	HEAST
Arsenic	ND	ND	1					IRIS
Beryllium	2E-05	2E-05						IRIS
Bis(2-ethylhexyl)phthalate (BEHP)	ND	ND						IRIS
Chloroform	ND	ND	1-					IRIS
Ethylbenzene	1E+00	ND	Inhalation	Low	Development toxicity	Rat/Rabbit	300 H,A,S	IRIS
Trichloroethene	ND	ND	L		A COLORED & COLORED			IRIS
Iron	Not Listed							
Manganese	5E-05	ND	Inhalation	Medium	Neurobehavioral impairment	Human	1,000 H.L.D	IRIS
Naphthalene	ND	ND						IRIS
Extractable Petroleum Hydrocarbons						1.1		1
C9-C18 Aliphatics	2.0E+00	2.0E+00 #	1	· · · · · · · · · · · · · · · · · · ·				MADEP, 1997
C19-C36 Aliphatics	ND	ND						MADEP, 1997
C10-C22 Aromatics	7.1E-02	7.1E-02 #						MADEP, 1997
Total Petroleum Hydrocarbons	6.0E-02 *	6.0E-02 #	1					1.1
Toluene	4E-01	ND	Inhalation	Medium	Neurological effects	Human	300 H,L,D	IRIS
Volatile Petroleum Hydrocarbons			-	-	the second s	-		
C5-C8 Aliphatics	2.0E-01	2.0E-01 #		i i consul				MADEP, 1997
C9-C12 Aliphatics	2.0E+00	2.0E+00 #		-	5			MADEP, 1997
C9-C10 Aromatics	6.0E-02	6.0E-02 #						MADEP, 1997
ND – No data available mg – milligram kg – kilogram µg – microgram DW – Drinking Water IRIS – Integrated Risk information System HEAST – Health Effects Assessment Summa MADEP = Massachusetts Department of Env	E. B. C. Martin, Collector, C. M. C.	 Source for all subct Value for C9-C10 u surrogate Subchronic RfD no surrogate. 	used as conserva	tive	 H – variation in human sensitivity A – animal to human extrapolation S – extrapolation from subchronic to chroit – extrapolation from LOAEL to NOAEL N – NOEL not attained D – Lack of supporting data M – additional modifying factor 	nic NOAEL		3.

5E-05 = 0.00005

SOURCES:

IRIS as of 4/98

HEAST, 1997

MADEP, 1997 - Interim Final Petroleum Report: Developement of Health-Based Alternatives to the TPH Parameter,

NON-INH.wkt

DERMAL DOSE/RESPONSE INFORMATION FOR CARCINOGENIC AND NONCARCINOGENIC EFFECTS

AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

1.0E+00 3.0E-04 2.0E-02 1.0E-02	ND 3.0E-04	ND 1 FE L 00	20%		(mg/kg-day)	(mg/kg-day)	(mg/kg-day) ⁻¹
2.0E-02		1 55 100	2070	Default ⁴	2.0E-01	ND	ND
	AID	1.5E+00	98%	Vahter, 1983	2.9E-04	2.9E-04	1.5E+00
1.0E-02	ND	1.4E-02	100%	Chadwick et al., 1982	2.0E-02	ND	1.4E-02
and the second se	1.0E-02	6.1E-03	100%	Brown et al., 1974	1.0E-02	1.0E-02	6.1E-03
6.0E-03	6.0E-03	1.1E-02	100%	Prout et al., 1985	6.0E-03	6.0E-03	1.1E-02
3.0E-01	ND	ND	2%	Goyer, 1991	6.0E-03	ND	ND
2.4E-02	ND	ND	4%	ATSDR, 1991	9.6E~04	ND	ND
4.0E-02	4.0E-02	ND	100%	Chang 1943	4.0E-02	4.0E-02	ND
4.0E-02	4.0E-02	ND	100%	Chang 1943	4.0E-02	4.0E-02	ND
							2
6.0E-01	ND	ND	50%	Default ⁴	3.0E-01	ND	ND
6.0E+00	ND	ND	50%	Default ⁴	3.0E+00	ND	ND
3.0E-02	3.0E-01	ND	91%	B(a)P ⁷	2.7E-02	2.7E-01	ND
3.0E-02	3.0E-01	ND	91%	B(a)P ⁷	2.7E-02	2.7E-01	ND
6.0E-02	ND	ND	80%	Default ⁴	4.8E-02	ND	ND
6.0E-01	ND	ND	80%	Default ⁴	4.8E-01	ND	ND
3.0E-02	3.0E-01	ND	91%	B(a)P ⁷	2.7E-02	2.7E-01	ND
	2.4E-02 4.0E-02 4.0E-02 6.0E-01 6.0E+00 3.0E-02 3.0E-02 6.0E-02 6.0E-02 6.0E-01	2.4E-02 ND 4.0E-02 4.0E-02 4.0E-02 4.0E-02 6.0E-01 ND 6.0E+00 ND 3.0E-02 3.0E-01 3.0E-02 3.0E-01 6.0E-02 ND 6.0E-02 ND 6.0E-01 ND	2.4E-02 ND ND 4.0E-02 4.0E-02 ND 4.0E-02 4.0E-02 ND 6.0E-01 ND ND 6.0E+00 ND ND 3.0E-02 3.0E-01 ND 6.0E-02 ND ND 6.0E-01 ND ND 6.0E-02 3.0E-01 ND 6.0E-02 ND ND 6.0E-01 ND ND	2.4E-02 ND ND 4% 4.0E-02 4.0E-02 ND 100% 4.0E-02 4.0E-02 ND 100% 4.0E-02 4.0E-02 ND 100% 6.0E-01 ND ND 50% 6.0E+00 ND ND 50% 3.0E-02 3.0E-01 ND 91% 3.0E-02 3.0E-01 ND 91% 6.0E-02 ND ND 80% 6.0E-01 ND ND 80%	2.4E-02 ND ND 4% ATSDR, 1991 4.0E-02 4.0E-02 ND 100% Chang, 1943 4.0E-02 4.0E-02 ND 100% Chang, 1943 4.0E-02 4.0E-02 ND 100% Chang, 1943 6.0E-01 ND ND 50% Defauk ⁴ 6.0E+00 ND ND 50% Defauk ⁴ 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 6.0E-02 ND ND 80% Defauk ⁴ 6.0E-01 ND ND 80% Defauk ⁴	2.4E-02 ND ND 4% ATSDR, 1991 9.6E-04 4.0E-02 4.0E-02 ND 100% Chang, 1943 4.0E-02 4.0E-02 4.0E-02 ND 100% Chang, 1943 4.0E-02 4.0E-02 4.0E-02 ND 100% Chang, 1943 4.0E-02 6.0E-01 ND ND 50% Defauk ⁴ 3.0E-01 6.0E+00 ND ND 50% Defauk ⁴ 3.0E+00 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 2.7E-02 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 2.7E-02 6.0E-02 ND ND 80% Defauk ⁴ 4.8E-02 6.0E-02 ND ND 80% Defauk ⁴ 4.8E-01	2.4E-02 ND ND 4% ATSDR 1991 9.6E-04 ND 4.0E-02 4.0E-02 ND 100% Chang 1943 4.0E-02 4.0E-02 6.0E-01 ND ND 50% Default ⁴ 3.0E-01 ND 6.0E+00 ND ND 50% Default ⁴ 3.0E+00 ND 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 2.7E-02 2.7E-01 3.0E-02 3.0E-01 ND 91% B(a)P ⁷ 2.7E-02 2.7E-01 6.0E-02 ND ND 80% Default ⁴ 4.8E-02 ND 6.0E-01 ND ND 80% Default ⁴ 4.8E-01 ND

NOTES:

1 - See preceding Dose/Response Tables

² - Dermal RfD = Oral RfD x Oral Absorption Efficiency

³ - Dermal CSF = Oral CSF / Oral Absorption Efficiency

Default⁴ - Compounds lacking specific absorption efficiencies are

assigned the following default efficiencies: volatiles and VPH - 80%,

semivolatiles, EPH, and pesticides - 50%, inorganics 20%

Heptachlor⁵ - Based on structural analogy to heptachlor

⁶The proposed alternate oral RfDs from

the Interim Final Petroleum Report: Development of Health-Based Alternatives to the Total Petroleum Hydrocarbon (TPH) Parameter (MADEP, 1997) are used as surrogates for the EPH and VPH fractions, see the oral RfD dose/response table.

B(a)P7 - Based on structural analogy to benzo(a)pyrene

* - Naphthalene used as a surrogate

** - Pyrene used as a surrogate

1E+00 = 1

ACRONYMS:

RfD - Reference Dose

CSF - Cancer Slope Factor

mg - milligram

kg - kilogram

ND - Not determined

ATSDR - Agency for Toxic Substances and Disease Registry

TABLE 9-11 QUANTITATIVE RISK SUMMARY AOC 69W

	CENTI	RAL TENDENCY	RME
	ELCR	HI ELCR	Hi
URRENT AND FUTURE USE		-	
URFACE SOIL (0 - 1 feet bgs)			
Incidental Ingestion of Surface Soil: Site Maintenance Worker Dermal Contact with Surface Soil: Site Maintenance Worker Inhalation of Particulates from Soil: Site Maintenance Worker TOTAL: SITE MAINTENANCE WORKER	5E-07 8E-07 <u>1E-09</u> 1E-06	0.02 3E-0 0.05 2E-0 <u>0.0006 4E-0</u> 0.07 5E-0	6 0.0 9 <u>0.000</u>
Incidental Ingestion of Surface Soil: Child Trespasser (6-18 years) Dermal Contact with Surface Soil: Child Trespasser (6-18 years) Inhalation of Particulates from Soil: Child Trespasser (6-18 years) TOTAL: CHILD TRESPASSER	1E-06 2E-06 <u>1E-09</u> 3E-06	0.04 3E-0 0.1 3E-0 <u>0.0007 3E-0</u> 0.1 6E-0	6 0 9 <u>0.000</u>
EDIMENT			
Incidental Ingestion of Downgradient Seciment: Child Trespasser (6-18 years) Dermal Contact with Downgradient Sediment: Child Trespasser (6-18 years) TOTAL: CHILD TRESPASSER	2E-07 <u>3E-07</u> 5 E-07	0.01 6E-0 0.04 4E-0 0.05 1E-0	7 0.0
ROUNDWATER (DISCHARGES TO SURFACE WATER)			
Incidental Ingestion of Surface Water. Child Trespasser (6-18 years) Wading Dermal Contact with Surface Water. Child Trespasser (6-18 years) Wading TOTAL: CHILD TRESPASSER	7E-07 <u>6E-07</u> 1E-06	0.02 2E-0 0.2 9E-0 0.2 2E-0	17 1
TOTAL RISK TO CHILD TRESPASSER EXPOSED TO SURFACE SOIL, SEDIMENT, and GROUNDWATER (DISCHARGES TO SURFACE WATER)	6E-06	0.4 1E-0	IS 0
UTURE USE			
SURFACE SOIL (0 - 1 feet bgs)	1.		
Incidental Ingestion of Surface Soil: Pupil (6-18 years) Dermal Contact with Surface Soil: Pupil (6-18 years) Inhalation of Particulates from Surface Soil: Pupil (6-18 years) TOTAL: PUPIL	2E-06 3E-06 <u>4E-09</u> 5E-06	0.06 5E-0 0.2 4E-0 0.002 6E-0 0.3 9E-0	06 0 09 <u>0.0</u>
Incidental Ingestion of Surface Soil: Excavation Worker Dermal Contact with Surface Soil: Excavation Worker Inhalation of Particulates from Soil: Excavation Worker TOTAL: EXCAVATION WORKER	1E-07 4E-08 <u>4E-11</u> 1E-07	0.07 3E-0 0.04 4E-0 <u>0.0009 1E-1</u> 0.1 3E-0	08 0. 00 0.00
SUBSURFACE SOIL (1 - 10 feet bgs)		-	
Incidental Ingestion of Subsurface Soil: Excavation Worker Dermal Contact with Subsurface Soil: Excavation Worker Inhalation of Particulates and Volatiles from Subsurface Soil: Excavation Worker TOTAL: EXCAVATION WORKER	4E-08 2E-08 2E-11 6 E-0 8	0.04 9E-0 0.04 2E-0 <u>0.8 6E-1</u> 0.9 1E-0	08 0. L1 (
TOTAL EXCAVATION WORKER EXPOSURE TO SURFACE AND SUBSURFACE SOIL	2E-07	1 464	7

TABLE 9-11 QUANTITATIVE RISK SUMMARY AOC 69W

REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS

	CEN	TRAL TENDENCY	RM	E
	ELCR	н	ELCR	Ш
UTURE USE (cont)		-		
SEDIMENT				
Incidental Ingestion of Downgradient Sediment: Pupil (6-18 years)	2E-07	0.01	6E-07	0.0
Dermal Contact with' Downgradient Sediment Pupil (6-18 years) TOTAL: PUPIL	<u>3E-07</u> 5E-07	<u>0.04</u> 0.05	<u>4E-07</u> 1E-06	<u>0.0</u> 0.0
PROUNDWATER (DISCHARGES TO SURFACE WATER)	100	1.1		
Incidental Ingestion of Surface Water: Pupil (6-18 years) Wading	7E-07 6E-07	0.02 0.2	2E-06 9E-07	0.0
Dermal Contact with Surface Water: Pupil (6-18 years) Wading TOTAL: PUPIL	1Ê-06	0.2	2E-06	0.
NDOOR AIR				
Inhalation of Vapors in Building: Pupil (6-18 years)] TOTAL: PUPIL	NC NC	0.4 4E-04	NC	<u>0.</u> 0.
TOTAL RISK TO PUPIL EXPOSED TO SURFACE SOIL, SEDIMENT, GROUNDWATER (DISCHARGES TO SURFACE WATER), and INDOOR AIR	6E-06	1.0	1E-05	
GROUNDWATER (POTABLE USE)				
Ingestion of Groundwater. Adult Resident	1E-04	4	3E-03	2
Ingestion of Groundwater, Child Resident TOTAL: RESIDENT	8 <u>E-05</u> 1E-04	8 NA	2E-03 5E-03	5 N

NOTES:

RME = Reasonable Maximum Exposure NC = Not Calculated because there were no potentially carcinogenic CPCs. NA = Not Additive 1E-06 = 1 in 1,000,000 ELCR = Excess Lifetime Cancer Risk HI = Hazard Index

TABLE 9–12 POTENTIAL SOURCES OF UNCERTAINTY AOC 69W

REMEDIAL INVESTIGATION REPORT FORT DEVENS, MA

UNCERTAINTY	EFFECT	JUSTIFICATION
Likelihood of exposure pathways	Overestimate	Future exposures may not actually occur
Degradation of chemicals not considered	Overestimate	Risk estimates are based on recent chemical concentrations. Concentrations will tend to decrease over time as a result of degradation, so future exposures may be to lower
		concentrations.
Extrapolation of animal toxicity data to humans	Unknown, probably overestimate	Animals and humans differ with respect to absorption, metabolism, distribution, and excretion of chemicals. The magnitude and direction of the difference will vary with each chemical. Animal studies typically involve high-dose exposures, whereas humans are exposed to low doses in the environment.
Use of linearized, multistage model to derive cancer slope factors	Overestimate	Model assumes a non-threshold, linear-at-low-dose relationship for carcinogens. Many compounds induce cancer by non-genotoxic mechanisms. Model results in a 95% uppe confidence limit of the cancer risk. The true risk is unlikely to be higher and may be as low as zero.
Summation of effects (cancer risks and hazard indices) from multiple substances	Unknown	The assumption that effects are additive ignores potential synergistic and/or antagnonistic effects. Assumes similarity in mechanism of action, which is not the case for many substances. Compounds may induce tumors or other toxic effects in different organs or systems.
Use of uncertainty factors in the derivation of reference doses.	Unknown	Ten-fold uncertainty factors are incorporated to account for various sources of uncertainty. Although some data seem to support the ten-fold factor, its selection is somewhat arbitrary.
Exclusion of analytes, such as TPHC and lead, from quantitative evaluation because no toxicity information is available.	Underestimate	The exclusion of analytes without toxicity values from quantitative evaluation may bias estimates of risk low.
The use of an oral absorption factor of 1	Overestimate	The assumption of 100% gastrointestinal absorption of chemicals on soil is conservative.
Dermal exposure to soil and sediment not evaluated because of lack of dermal absorption factors.	Underestimate	Dermal contact with soils and sediment may produce some incremental risk and the inability to quantify the risk may bias the total risk estimate low.

	E	Table 9-13 Endpoints for Ecological A AOC 69W	Assessment						
Remedial Investigation Report Devens, Massachusetts									
Medium	Receptor	Assessment Endpoint	Measurement Endpoint						
Surface Soil	Wildlife	Survival and propagation of wildlife populations.	Oral contaminant doses (mg/kg BW-day) based on mea- sured adverse effects on growth, reproduction, or survival (e.g., LD ₅₀ studies, LOAELs, and NOAELs) of mammali- an or avian laboratory test populations.						
	Terrestrial invertebrates	Survival and propagation of terrestrial invertebrate populations.	Contaminant concentrations in surface soil $(\mu g/g)$ that correlate with adverse effects on survival (e.g., LC_{50} stud- ies) of terrestrial invertebrates. When no survival studies are available, measured adverse effects on reproduction and growth are used.						
	Terrestrial plants	Survival and propagation of plant populations,	Contaminant concentrations in surface soil ($\mu g/kg$) that correlate with adverse effects on growth, reproduction, or survival of terrestrial plants.						
Sediment	Wildlife	Survival and propagation of wildlife populations.	Oral contaminant doses (mg/kg BW-day) based on mea- sured adverse effects on growth, reproduction, or survival (e.g., LD ₅₀ studies, LOAELs, or NOAELs) of mammalian or avian laboratory test populations.						
	Aquatic organisms (invertebrates, plants, and amphibians)	Survival and propagation of invertebrate, amphibian, and aquatic plant populations.	Chemical concentrations in sediment (µg/g) associated with adverse effects to growth, reproduction, and survival of aquatic organisms.						
	Aquatic invertebrates	Survival and propagation of benthic macroinvertebrate populations.	Direct measurement of survival and growth of the midge (Chironomus tentans) and amphipod (Hyalella azteca) in laboratory toxicity tests.						
Groundwater Discharge	Aquatic organisms (invertebrates, plants, and amphibians)	Future survival and pro- pagation of invertebrate, amphibian, and aquatic plant populations.	Chemical concentrations in surface water $(\mu g/l)$ associated with adverse effects to growth, reproduction, survival, and biodiversity of aquatic organisms.						
$LD_{50} = $ lethal dose $LC_{50} = $ lethal conc	is per gram	on population							

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN SURFACE SOIL¹

AOC 69W

Analyte	Frequency				F	lange o	of	Background	Contaminant	Average of	Expose	are Point
	of	R	lange of	•	I	Detecte	d	Surface Soil	of Potential	All	Conce	ntrations
Lange -	Detection ²	6	SQLs ¹		Con	centra	tions	Concentration ⁴	Ecological Concern?5	Concentrations*	RME'	Average
PAL METALS (ug/g)												
Aluminum	6/6		NA		5,210		6,160	18,000	NO ⁹	6,160		
Arsenic	6/6		NA		7.7	÷	18.0	19	NO ⁹	12		
Barium	6/6		NA		14.1		22.4	54	NO ⁹	18.2		
Beryllium	1/6		0.50		0.62		0.85	0.81	YES	0.41	0.85	0.41
Calcium	6/6		NA		333		908	810	NO ¹⁰	680		
Chromium	6/6		NA		12.1		28.1	33	NO ⁹	18		
Cobalt	6/6		NA		2.5		5.4	4.7	YES	4.1	5.4	4.1
Copper	6/6		NA		5.6		29.9	13.5	YES	12	29.9	12
Iron	6/6		NA		6,780		10,300	18,000	NO ^{9,10}	8,820		
Lead	6/6		NA		11.4		238	48	YES	61.2	238	61.2
Magnesium	6/6		NA		1,360		2,670	5,500	NO ^{9,10}	2,405		
Manganese	6/6		NA		52.4	÷	240	380	NO ⁹	167		
Mercury	2/6		0.050		0.076	+	0.078	NA	YES	0.042	0.078	0.042
Nickel	6/6		NA		6.0		18.1	14.6	YES	13.3	18.1	13.3
Potassium	6/6		NA		367		993	2,400	NO ^{9,10}	630		
Selenium	1/6		0.25			0.36		NA	YES	0.16	0.51	0.16
Sodium	6/6		NA		241		506	131	NO ¹⁰	347		
Vanadium	6/6		NA		10.6		19.1	32.3	NO ⁹	14.1		
Zinc	6/6		NA		18.9		71.7	43.9	YES	32.5	71.7	32.5
PAL SEMIVOLATILE ORO	GANICS (ug/g)											
Acenaphthylene	1/6	0.033	2	3.0		2.0		NA	YES	0.71	2.0	0.71
Anthracene	1/6	0.033		3.0		1.0		NA	YES	0.54	1.0	0.54
Benzo[k]fluoranthene	1/6	0.066		7.0		2,0		NA	YES	1.09	2.0	1.09
Chrysene	2/6	0.12		10	0.17	-	5.0	NA	YES	2.04	5.0	2.04
Fluoranthene	4/6	0.068		1.0	0.19		9.0	NA	YES	3.3	9.0	3.3
Fluorene	1/6	0.033		3.0		1.0		NA	YES	0.54	1.0	0.54
Phenanthrene	5/6	0.20		0.70	0.065		9.0	NA	YES	3.09	9.0	3.09
Pyrene	5/6	0.20		0.70	0.075	-	10	NA	YES	3.8	10	3.8

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN SURFACE SOIL¹

AOC 69W

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

Analyte	Frequency of	Range of	Range of Detected	Background Surface Soil	Contaminant of Potential	Average of All		are Point ntrations
	Detection ²	SQLs ³	Concentrations	Concentration ⁴	Ecological Concern?5	Concentrations"	RME ⁷	Average
PAL VOLATILE ORGANICS	S (ug/g)							
Acetone	1/6	0.017	0.069	NA	YES	0.019	0.069	0.019
Toluene	3/6	0.00078	0.00096 - 0.0021	NA	YES	0.00091	0.0021	0.00091
Trichlorofluoromethane	2/6	0.0059	0.0055 - 0.0072	NA	YES	0.0041	0.0077	0.0041
Xylenes	1/6	0.0015	0.0027	NA	YES	0.0011	0.0027	0.0011
OTHER (ug/g)								
ТРН	5/6	28	53 - 940	NA	YES	390	940	390

NOTES:

¹ Sample locations include ZWS-95-35X, ZWS-95-39X, ZWS-95-42X, ZWS-95-45X, ZWS-95-46X, and ZWS-95-47X.

² Frequency of detection is equal to the number of samples in which the analyte is detected in relation to the total number of samples analyzed

³ Sample Quantitation Limits (SQLs) are equal to the detection limit adjusted for percent moisture (solid media only) and dilutions (if any are performed).

⁴95th percentiles of inorganic background concentrations from the Ft Devens background surface soil database (developed in 1993) were used to screen CPCs

⁵ Contaminant of Potential Concern (CPC) for wildlife receptors.

⁶ The average of all concentrations assigns a value of 1/2 the detection limit to all non-detects.

⁷ Reasonable Maximum Exposure (RME) concentrations are equal to the maximum detected concentration; the 95th percent UCL was not calculated because there are fewer than 10 samples in the data set.

*Average Exposure Point Concentrations (EPCs) are equal to the average of all concentrations. If the average is greater than the RME, then the RME was used instead.

⁹ Maximum analyte concentration is less than the background concentration.

10 Analyte is an essential nutrient, and not considered toxic except at high concentrations.

AOC = Area of Contamination

NA = Not available

ug/g = micrograms per gram

EPC = exposure point concentration

TPH = total petroleum hydrocarbons

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN UPGRADIENT SEDIMENT¹

AOC 69W

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

Analyte	Frequency of	Range of		inge o stectes		Background Surface Soil	Contaminant of Potential	Average of All		sure Point entrations
Detection ²		SQLs ³	Concentrations			Concentration ⁴	Ecological Concern? ⁵	Concentrations ⁶	RME ⁷	Average
PAL METALS (ug/g)										1
Aluminum	3/3	NA	3,120		6,010	18,000	NO ⁹			
Arsenic	3/3	NA	6.4	+	10.0	19	NO ⁹			
Barium	3/3	NA	9.5		18.6	54	NO ⁹			
Calcium	3/3	NA	674		992	810	NO ¹⁰			
Chromium	3/3	NA	10.5		35.5	33	YES	17.4	35.5	17.4
Cobalt	3/3	NA	2.0		4.2	4.7	NO ⁹			
Copper	3/3	NA	7.9		25.1	13.5	YES	16.1	25.1	16.1
Iron	3/3	NA	7,020	Υ.	15,400	18,000	NO ⁹			
Lead	3/3	NA	29.0		55.9	48	YES	39.2	55.9	39.2
Magnesium	3/3	NA	1,670		3,100	5,500	NO ^{9,10}			
Manganese	3/3	NA	93.6		230	380	NO ⁹			
Nickel	3/3	NA	8.7		17.3	14.6	YES	11.8	17.3	11.8
Potassium	3/3	NA	371	2	783	2,400	NO ^{9,10}		3	
Sodium	3/3	NA	309		330	131	NO ¹⁰			
Vanadium	3/3	NA	7.4	-	15.2	32.3	NO ⁹		C	
Zinc	3/3	NA	39.7		71.4	43.9	YES	50,5	71.4	50.5
PESTICIDES/PCBs (ug/g)										
alpha-Chlordane	1/3	0.0050	C	0.013		NA	YES	0.0060	0.013	0.0060
Aroclor-1260	1/3	0.080		0.51		NA	YES	0.20	0.51	0.20
4,4'-DDD	3/3	NA	0.027	+	2.1	NA	YES	0.75	2.1	0.75
4,4'-DDE	3/3	NA	0.012		0.081	NA	YES	0.036	0.081	0.036
4,4'-DDT	3/3	NA	0.044	*	0.40	NA	YES	0.19	0.40	0.19
Dieldrin	1/3	0.0063	(0.060		NA	YES	0.022	0.060	0.022
Endosulfan II	1/3	0.0066	.0	0.050		NA	YES	0.019	0.050	0.019
gamma-Chlordane	1/3	0.0050	(0.024		NA	YES	0.0097	0.024	0.0097

1.00

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN UPGRADIENT SEDIMENT⁴

AOC 69W

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

Analyte	Frequency of		Range of	r		ange ()etecte		Background Surface Soil	Contaminant of Potential	Average of All	0.00	ure Point ntrations
and the second	Detection ²	_	SQLs ³		Con	centra	tions	Concentration ⁴	Ecological Concern? ⁵	Concentrations ⁶	RME ⁷	Average
PAL SEMIVOLATILE OR	GANICS (ug/g)											
Benzo[k]fluoranthene	1/3	0.30		2.0		0.80		NA	YES	0.65	0.80	0.65
Chrysene	1/3	1.0	-	3.0	0.90	1.4	1.0	NA	YES	0.98	1.0	0.98
Fluoranthene	2/3		2.0		1.0		2.0	NA	YES	1.5	2.0	1.5
Phenanthrene	2/3		0.80		0.60		0.80	NA	YES	0.57	0.80	0.57
Pyrene	3/3		NA		1.0		2.0	NA	YES	1.5	2.0	1.5
PAL VOLATILE ORGANI	CS (ug/g)											
Methylene chloride	1/3		0.012			0.014		NA	YES	0.0073	0.014	0.0073
Trichlorofluoromethane	2/3		0.0059		0.0080		0.011	NA	YES	0.0073	0.011	0.0073
TPH BY GC (ug/g)												
Diesel Fuel	3/3		NA		9.4		52	NA	YES	29	52	29
TPH, Gas Fraction	2/3		8.0		17	÷	22	NA	YES	14	22	14
OTHER (ug/g)										1		
TOC	3/3		NA		3,400		12,000	NA	NA			
ТРН	3/3		NA		360		1,200	NA	YES	740	1,200	740

NOTES:

¹ Sample locations include ZWD-95-01X, ZWD-95-04X, and ZWD-95-05X. A duplicate sample was collected at ZWD-95-01X.

² Frequency of detection is equal to the number of samples in which the analyte is detected in relation to the total number of samples analyzed.

³ Sample Quantitation Limits (SQLs) are equal to the detection limit adjusted for percent moisture (solid media only) and dilutions (if any are performed).

⁴95th percentiles of inorganic background concentrations from the Ft. Devens background surface soil database (developed in 1993) were used to screen CPCs.

⁵ Contaminant of Potential Concern (CPC) for wildlife and aquatic receptors.

⁶ The average of all concentrations assigns a value of 1/2 the detection limit to all non-detects.

⁷ Reasonable Maximum Exposure (RME) concentrations are equal to the maximum detected concentration; the 95th percent UCL was not calculated because there are fewer than 10 samples in the data set.

*Average Exposure Point Concentrations (EPCs) are equal to the average of all concentrations. If the average is greater than the RME, then the RME was used instead

⁹ Maximum analyte concentration is less than the background concentration.

¹⁰ Analyte is an essential nutrient, and not considered toxic except at high concentrations.

AOC = Area of Contamination NA = not available GC = gas chromatography TOC = total organic carbon TPH = total petroleum hydrocarbons ug/g = micrograms per gram EPC = exposure point concentration

χ.,

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN DOWNGRADIENT SEDIMENT¹

AOC 69W

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

Analyte	Frequency		R	ange o	of	Background	Contaminant	Average of	Expose	ure Point
	of	Range of	D	etecte	d	Surface Soil	of Potential	All	Conce	ntrations
	Detection ²	SQLs ³	Concentrations			Concentration ⁴	Ecological Concern? ⁵	Concentrations ⁶	RME ⁷	Average
PAL METALS (ug/g)							14. C			
Aluminum	3/3	NA	2,930		4,840	18,000	NO ⁹			
Arsenic	3/3	NA	6.4		14.0	19	NO ⁹	-		
Barium	3/3	NA	7.1		11.4	54	NO ⁹			
Calcium	3/3	NA	536	*	736	810	NO ^{9,10}			
Chromium	3/3	NA	11.5	*	16.1	33	NO ⁹			
Cobalt	3/3	NA	2.8		6.9	4.7	YES	4.5	6.9	4.5
Copper	3/3	NA	10.4		23.4	13.5	YES	14.9	23.4	14.9
Iron	3/3	NA	8,030		10,900	18,000	NO ⁹			
Lead	3/3	NA	15.0	71	30.0	48	NO ⁹			
Magnesium	3/3	NA	1,580		2,630	5,500	NO ^{9,10}			
Manganese	3/3	NA	93.9		186	380	NO ⁹			
Nickel	3/3	NA	9.6		18.1	14.6	YES	13.1	18.1	13.1
Potassium	3/3	NA	364		426	2,400	NO ^{9,10}			
Sodium	3/3	NA	259		307	131	NO ¹⁰			
Vanadium	3/3	NA	8.5		10.4	32.3	NO ⁹			
Zinc	3/3	NA	27.9	-	39.6	43.9	NO ⁹			
PESTICIDES/PCBs (ug/g)								1.000		
4,4'-DDD	3/3	NA	0.017	4	0.12	NA	YES	0.068	0.12	0.068
4,4'-DDE	1/3	0.0077		0.015		NA	YES	0.0076	0.015	0.0076
4,4'-DDT	3/3	NA	0.021		0.046	NA	YES	0.031	0.046	0.031
PAL SEMIVOLATILE OF	RGANICS (ug/g)									
Benzo[k]fluoranthene	1/3	0.30		0.40		NA	YES	0.23	0.40	0.23
Fluoranthene	2/3	0.30		1.0		NA	YES	0.72	1.0	0.72
Phenanthrene	2/3	0.20	0.50		0.90	NA	YES	0.50	0.90	0.50
Pyrene	2/3	0.20		1.0		NA	YES	0.70	1.0	0.70

12.1

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN DOWNGRADIENT SEDIMENT¹

AOC 69W

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

Analyte	Frequency of	Range of		ange (etecte		Background Surface Soil	Contaminant of Potential	Average of All		ure Point ntrations
	Detection ²	SQLs ³	Conc	entra	tions	Concentration ⁴	Ecological Concern? ⁵	Concentrations ⁶	RME ⁷	Average
PAL VOLATILE ORGANI	CS (ug/g)									
Trichlorofluoromethane	3/3	NA	0.0082		0.0096	NA	YES	0.0091	0.0096	0.0091
TPH BY GC (ug/g)										
Diesel Fuel	1/3	8.0		13		NA	YES	6.9	13	6.9
TPH, Gas Fraction	1/3	8.0		19		NA	YES	9.1	19	9.1
OTHER (ug/g)								1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1		
TOC	3/3	NA	2,400		5,400	NA	NA			
ТРН	3/3	NA	130		290	NA	YES	200	290	200

NOTES:

¹ Sample locations include ZWD-95-02X, ZWD-95-03X, and ZWD-95-06X.

² Frequency of detection is equal to the number of samples in which the analyte is detected in relation to the total number of samples analyzed.

³ Sample Quantitation Limits (SQLs) are equal to the detection limit adjusted for percent moisture (solid media only) and dilutions (if any are performed).

⁴95th percentiles of inorganic background concentrations from the Ft. Devens background surface soil database (developed in 1993) were used to screen CPCs.

⁵ Contaminant of Potential Concern (CPC) for wildlife and aquatic receptors.

⁶ The average of all concentrations assigns a value of 1/2 the detection limit to all non-detects.

⁷ Reasonable Maximum Exposure (RME) concentrations are equal to the maximum detected concentration; the 95th percent UCL was not calculated because there are fewer than 10 samples in the data set.

*Average Exposure Point Concentrations (EPCs) are equal to the average of all concentrations. If the average is greater than the RME, then the RME was used instead.

⁹ Maximum analyte concentration is less than the background concentration.

¹⁰ Analyte is an essential nutrient, and not considered toxic except at high concentrations.

AOC = Area of Contamination

NA = not available

GC = gas chromotography

TOC = total organic carbon

TPH = total petroleum hydrocarbons

ug/g = micrograms per gram

EPC = exposure point concentration

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN GROUNDWATER'

AOC 69W

Analyte	Frequency					Rang	e of	Background	Average	Contaminant	Average of	Exposi	re Point
	of	R	ange of			Detec	ted	Groundwater	Upgradient	of Potential	All	Conce	ntrations
	Detection ²	1	SQLs ³	1.0	C	oncent	rations	Concentration ⁴	Concentration ⁵	Ecological Concern? ⁶	Concentrations ⁷	RME	Average
PAL UNFILTERED METALS	(ug/L)	-									1		
Aluminum	2/8		141		207		390	6,870	NA	NO ¹⁰			
Arsenic	6/8		2.5		5.2		190	10.5	NA	YES	50.1	190	50.1
Barium	8/8		NA		4.6	-	16.8	39.6	NA	NO ¹⁰			
Calcium	8/8		NA		16,000	-	25,000	14,700	22,700	NO ¹¹			
Iron	7/8		38.8		441	-	26,400	9,100	NA	YES	6,500	26,400	6,500
Lead	2/8		1.3		1.45	-	2	4	NA	NO ¹⁰			
Magnesium	8/8		NA		1,700		2,800	3,480	2,720	NO10,11	1.		
Manganese	8/8		NA		16		2,700	291	NA	YES	820	2,700	820
Potassium	8/8		NA		1,700		5,000	2,370	1,480	NO ^u			
Sodium	8/8		NA		23,500	4	38,000	10,800	34,800	NO ¹¹			
PAL FILTERED METALS (ug	/L)										1		
Arsenic	6/8		2.5		4.6		180	10,5	NA	YES	46.6	150	46.6
Barium	7/8		5.0		7	-	15,3	39.6	NA	NO ¹⁰			
Calcium	8/8		NA		15,300		24,300	14,700	22,300	NO			
Iron	7/8		38.8		447		22,600	9,100	NA	YES	5,670	22,600	5,670
Lead	2/8		1.3		1.64		1.67	4	NA	NO ¹⁰			
Magnesium	8/8		NA		1,700		2,680	3,480	2,680	NO ^{9,10,11}			
Manganese	8/8		NA		3.5	-	2,140	291	NA	YES	730	2,140	730
Potassium	8/8		NA		90		4,760	2,370	1,610	NO ¹¹			
Sodium	8/8		NA		21,700		38,100	10,800	32,400	NO			
PESTICIDES/PCBs (ug/L)											1 million 1		
gamma-Chlordane	1/8		0.075			0.1	2	NA	NA	YES	0.047	0.12	0.047
Heptachlor epoxide	1/8		0.025			0.0	59	NA	NA	YES	0.018	0.059	0.018
PAL SEMIVOLATILE ORGAN	NICS (ug/L)												
2-Methylnaphthalene	2/8		1.7		8.2	÷.,	550	NA	NA	YES	70	550	70
Acenaphthene	1/8	1.7		60		2.:	2	NA	NA	YES	4.7	2.2	2.2
Bis(2-ethylhexyl)phthalate	2/8		4.8		3.5	40	500	NA	6.5	YES	65	500	65
Dibenzofuran	1/8	1.7	14	60		2.4	1	NA	NA	YES	4.7	2.4	2.4
Diethylphthalate	3/8	2.0	4	110	2.03	-	3	NA	NA	YES	8.3	3	3
Naphthalene	2/8		0.50		16	-	200	NA	NA	YES	27	200	27
Phenanthrene	2/8		0.50		2.2	2	150	NA	NA	YES	19	150	19

ECOLOGICAL CONTAMINANTS OF POTENTIAL CONCERN IN GROUNDWATER¹

AOC 69W

Analyte	Frequency					Ran	ge of	Background	Average	Contaminant	Average of	Exposu	re Polnt
	of		Range	f		Dete	ected	Groundwater	Upgradient	of Potential	All	Concer	trations
	Detection ²		SQLs	Ê.	С	oncen	trations	Concentration ⁴	Concentration ⁵	Ecological Concern? ⁶	Concentrations ⁷	RME"	Average
PAL VOLATILE ORGANICS	(ug/L)			10.00	-								-
1,1,1-Trichloroethane	1/8	0.5		1.25		0.8	8	NA	NA	YES	0.38	0.88	0.38
Acetone	1/8	13	. 8.	37		13	1	NA	NA	YES	8.8	13	8.8
Ethylbenzene	3/8		0.50		1.9	-	17	NA	NA	YES	2.7	17	2.7
Toluenc	6/8		0.50		0.45		9	NA	NA	YES	0.85	9	0.85
Trichloroethylene	2/8	0.5		1.25	0.73	-	1.8	NA	NA	YES	0.55	1.8	0.55
Xylene	1/8	0.84	-	4		1.	7	NA	NA	YES	0.58	1.7	0.58
WET CHEMISTRY (ug/L)													
Alkalinity	8/8		NA		20,000	-	89,000	NA	18,000	YES	45,000	89,000	45,000
Chloride	8/8		NA		27,000	-	55,000	NA	66,000	NO ¹⁰			
Nitrite	8/8		NA		760	1	3,000	NA	3,000	NA			
Nitrogen	10/13		180		200	-	1,200	NA	160	NA			
Phosphate	10/13		13		19		1060	NA	NA	NA			
Sulfate	7/8		10,000	t i	7,500	-	24,000	NA	17,000	NA			
Total Dissolved Solids	13/13		NA		140,000	-	220,000	NA	190,000	NA			
Total Hardness	13/13		NA		44,000		137,000,000	NA	36,000,000	NA			
Total Suspended Solids	9/13		4,000		5,000	-	37,000	NA	6,000	YES	11,000	37,000	11,000
OTHER (ug/L)													
Total Organic Carbon	1/5		1,000			6,1	00	NA	NA	NA			
Total Petroleum Hydrocarbons	11/16	170	1 G.	170,000	185	4	194,000	NA	NA	YES	35,000	194,000	35,000

REMEDIAL INVESTIGATION DEVENS, MASSACHUSETTS

¹ Sample locations include 69W-94-10 through 69W-94-14, ZWM-95-15X, ZWM-95-16X, and ZWM-95-18X through ZWM-95-20X. Duplicate samples were collected at 69W-94-11 and ZWM-96-19X. Data from two different rounds of sampling were available for 69W-94-10, 69W-94-11, and ZWM-95-16X. Filtered data are not available for ZWM-95-19X and ZWM-95-20X.

² Frequency of detection is equal to the number of samples in which the analyte is detected in relation to the total number of samples analyzed

¹ Sample Quantitation Limits (SQLs) are equal to the detection limit adjusted for percent moisture (solid media only) and dilutions (if any are performed).

⁴ 95th percentiles of inorganic background concentrations from the Pt. Devens background groundwater database (developed in 1993) were used to screen CPCa.

⁵Average upgradient groundwater concentration from monitoring wells ZWM-95-17X and ZWM-96-21X.

⁶ Contaminant of Potential Concern (CPC) for aquatic receptors.

⁷ The average of all concentrations assigns a value of 1/2 the detection limit to all non-detects.

⁶ Reasonable Maximum Exposure (RME) concentration is equal to the maximum detected concentration; the 95th percent UCL is not presented

* Average EPCs are equal to the arithmetic mean of all concentrations When the arithmetic mean is greater than the RME, the RME is used instead

10 Maximum analyte concentration is less than the background concentration.

¹¹ Analyte is an essential nutrient, and not considered toxic except at high concentrations.

12 Maximum analyte concentration is less than the upgradient concentration

NOTES:

AOC = Area of Contamination NA = not available/not applicable ug/L = micrograms per Liter EPC = exposure point concentration UCL = upper confidence limit

Table 9-18 Ecological Receptors Evaluated at AOC 69W Remedial Investigation Report Devens, Massachusetts										
	Recepto	r Evaluated	Me	dia .						
Method of Evaluation	Common Name	Scientific Name	Surface Soil	Sediment						
Food Web Modeling	White-footed mouse	Peromyscus leucopus	x							
	Short-tailed shrew	Blarina brevicauda	x	х						
	American robin	Turdus migratorius	x							
	Red-winged blackbird	Agelaius phoeniceus	x	x						
	Raccoon	Procyon lotor	x	х						
Benchmark Comparison	Terrestrial plants		x							
	Soil invertebrates	10	x							
	Aquatic plants			х						
	Benthic invertebrates			x						
	Amphibians			x						
Toxicity Testing	Midge	Chironomus tentans		х						
	Amphipod	Hyalella azteca		x						

Table 9-19 Model for Estimation of Contaminant Exposures for Representative Wildlife Species AOC 69W											
	Remedial Investigation Report Devens, Massachusetts										
Estimation of Contaminant Exposu	ures Related to Surface Soll										
Description:	Estimates the amount (dose) of a contaminant ingested and accumulated by a species via incidental ingestion of contaminated surface soil and ingestion of contaminated food items.										
Soil Contaminant Concentration:	Maximum: The maximum detected concentration of the ecological chemicals of potential concern (CPC) when the sample size is ≤ 9, and the lesser of the maximum detected concentration or the 95th percent upper confidence limit (UCL) when the sample size is ≥ 10.										
	Average: Average of all concentrations. If the average is greater than the maximum exposure point concentration (EPC), the maximum EPC was selected.										
Soil Exposure:	Soil Exposure = (% of Diet Soil (mg/kg) (mg/kg) (mg/kg)										
Concentration of a Contaminant in Primary Prey Items (T,):	Primary Prey Item = (BAF _{inv or plant} x Concentration) Concentration = (mg/kg) (mg/kg)										
Concentration of a Contaminant in Secondary Prey Items (T _N):	Secondary Tissue Prey Item = (BAF _{man or bird} x Prey Items*) (mg/kg) (mg/kg)										
	where BAF = Bioaccumulation Factor or mg/kg fresh weight tissue over mg/kg dry weight soil for invertebrates and plants, and mg/kg fresh weight tissue over mg/kg fresh weight food for small mammals and small birds.										
(1)	' For a discussion of the weighted contaminated concentration in prey items, see explana- tion of the PDE term below										
Total Exposure Related to Surface Soil:	$\frac{PDE}{(mg/kgBW-day)} = \frac{\left[P_1 \times T_1 + \ldots + P_N \times T_N + \frac{soil}{exposure}\right] \times IR_{piet} \times SFF \times ED}{BW}$										
	 where PDE = Potential Dietary Exposure (mg/kgBW-day), P_N = percent of diet composed of food item N, T_N = contaminant concentration in food item N (mg/kg), IR_{Diet} = food ingestion rate of receptor (kg of food or dietary item per day), BW = body weight (kg) of receptor, SFF = Site Foraging Frequency (site area [acres] divided by home range [acres]), which cannot exceed 1, and ED = Exposure Duration (fraction of year species is expected to occur onsite). 										

Model for Estima	Table 9-19 Ition of Contaminant Exposures for Representative Wildlife Species AOC 69W					
Remedial Investigation Report Devens, Massachusetts						
Estimation of Contaminant Ex	posures Related to Sediment					
Description:	Estimates the amount of a contaminant ingested and accumulated by a species resulting from incidental ingestion of sediment.					
Contaminant Concentration:	Same as described above for soil.					
Sediment Exposure:	Sediment Sediment IR _{pier}) Exposure = (as Sediment x Concentration x (kg/day)) (mg/day) (mg/kg)					
	where IR _{DIET} = dietary Ingestion rate of receptors (kg of diet per day)					
Total Exposure Related to Sediment:	$\frac{PDE}{(mg/kgBW-day)} = \frac{\frac{Sediment}{(Exposure + P_A \times T_A \times IR_{DIBT}) \times SFF \times ED}{(mg/day)}}{BW}$					
	where P_A = percent of diet comprised of aquatic prey items T_A = tissue concentration in aquatic prey items (mg/kg)					
	Note: Food chain exposures from incidental ingestion of sediment are expected to be minimal. Food chain exposures from ingestion of aquatic prey items are evaluated in the same way as primary prey items. Food chain exposures for consumers eating secondary prey items (i.e., small mammals and birds) are estimated by adding the body-weight normalized doses of the prey items.					
Notes:						
AOC = Area of contamination mg/kg = milligrams per kilogi kg/day = kilograms per day kg = kilograms % = percent mg/kg BW-day = milligrams p	ram ber kilogram of body weight per day					

		Remedial	Table 9-20 Bioaccumulation Factors AOC 69W Investigation Report as, Massachusetts
Recepto	or Group	Nature of Approach	General Approach
Terrestr	ial Receptors		
Plants			
Unit:	mg/kg wet tissue per mg/kg dry soil	Literature Values	When available, literature values were used to estimate plant BAFs. Evidence from the literature (Levine <u>et al.</u> , 1989) suggests that lead does not bioaccumulate in plant tissue; therefore, a BAF of zero was assigned (i.e., a zero does not imply that literature information is lacking).
		SAR	When literature values were not available, plant BAFs for semivolatile organic compounds (SVOCs) were calculated using a regression equation based on the uptake of organic chemicals into plant tissue from Travis and Arms (1988). ¹
		Extrapolation and Empirical Data	When literature values were not available, plant BAFs for inorganic compounds were obtained from Baes et al. (1984). ²
		Assumption	Although evidence suggests that plants may transport organic analytes with log K _w s < 5 (i.e., volatile organic compounds [VOCs]) from the roots into leafy portions (Briggs <u>et al.</u> , 1982; Briggs <u>et al.</u> , 1983), bioaccumulation data for VOCs is generally lacking in the scientific literature. In addition, evidence in the literature (Suter, 1993; Maughan, 1993) suggests that analytes with log K _w s < 3.5 are not bioaccumulated into animal tissue. Therefore, it was assumed that transfer of VOCs from plant tissue to animal tissue does not occur.
Terresti	ialinvertebrates		
Ünit:	mg/kg wet tissue per mg/kg dry soil	Literature Values	Literature values were used to estimate BAFs for Invertebrates.
		Assumption	Earthworm data were used to represent all invertebrates.
		Empirical Data and Assumption	A single BAF for PAHs was calculated using data presented in Beyer (1990); dry weight was converted to wet weight assuming earthworms are 80 percent water.
		Surrogate Values	When no literature values for invertebrates were available, mammal values were used as surrogates.
		Assumption	Bioaccumulation data for VOCs is generally lacking in the scientific literature. In addition, evidence in the literature (Suter, 1993; Maughan, 1993) suggests that analytes with log $K_{\rm w}$ s < 3.5 are not bioaccumulated into animal tissue. Therefore, it was assumed that soll invertebrates do not bioaccumulate VOCs.

.

4

Table 9-20 Estimation of Bioaccumulation Factors AOC 69W Remedial Investigation Report Devens, Massachusetts					
Recept	or Group	Nature of Approach	General Approach		
<u>Small N</u> Unit:	<u>Mammals</u> mg/kg wet tissue per mg/kg wet food	Literature Values	When available, literature values were used to estimate BAFs for small mammals.		
		SAR	When literature values were not available for SVOCs, BAFs for small mammals were estimated using a regression equation based on the uptake of organic chemicals into beef tissue from Travis and Arms (1988) ³ .		
		Extrapolation/ Empirical Data	When literature values were not available, BAFs for small mam- mals for inorganics were derived from ingestion-to-beef biotransfer factors (BTFs) presented in Baes et al. (1984) ⁴ .		
		Assumption	Bioaccumulation data for VOCs are generally lacking in the scientific literature. In addition, evidence in the literature (Suter, 1993; Maughan, 1993) suggests that analytes with log K_{ov} s < 3.5 are not bioaccumulated into animal tissue. Therefore, it was assumed that small mammals do not bioaccumulate VOCs because of their generally low K_{ov} s.		
Small B	Birds				
Unit:	mg/kg wet tissue per mg/kg wet food	Literature Values	When available, literature values were used to estimate BAFs for small birds.		
		Surrogate Values	BAFs were not available for many SVOCs or inorganic compounds as there is little bioaccumulation data available for birds. In these situations, mammal data were used as surrogates. It was assumed that small birds do not accumulate VOCs because of their generally low $K_{ov}s$.		
Semi-ad	quatic Receptors				
Sedime	nt				
Unit:	mg/kg wet tissue per mg/kg wet sediment	Literature Values	Literature values were used to estimate BAFs for aquatic plants and macroinvertebrates. When literature values were not available, terrestrial plant and invertebrate BAFs were used instead.		
		Assumption and Surrogate Values	Bioaccumulation data for VOCs are generally lacking in the scientific literature. Therefore, it was assumed that semi-aquatic wildlife do not bioaccumulate VOCs. Little to no literature values exist for sediment to biota accumulation; when these data were not available, small mammal values were used as surrogates.		

Table 9-20 Estimation of Bioaccumulation Factors AOC 69W Remedial Investigation Report Devens, Massachusetts						
Recep	otor Group	Nature of Approach	General Approach			
1		lated using the following Travis and Arms to 0.578 log K	(1988) regression:	*		
	chemical and ph vegetative and re weight basis ass	m Baes et al. (1984). Values are based or ysical parameters, or comparisons of obse productive plant material and soil. Data a uming that plants are 80 percent water. T	erved and predicted elemental concentrat	ions in ted to a fresh		
	contain a much I	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores.		93). Grains		
3	contain a much I underestimates e Small mammal E log BTF = log K	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis ow - 7.6	10 percent), therefore, this assumption lik and Arms (1988) regression:	93). Grains		
3	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		
3 4 Notes:	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve kg (dry weight) p	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis 7.6 otransfer factor (mg/kg tissue divided by m orted to a BAF (mg/kg tissue divided by m	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		
4 Notes: mg/kg	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve kg (dry weight) p ; g = milligrams per ki	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis ow- 7.6 transfer factor (mg/kg tissue divided by m orted to a BAF (mg/kg tissue divided by m ore day (average intake for lactating and n	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		
A Notes: mg/kg AOC =	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve kg (dry weight) p ; g = milligrams per ki Area of contaminat	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis ow- 7.6 transfer factor (mg/kg tissue divided by mored to a BAF (mg/kg tissue divided by more day (average intake for lactating and n logram	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		
4 Notes: mg/kg AOC = BAFs	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve kg (dry weight) p ; g = milligrams per ki = Area of contaminat = bioaccumulation fa	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis ow- 7.6 transfer factor (mg/kg tissue divided by m orted to a BAF (mg/kg tissue divided by m ore day (average intake for lactating and n logram ion actors	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		
A Notes: mg/kg AOC = BAFs PAH =	contain a much I underestimates e Small mammal E log BTF = log K where BTF = bio BTFs were conve kg (dry weight) p ; g = milligrams per ki Area of contaminat	percent water) and leafy vegetables (87 to ower percentage of water (approximately exposure to graminivores. BAFs calculated using the following Travis ow- 7.6 transfer factor (mg/kg tissue divided by m orted to a BAF (mg/kg tissue divided	10 percent), therefore, this assumption lik and Arms (1988) regression: ng chemical ingested per day). Ig/kg food) by multiplying by a food inge	993). Grains rely estion rate of 12		

Table 9-21 Results of Sediment Toxicity Testing 1 AOC 69W Remedial Investigation Report Devens, Massachusetts							
Sample Location		Midge ronomus tentans) ochronic Toxicity Test	(H	Amphipod Iy <i>alella azleca</i>) Acute Toxicity Test			
	Mean % Survival (Standard Deviation)	Mean Growth (mg dry weight) (Standard Devlation)	Mean % Survival (Standard Deviation)	Mean Growth (mg dry weigh (Standard Deviation)			
Control	74 (19)	1.70 (0.32)	64 (18) ²	0.10 (0.05)			
ZWD-95-02X	75 (15)	2.24 (0.85)	55 (24) ³	0.15 (0.07)			
ZWD-95-03X	88 (14)	2.94 (0.67)	66 (18)	0.10 (0.05)			
ZWD-95-06X	60 (19)	2.41 (0.93)	36 (23) ^{3,4}	0.11 (0.07)			
57D-95-08X (reference)	84 (12)	1.81 (0.30)	80 (21)	0.10 (0.03)			

² The control survival did not meet the acceptance criteria of 80%.
 ³ Amphipod survival in this sample was significantly less than the reference sample (57D-95-08X).
 ⁴ Amphipod survival in this sample was significantly less than the control.

AOC = Area of contamination Notes:

% = percent

mg = milligrams

69wtox.wp

Table 9-22 Results of Food-Web Modelling for Surface Soil and Sediment [a] AOC 69W Remedial Investigation Report Devens, Massachusetts							
Media Ev		Risk from Exposure to RME Concentrations Concentra		Primary Risk Contributor			
Surface S	Ecological Receptor						
	White-footed mouse	0.47	0.15	NA			
	Short-tailed shrew	2.8	0.88	Lead			
	American robin	0.29	0.14	NA			
	Red-winged blackbird	0.20	0.099	NA			
	Raccoon	0.0017	0.00051	NA			
Sediment							
	Short-tailed shrew	0.040	NE	NA			
	Red-winged blackbird	0.024	NE	NA			
	Raccoon	0.0000069	NE	NA			

Summary of I	Ecological Risk fo Remedial	Table 9-23 r Plants and AOC 69W Investigatio s, Massach	n Report	ates in Surface	e Soil	
Analyte		re Point rations ¹	RTV	⁷ (µg/g)	RTV Exceeded? ³ (by Max./by Ave.)	
	RME	Average	Plant ²	Invertebrate ²	Plant	Invertebrat
PAL Metals (µg/g)						
Beryllium	0.85	0.41	10	NA	No/No	NA
Cobalt	5.4	4.1	20	NA	No/No	NA
Copper	29.9	12.0	100	30	No/No	No/No
Lead	238	61.2	50	1,190	Yes/Yes	No/No
Mercury	0.078	0.042	0.3	36	No/No	No/No
Nickel	18.1	13.3	30	400	No/No	No/No
Selenium	0.51	0.16	1	NA	No/No	NA
Zinc	71.7	32.5	50	130	Yes/No	No/No
PAL Semivolatile Organics (µg/g)						
Acenaphthylene	2.0	0.71	25	34	No/No	No/No
Anthracene	1.0	0.54	25	34	No/No	No/No
Benzo(k)fluoranthene	2.0	1.09	25	34	No/No	No/No
Chrysene	5.0	2.04	25	34	No/No	No/No
Fluoranthene	9.0	3.3	25	34	No/No	No/No
Flourene	1.0	0.54	25	34	No/No	No/No
Phenanthrene	9.0	3.09	25	34	No/No	No/No
Pyrene	10	3.8	25	34	No/No	No/No
PAL Volatile Organics (µg/g)						
Acetone	0.069	0.019	NA	NA	NA	NA
Toluene	0.0021	0.00091	200	21	No/No	No/No
Trichlorofluoromethane	0.0077	0.0041	NA	NA	NA	NA
Xylenes	0.0027	0.0011	>1,000	21	No/No	No/No
Other (µg/g)						
Total Petroleum Hydrocarbons	940	390	NA	NA	NA	NA

Table 9-23 Summary of Ecological Risk for Plants and Invertebrates in Surface Soil AOC 69W Remedial Investigation Report Devens, Massachusetts							
Analyte	Exposure Point	Exposure Point Concentrations ¹		V (mg/g)	RTV Exceeded? ³ (by Max./by Ave.)		
	RME	Average	Plant ²	Invertebrate ²	Plant	Invertebrate	
² Plant and invertebrate RTVs a from among plant growth stud among chemicals in the same	dies on plants in solid me chemical class (applies to	dia, and invertebrate o organic compound	RTVs are the l s). A conservat	owest LC ₅₀ (14-day s ive factor of 0.2 was	oil test on Eise	enia foetida) from	
from among plant growth stud among chemicals in the same resultant value should be prot ³ Comparison shown is maximu AOC = Area of contamination RTV = Reference toxicity value ug/g = micrograms per gram LC ₅₀ = concentration lethal to 50	dies on plants in solid me chemical class (applies to ective of 99.9% of the po um EPC to RTV/average 0% of the test population	dia, and invertebrate o organic compound pulation from lethal EPC to RTV.	RTVs are the l s). A conservat	owest LC ₅₀ (14-day s ive factor of 0.2 was	oil test on Eise	enia foetida) from	
from among plant growth stud among chemicals in the same resultant value should be prot	dies on plants in solid me chemical class (applies to ective of 99.9% of the po um EPC to RTV/average 0% of the test population concentration	dia, and invertebrate o organic compound pulation from lethal EPC to RTV.	RTVs are the l s). A conservat	owest LC ₅₀ (14-day s ive factor of 0.2 was	oil test on Eise	enia foetida) fron	

	Comparison	of Downgra	dient Sediment Exp	AOC 69W		with Toxicity]	Benchmark Value	s ¹	
			Remedia Deve	l Investigatio ns, Massachu	n Report setts				
Analyte	Exposure Point Concentrations		USEPA Sediment Quality Guidelines ²	USEPA Sediment Quality	NOAA.3		OME LEL ⁴	NYSDEC LEL 5	Result
	Maximum	Average		ER-L	ER-M				
PAL Metals (µg/g)									
Cobalt	6.9	4.5	NA	NA	NA	NA	NA	No benchmark available	
Copper	23.4	14.9	NA	34	270	16	16	Exceeded	
Nickel	18.1	13.1	NA	20.9	51.6	16	16	Exceeded	
Pesticides/PCBs (µg/g)									
4,4'-DDD	0.12	0.068	⁶ 0.00828	⁶ 0.00158	⁸ 0.0461	0.008	⁶ 0.01	Exceeded	
4,4'-DDE	0.015	0.0076	⁶ 0.00828	0.0022	0.027	0.005	⁶ 0.01	Exceeded	
4,4'-DDT	0.046	0.031	0.00828	0.00158	0.0461	0.007	0.01	Exceeded	
PAL Semivolatile Organic	s (µg/g)								
Benzo(k)fluoranthene	0.40	0.23	NA	NA	NA	0.240	NA	Exceeded	
Fluoranthene	1.0	0.72	6.2	0.6	5.1	0.750	10.2	Exceeded	
Phenanthrene	0.90	0.50	1.8	0.24	1.5	0.560	1.19	Exceeded	
Ругеле	1.0	0.70	NA	0.665	2.6	0.490	NA	Exceeded	
PAL Volatile Organics (µg	g/g)								
Trichlorofluoromethane	0.0096	0.0091	NA	NA	NA	NA	NA	No benchmark available	
TPH by GC (µg/g)									
Diesel Fuel	13	6.9	NA	NA	NA	NA	NA	No benchmark available	
TPH Gas Fraction	19	9.1	NA	NA	NA	NA	NA	No benchmark available	
Other (µg/g)									
ТРН	290	200	NA	NA	NA	NA	NA	No benchmark available	

69SDBENCH.DOC

1

8/19/98

	Compariso	n of Downgr	Remedi	Table 9-24 posure Conc AOC 69W al Investigati ens, Massach	entrations	with Toxicity	Benchmark Values ¹	
Analyte	Exposure Point USEPA Concentrations Sediment Quality			NOAA ³		OME LEL 4	NYSDEC LEL 5	Result
	Maximum	Average		ER-L	ER-M	1		
 ² U.S. Environmenta (FCVs); when no F ³ National Oceanic a 90th percentile and 	CVs were available, Final and Atmospheric Administ the 50th percentile of the	PA, 1988) mean Residue Values ration (NOAA) E test populations,	Sediment Quality Criteria ((FRVs) were used instead. ffects Range-Low (ER-L) a respectively (Long et al., 19	nd Effects Rang 194).	e-Median (ER-	M) Sediment Guide	(TOC) of 1%. All values repre	tration that is protective of
 ² U.S. Environmenta (FCVs); when no F ³ National Oceanic a 90th percentile and ⁴ Ontario Ministry o of benthic organism ⁵ New York State D TOC content of 1% ⁶ Value for 4,4'-DD⁷ Notes: Shading indicates that 	 Il Protection Agency (USE CVs were available, Final and Atmospheric Administ I the 50th percentile of the f the Environment (OME) ns. epartment of Environment 6. The lowest effect levels I used as a surrogate. the concentration exceeds 	PA, 1988) mean Residue Values (ration (NOAA) E test populations, I Low Effects Leve al Conservation ((LELs) for metal	Sediment Quality Criteria ((FRVs) were used instead. (FRVs) were used instead. (Fects Range-Low (ER-L) a respectively (Long et al., 19 el (LEL) Provincial Sediment (LEL) Provincial Sediment NYSDEC) sediment criteria (s are also presented.	nd Effects Rang 194). nt Quality Guide	e-Median (ER- elines (Persaud	M) Sediment Guide et al., 1996) corres		tration that is protective of n be tolerated by the major
 ² U.S. Environmenta (FCVs); when no F ³ National Oceanic a 90th percentile and ⁴ Ontario Ministry o of benthic organism ⁵ New York State D TOC content of 1% ⁶ Value for 4,4'-DD Notes: 	 Il Protection Agency (USE CVs were available, Final and Atmospheric Administ I the 50th percentile of the f the Environment (OME) ns. epartment of Environment 6. The lowest effect levels I used as a surrogate. the concentration exceeds mination per gram 	PA, 1988) mean Residue Values (ration (NOAA) E test populations, I Low Effects Leve al Conservation ((LELs) for metal	Sediment Quality Criteria ((FRVs) were used instead. (FRVs) were used instead. (Fects Range-Low (ER-L) a respectively (Long et al., 19 el (LEL) Provincial Sediment (LEL) Provincial Sediment NYSDEC) sediment criteria (s are also presented.	nd Effects Rang 194). nt Quality Guide	e-Median (ER- elines (Persaud	M) Sediment Guide et al., 1996) corres	elines correspond to the concen pond to a concentration that ca	tration that is protective of n be tolerated by the major

Comparison of Groun	dwater Exposure Remedial	AOC 69W	tions with		Values ¹
Analyte			AWQC ² (µg/l)	AQUIRE Lowest Reported Adverse Effect Concentration ³ (µg/l) Test Species	Result
				(15)	1
PAL Unfiltered Metals (µg/l)				1.7.16.10	
Arsenic	190	50.1	190	1,700/water flea LC50	Not exceeded
Iron	26,400	6,500	1,000	3,700/duckweed growth	Exceeded
Manganese	2,700	820	NA	280/phytoplankton population endpoints	Exceeded
PAL Filtered Metals (µg/l)					
Arsenic	150	46.6	190	1,700/water flea LC_{50}	Not exceeded
Iron	22,600	5,670	1,000	3,700/duckweed growth	Exceeded
Manganese	2,140	730	NA	280/photoplankton population endpoints	Exceeded
Pesticides/PCBs (µg/l)					
gamma-Chlordane	0.12	0.047	4 0.0043	7.1/bluegill LC ₅₀	Exceeded
Heptachlor epoxide	0.059	0.018	0.0038	NA	Exceeded
PAL Semivolatile Organics (µg/l)					
2-Methylnaphthalene	550	70	NA	2,000,000/green algae growth	Not exceeded
Acenaphthalene	2.2	2.2	520	NA	Not exceeded
Bis(2-ethylhexyl)phthalate	500	65	⁵ 160	0.89/moorfrog hatchability	Exceeded
Dibenzofuran	2.4	2,4	NA	280	Not exceeded
Diethylphthalate	3	3	NA	NA	No benchmark available
Naphthalene	200	27	620	NA	Not exceeded
Phenanthrene	150	19	⁵ 6.3	NA	Exceeded
PAL Volatile Organics (µg/l)					
1,1,1-Trichlorothane	0.88	0.38	NA	1,300/Water flea reproduction	Not exceeded
Acetone	13	8.8	NA	550,000/water flea mortality	Not exceeded
Ethylbenzene	17	2.7	NA	4,600/green algae growth	Not exceeded
Toluene	9	0.85	NA	1,000/water flea reproduction	Not exceeded
Trichloroethylene	1,8	0.55	21,900	1,900/medaka mortality	Not exceeded
Xylene	1.7	0.58	NA	NA	No benchmark available

		Investigati 18, Massach			
Analyte	Exposure Point Concentrations		AWQC ² (µg/l)	AQUIRE Lowest Reported Adverse Effect Concentration ³ (µg/1) Test Species	Result
	Maximum	Average			
Wet Chemistry (µg/l)			-		
Alkalinity	89,000	45,000	20,000	NA	Exceeded
Chloride	55,000	42,000	230,000	NA	Not exceeded
Total Suspended Solids	37,000	11,000	NA	NA	No benchmark available
Other (µg/l)					
Total Petroleum Hydrocarbons	194,000	35,000	NA	NA	No benchmark available

However, for gamma-Chlordane, data was only available for bluegill.

⁴ Value for chlordane used as a surrogate.

⁵ Proposed criterion.

Notes:

AOC = Area of contamination

CPC = contaminant of potential concern

 $\mu g/l = micrograms per liter$

AWQC = Ambient Water Quality Criteria (guidance criteria established under the Clean Water Act)

NA = Not available

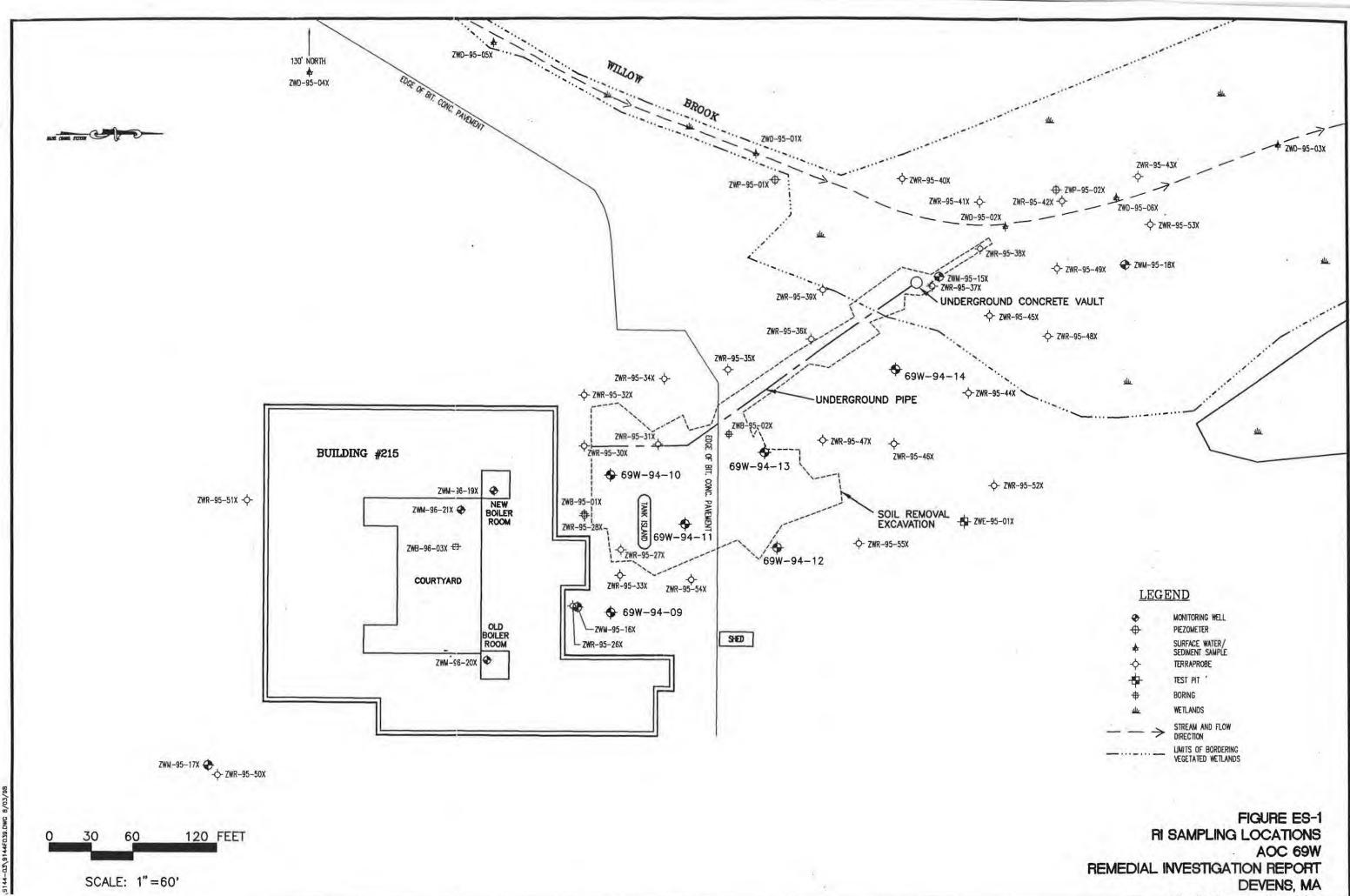
Shading indicates an exceedance of a toxicity benchmark value.

Table 9-26 Potential Sources of Uncertainty in Ecological Risk Assessment AOC 69W Remedial Investigation Report Devens, Massachusetts				
Uncertainties Associated with CPC Select	ion Process			
Degradation of chemicals not considered	Overestimate	Risk estimates are based on recent chemical concentrations. Concentrations will tend to decrease over time from degradation and the formation of daughter products.		
No evaluation of Tentatively Identified Compound (TIC) data	Underestimate	Risk was not calculated for potential exposure to TICs.		
Use of estimated data	Unknown	Using estimated data in the risk assessment may over- or underestimate the actual concentration of an analyte in site media.		
Uncertainties Associated with Exposure Assessment				
Surface soil sampling depths	Underestimate	Most terrestrial receptors wil be exposed only within the first six inches of soil where contaminant concentrations are typically greatest. Sampling the upper two feet of soil provides a diluted soil exposure concentration.		
Food chain assumed to occur at site	Unknown	Occurrence of the food chain used in the models at the sites is unknown.		
Food chain model exposure parameter assumptions	Unknown	Some exposure parameters are from the literature and some are estimated. Efforts were made to select exposure parameters representative of a variety of species or feeding guilds, so that exposure estimates would be representative of more than a single species.		
Assumption that receptor species will spend equal time at all habitats within home range	Unknown	Organisms will spend varying amounts of time in different habitats, thus affecting their overall exposures.		
Extrapolation of literature values from test species to representative wildlife species	Unknown	Species differ with respect to absorption, metabolism, distribution, and excretion of chemicals. The magnitude and direction of the difference will vary with each chemical.		
Organism-specific state variables	Underestimate	Surrogate laboratory animals are well- maintained and kept under controlled conditions. Field species must tolerate general environmental stressors that can exacerbate contaminant-induced stress.		

Table 9-26 Potential Sources of Uncertainty in Ecological Risk Assessment AOC 69W Remedial Investigation Report Devens, Massachusetts				
Consumption of contaminated prey	Unknown	Toxicity to receptors may result in sickness or mortality, thus making fewer prey items available to predators Predators may stop foraging in areas with reduced prey populations, or discriminate against, or, conversely, select contaminated prey. Furthermore anthropogenic sources of contamination may not even have as great an impact or the predator-prey relationship as do broader environmental stressors (e.g. climatic effects).		
No evaluation of dermal or inhalation exposure pathways	Underestimate	The dermal and inhalation exposure pathways are generally considered insignificant due to protective fur, feathers, chitinous exoskeletons, and the low concentration of contaminants under natural atmospheric conditions. However, under certain conditions, these exposure pathways may occur.		
Maximum exposure scenarios	Overestimate	It is unlikely any receptor would be exposed concurrently to maximum concentrations of all CPCs.		
Use of surrogate values for invertebrate BAFs	Underestimate	Bioaccumulation data for earthworms are lacking for several metals (e.g., alumi- num, antimony, barium, cobalt, manganese, and vanadium); therefore, mammal BAFs were used as surrogates. However, earthworms may actually bioaccumulate these metals to a greater degree than mammals.		
Continuous uptake and bioaccumulation of CPCs by soil biota	Unknown	Tissue and organ responses to CPC uptake are represented by a linear function which is an oversimplification of a more complex system (i.e., trophic states and lipid concentrations may affect bioaccumulation, or contaminants may only be seasonally available).		
Bioaccumulation of CPCs in leafy portions of plants	Overestimate	Ryan et al. (1988) states that compounds with log $K_{ows} > 5$ are unavailable to plants due to soil sorption. Compounds with log $K_{ows} > 5$ will be taken into the roots of plants, but are not easily transported into the leafy parts of plants (Briggs et al., 1982; 1983). The surface soil ingestion exposure model overesti- mates CPC exposure via plant ingestion to those receptors that only eat the leafy portions of plants. Levine et. al. (1989) suggests that lead does not bioaccumulate in plant tissue and so a BAF of zero was assigned.		

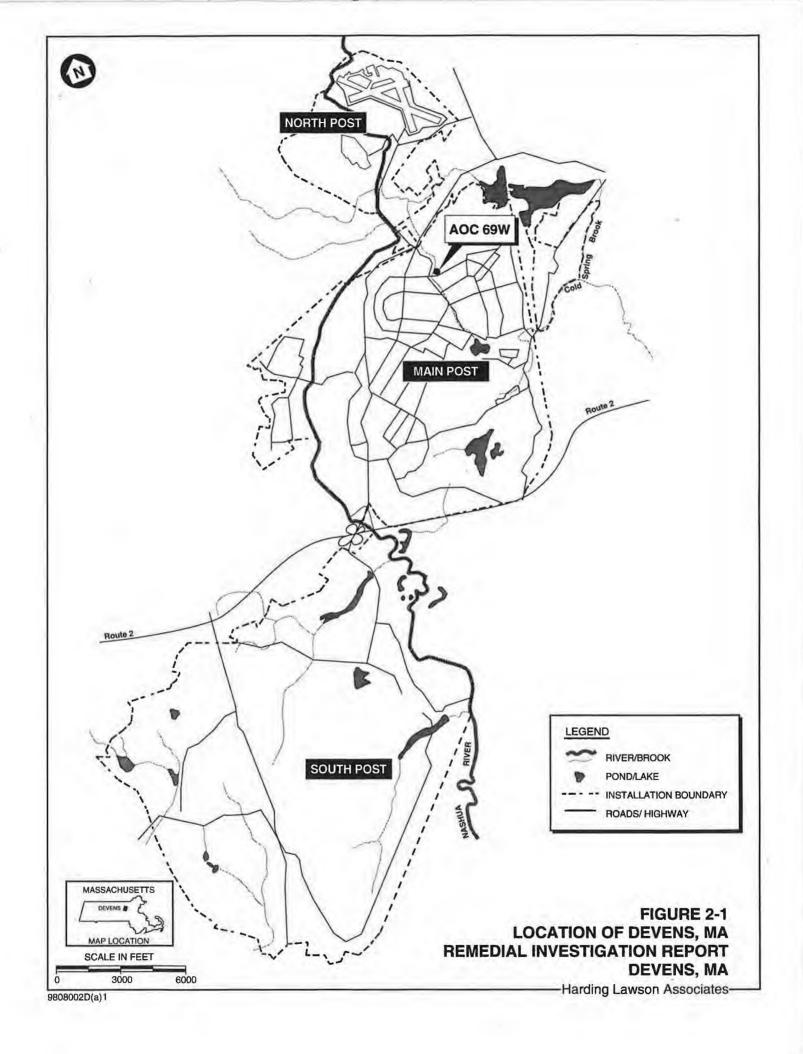
•

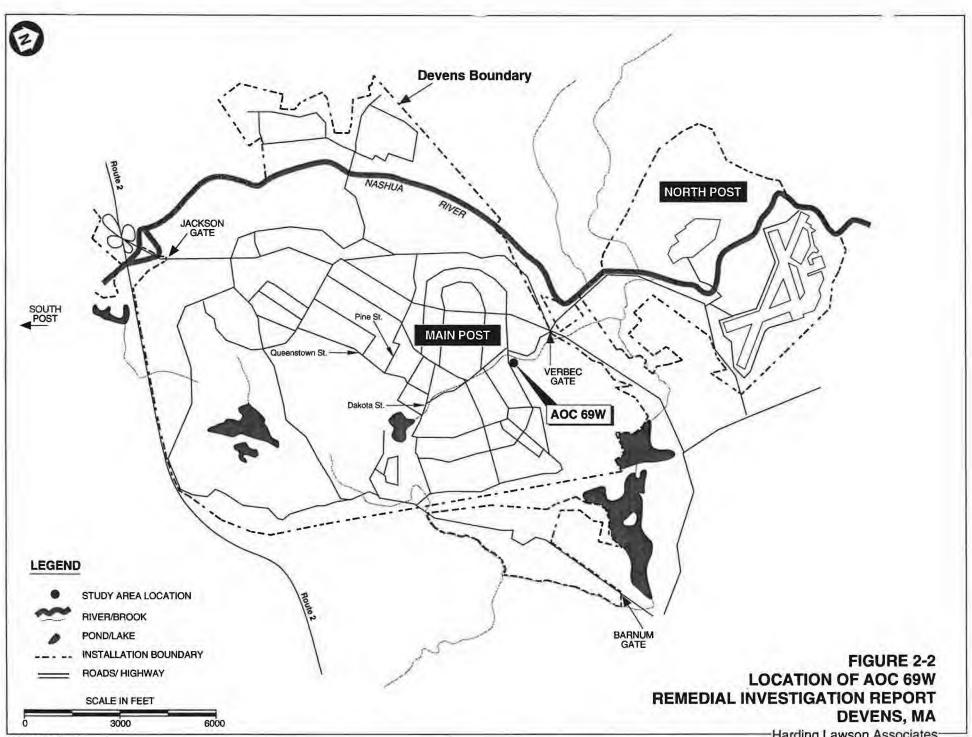
Table 9-26 Potential Sources of Uncertainty in Ecological Risk Assessment AOC 69W Remedial Investigation Report Devens, Massachusetts				
Seasonal changes in receptor foraging habits	Unknown		The food-chain model does not conside variations in a receptor's foraging habit due to seasonal changes and breeding.	
Relative uptake of inorganics by different plant species	Unknown		Estimated plant BAFs for certain inorganics were based on BAF data for leafy produce grown in sewage sludge. Variability in type of plant and substrate may make the chosen BAF values an overestimate or underestimate of actual uptake.	
Assumption of 1% TOC	Unknown		For simplicity, a 1% TOC is assumed, although a TOC of less than 1% exists a AOC 69W. This assumption may affect toxicity benchmark value adjustments and the number of benchmark exceedances.	
Uncertainties Associated with Effects				
Lack of ingestion toxicity information for reptile and amphibian species	Unknown		Information is not available on the toxicity of contaminants to reptiles of amphibians resulting from dietar exposures; as a result, dietary exposure to these receptors were not quantitativel evaluated in the AOC 69W ERA Assuming the toxicities of analytes to mammals and birds are similar for these receptors, and to the extent that the dietary exposures for reptiles and amphibians are the same as for the tertiary consumers evaluated in the AOC 69W ERA, an assumption can be made that dietary exposures to reptiles and an phibians would result in similar riss levels that were predicted for predator	
			mammals. However, risks to reptiles an amphibians are unknown.	
Use of measurement endpoints	Overestimate		Although an attempt was made to have measurement endpoints reflect asses- sment endpoints, limited available ecotoxicological literature resulted in the selection of certain measurement end- points that may overestimate assessment endpoints.	
Failure to address potential population-level effects	Underestimate		Bulk toxicity studies to assess population risks to the midge (<i>Chironomus tentans</i> and amphipod (<i>Hyalella azteca</i>) do not address the issue of effects on community structure and biodiversity.	



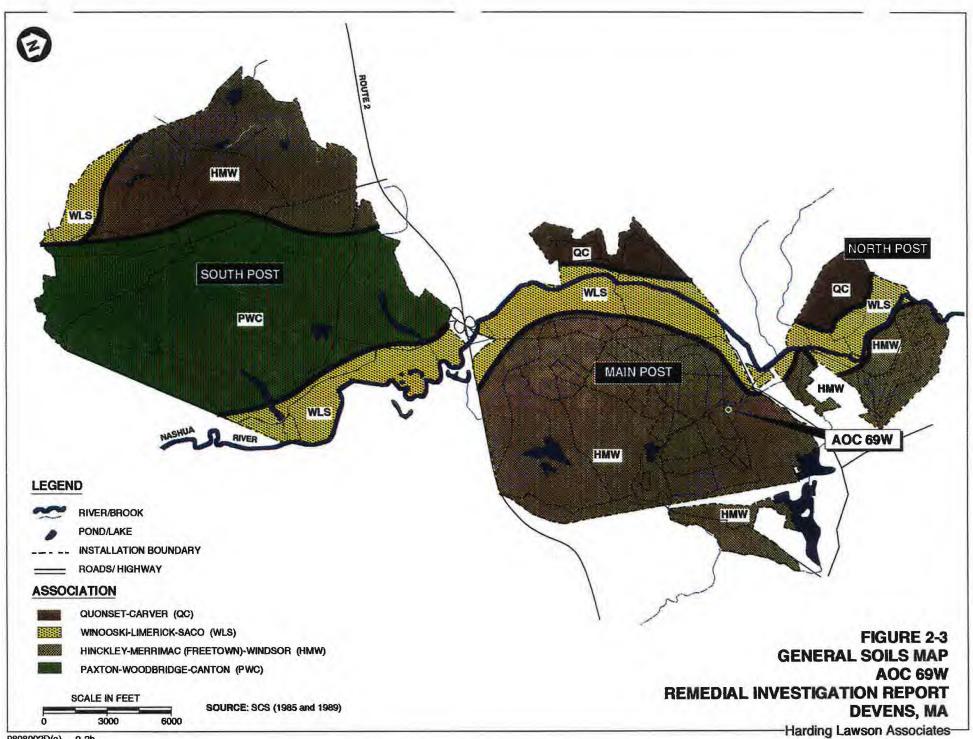
•	MONITORING WELL
\$	PIEZOMETER
*	SURFACE WATER/ SEDIMENT SAMPLE
- 0 -	TERRAPROBE
+	TEST PIT '
+	BORING
ste	WETLANDS
>	STREAM AND FLOW DIRECTION
	LIMITS OF BORDERING VEGETATED WETLANDS

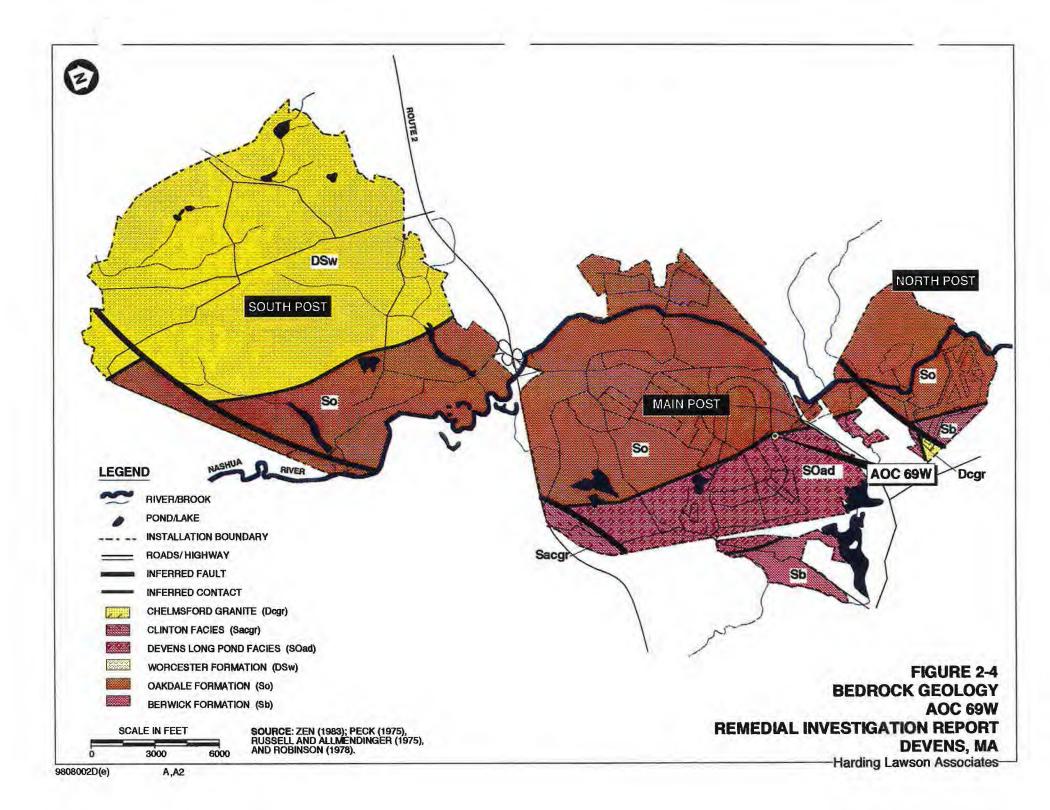
- Harding Lawson Associates

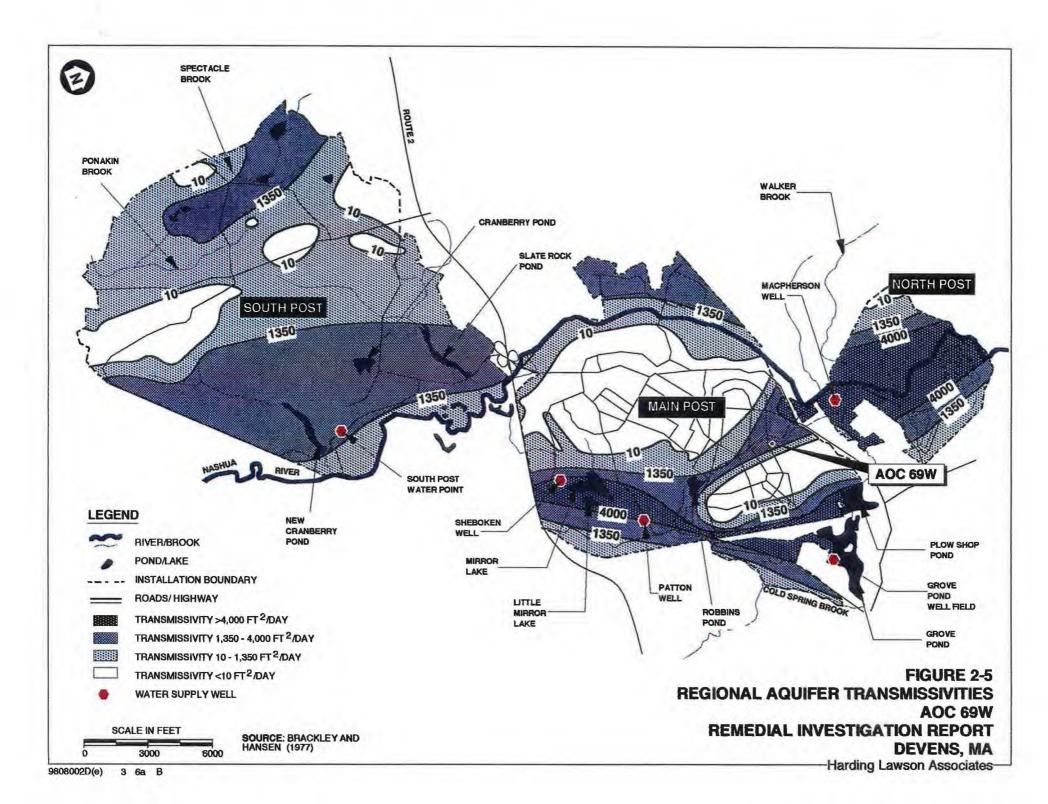


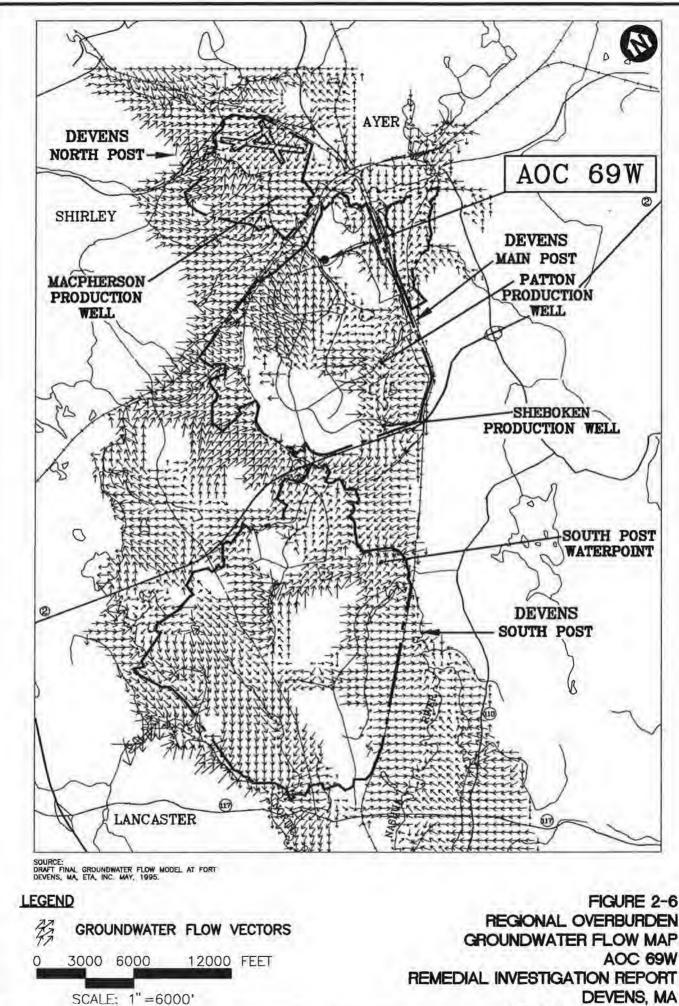


-Harding Lawson Associates-



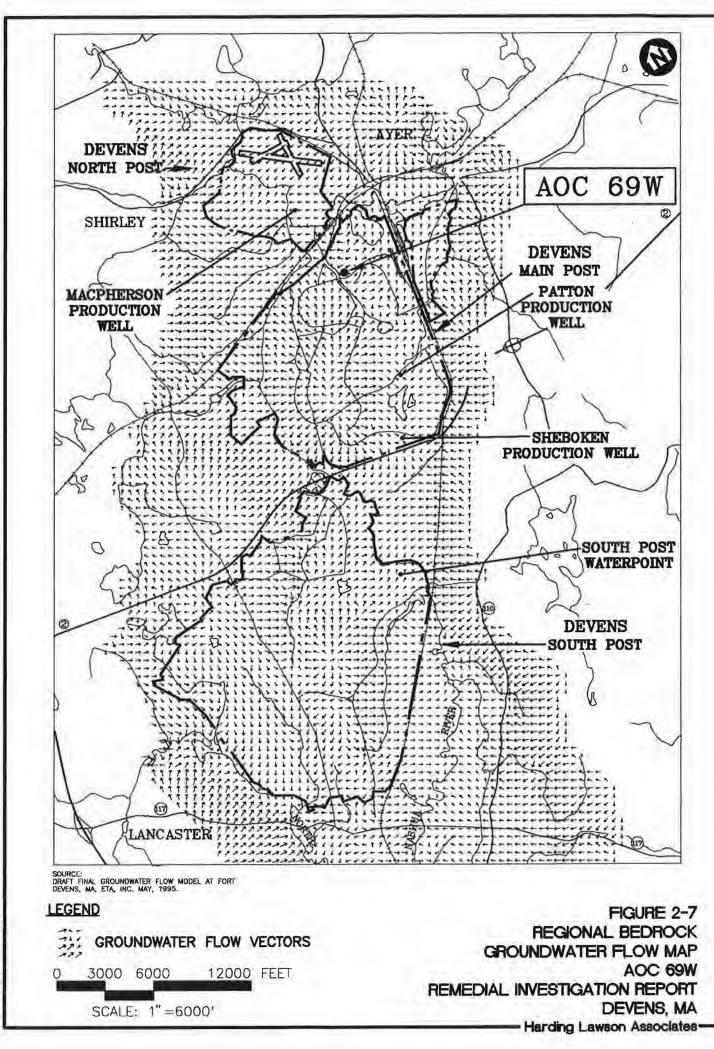




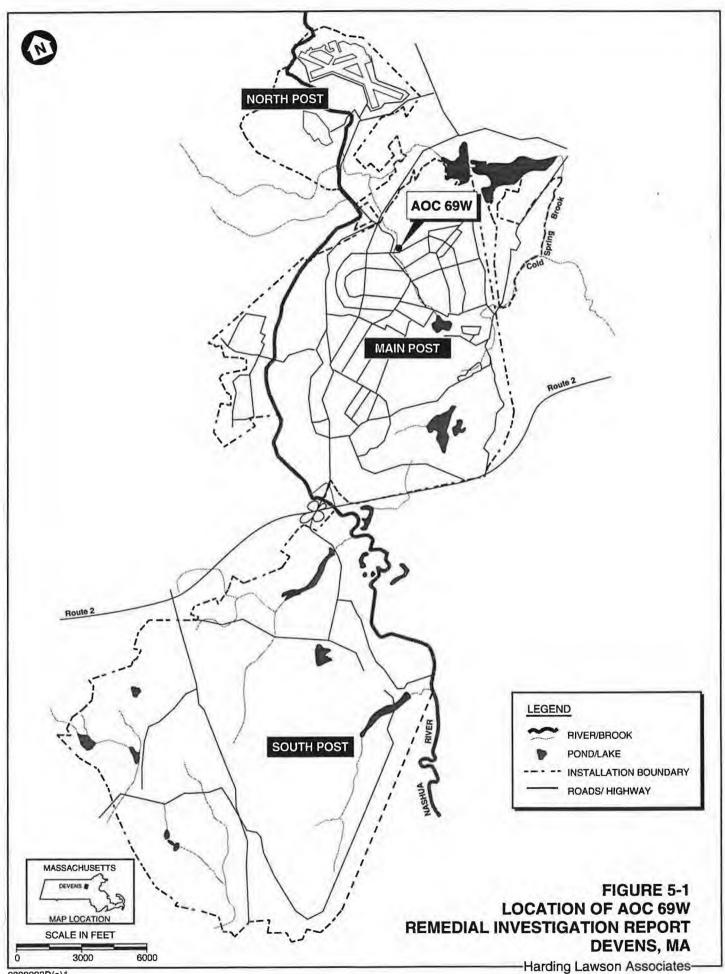


-Harding Lawson Associates-

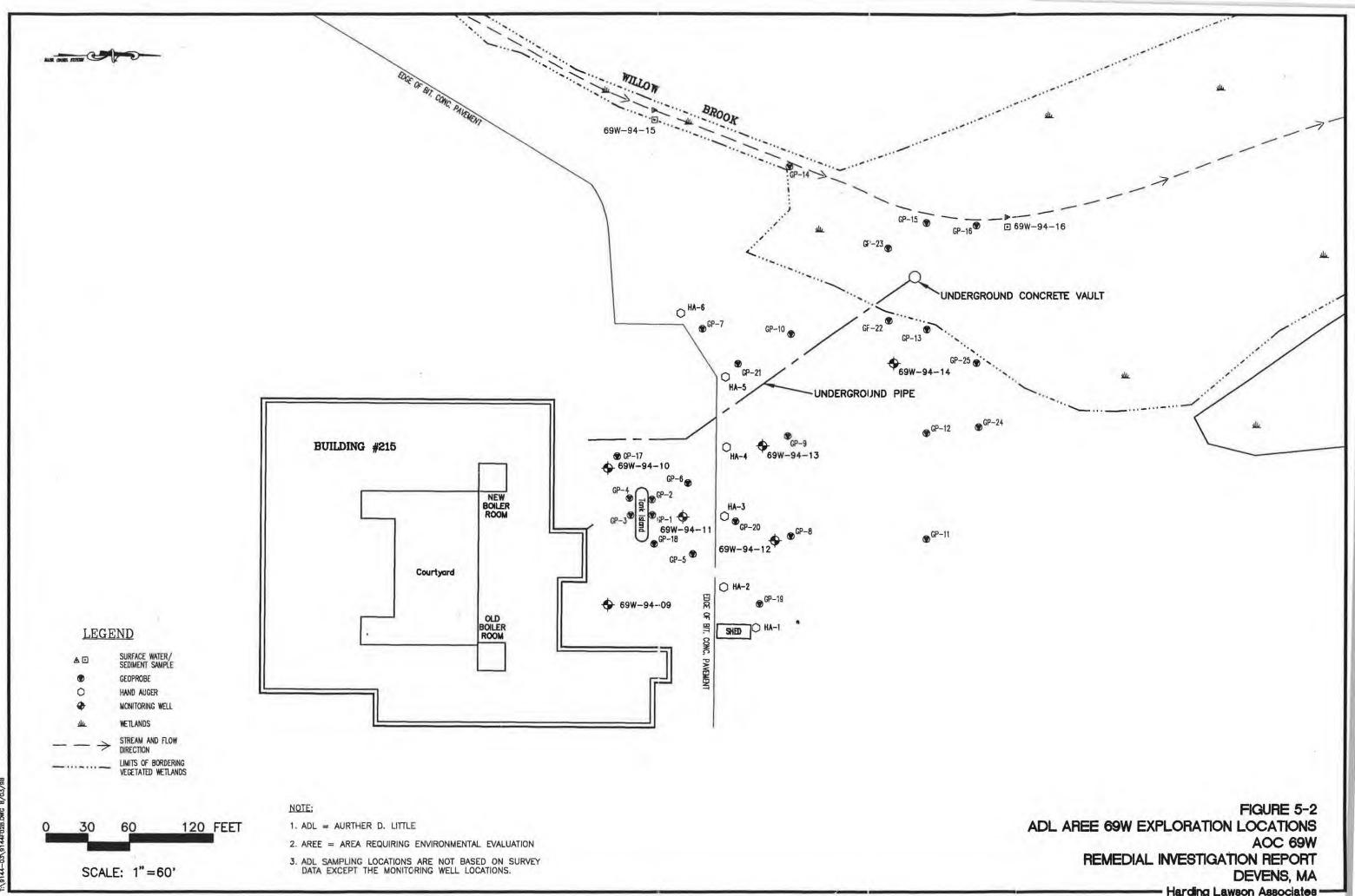
14-23 ETAFUOE DWG 8/04/



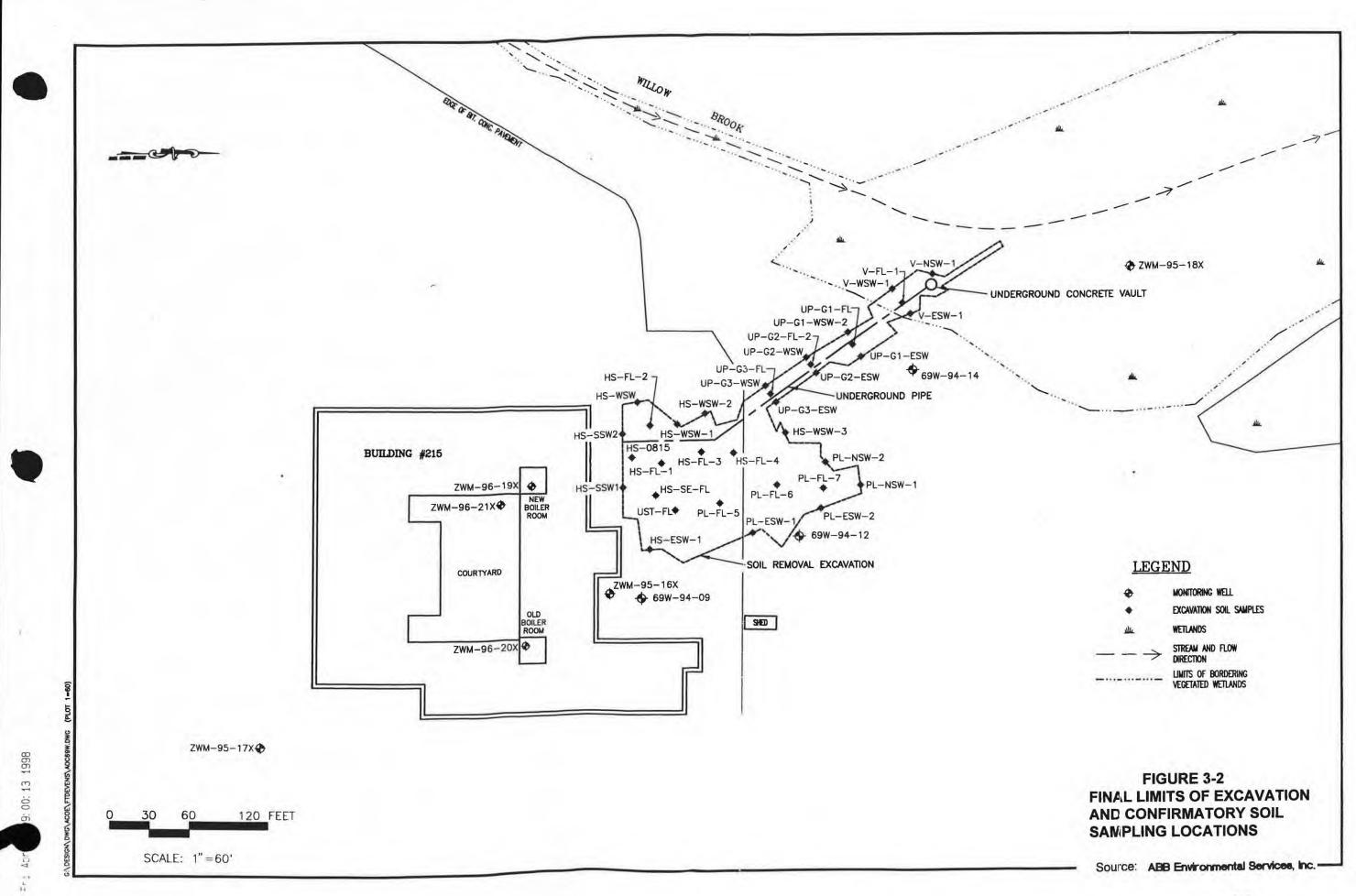
BIAFOOS DWG BY 04/88

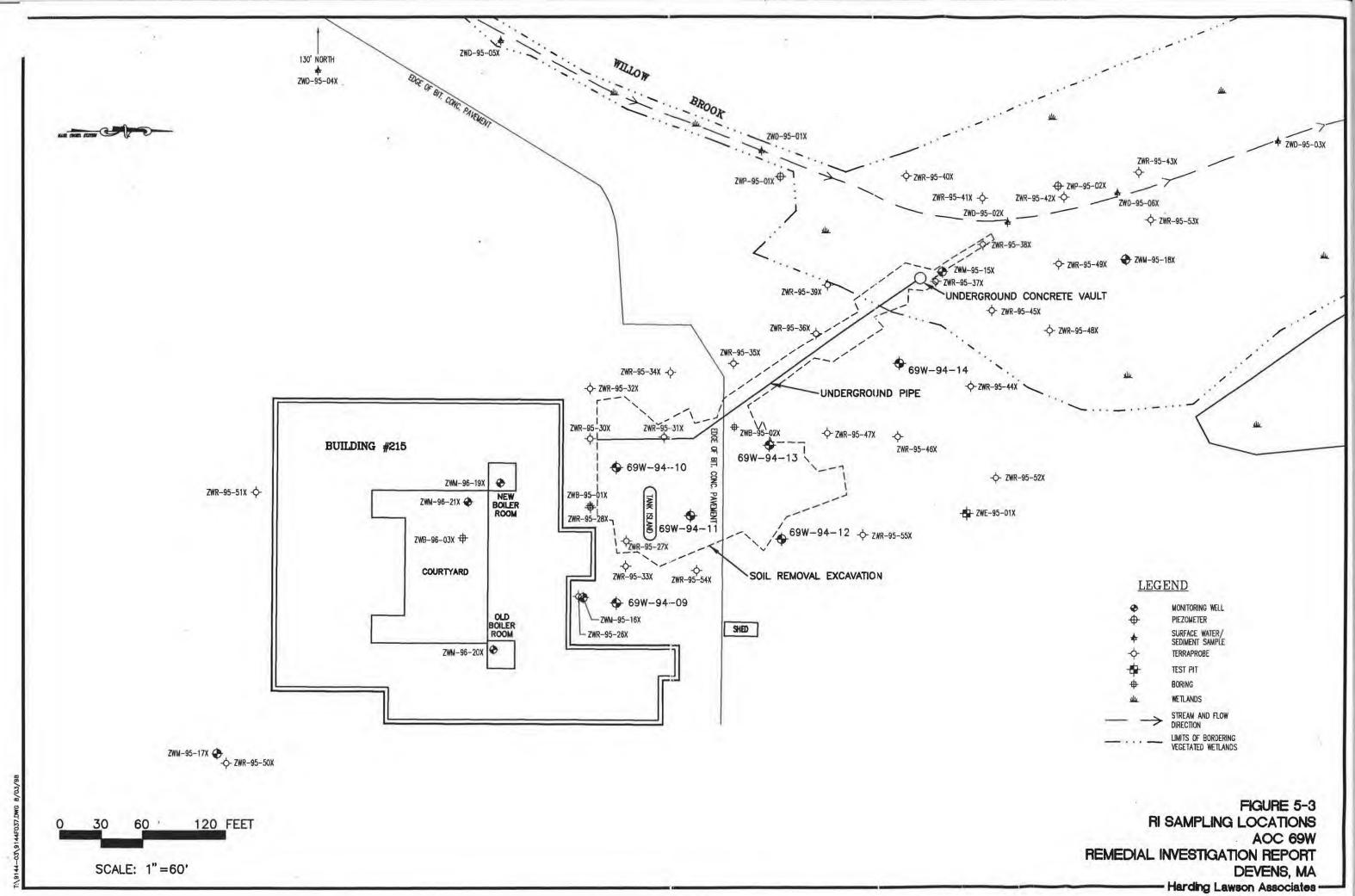


9808002D(a)1

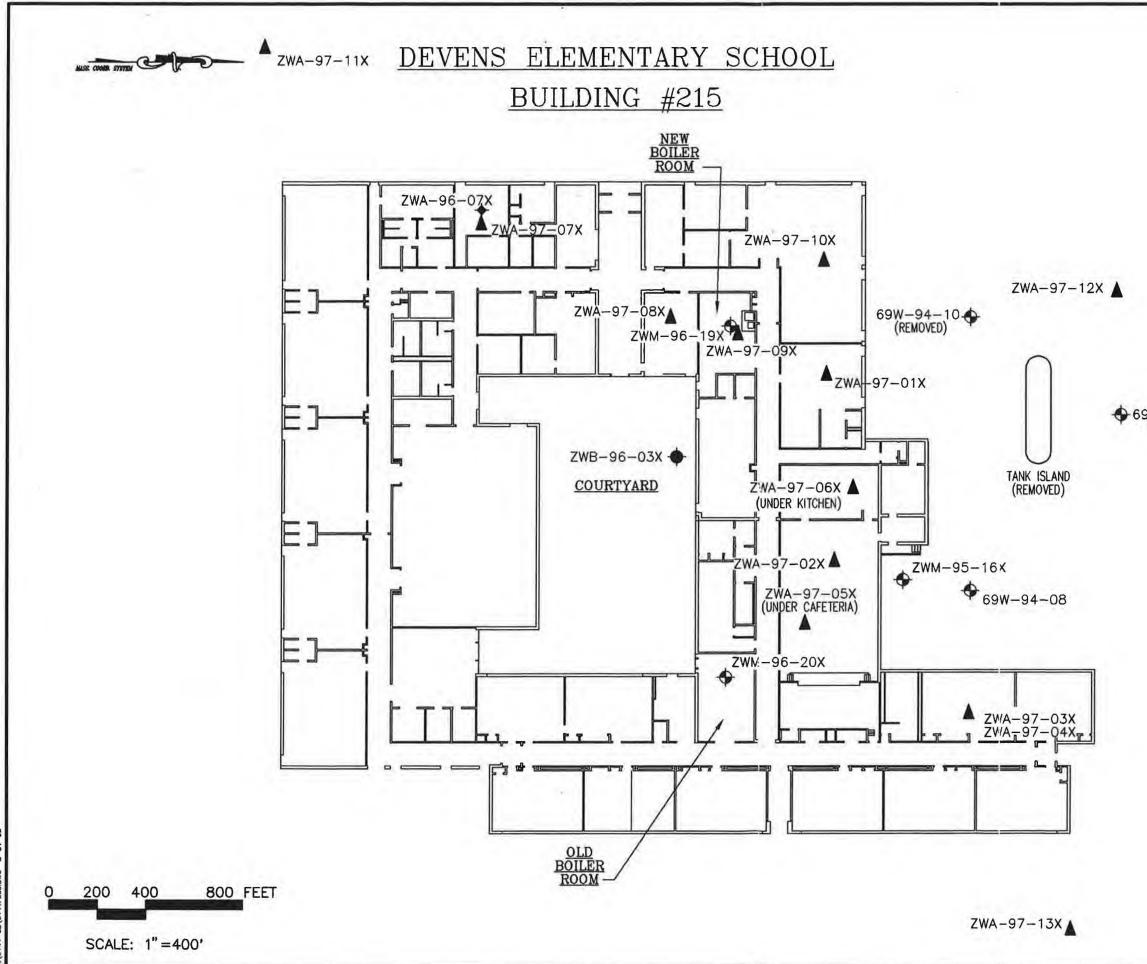


- Harding Lawson Associates





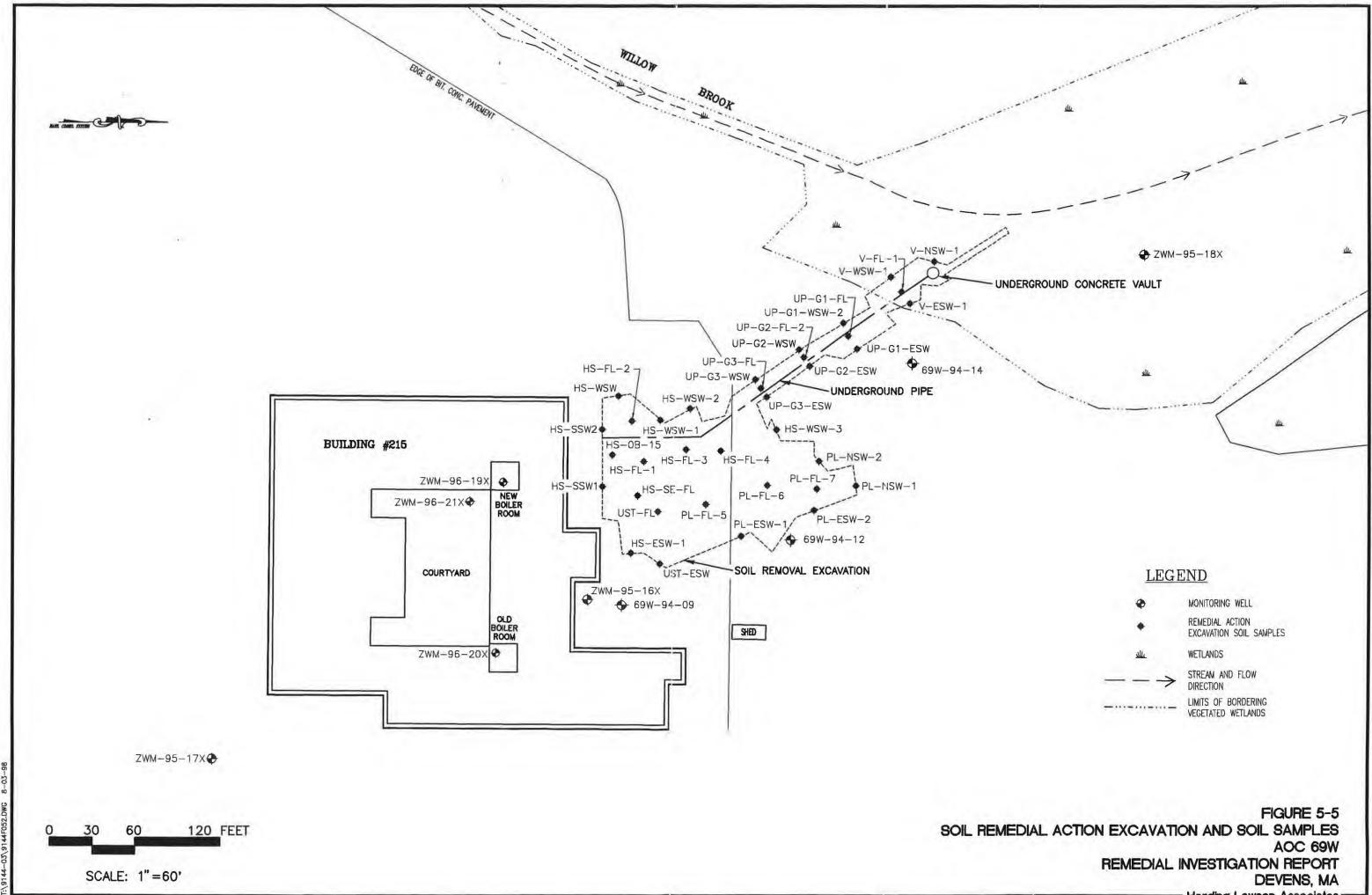
	MONITODINO BELL
Ŷ	MONITORING WELL
	PIEZOMETER
*	SURFACE WATER/ SEDIMENT SAMPLE
-0-	TERRAPROBE
+	TEST PIT
+	BORING
عللد	WETLANDS
\rightarrow	STREAM AND FLOW DIRECTION
	LIMITS OF BORDERING VEGETATED WETLANDS



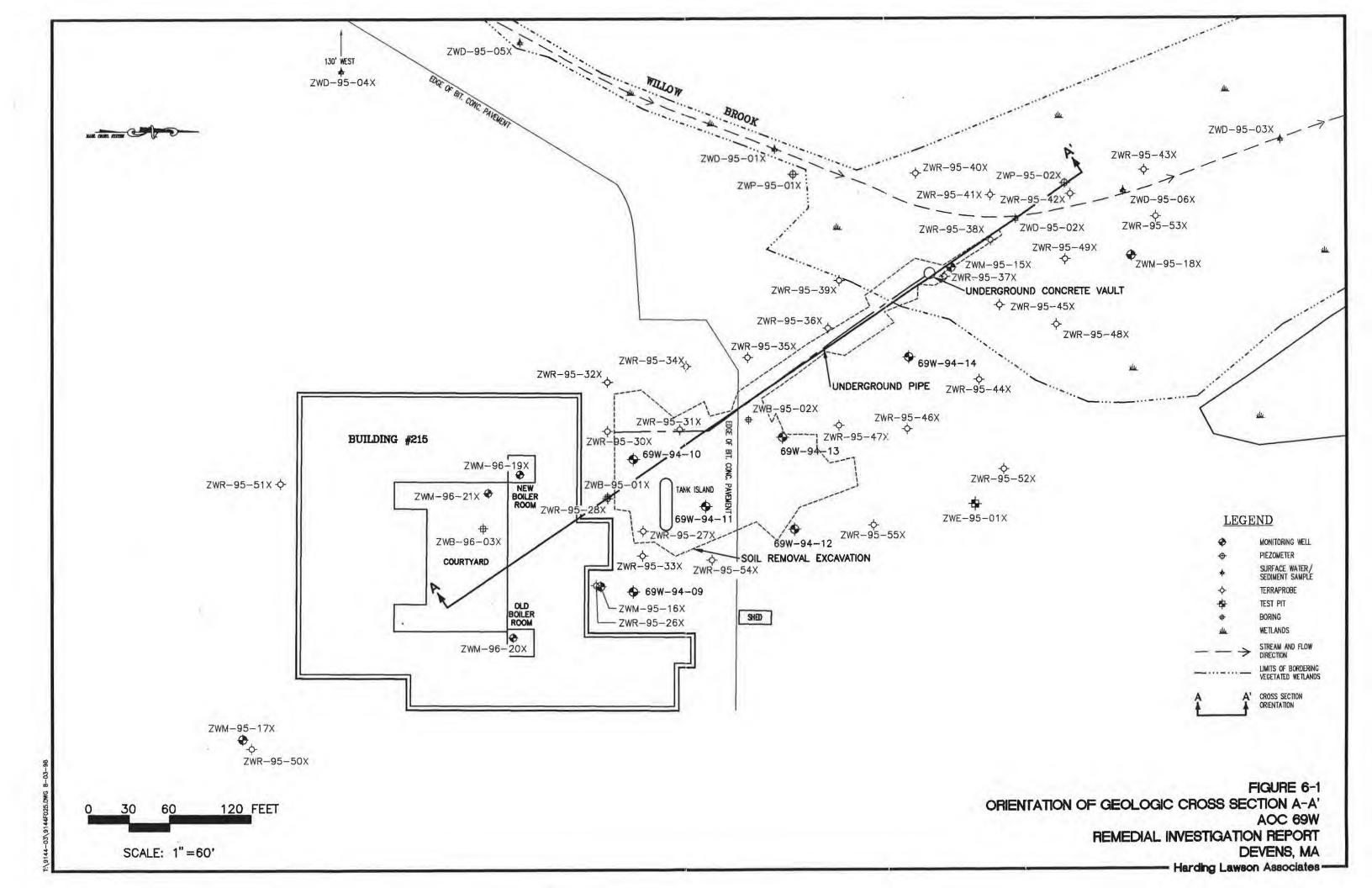
LEGEND EXISTING MONITORING WELL LOCATION EXISTING BORING LOCATION AIR SAMPLING LOCATION (1997) (REMOVED) REMOVED DURING 1997-1998 SOIL REMOVAL ACTION

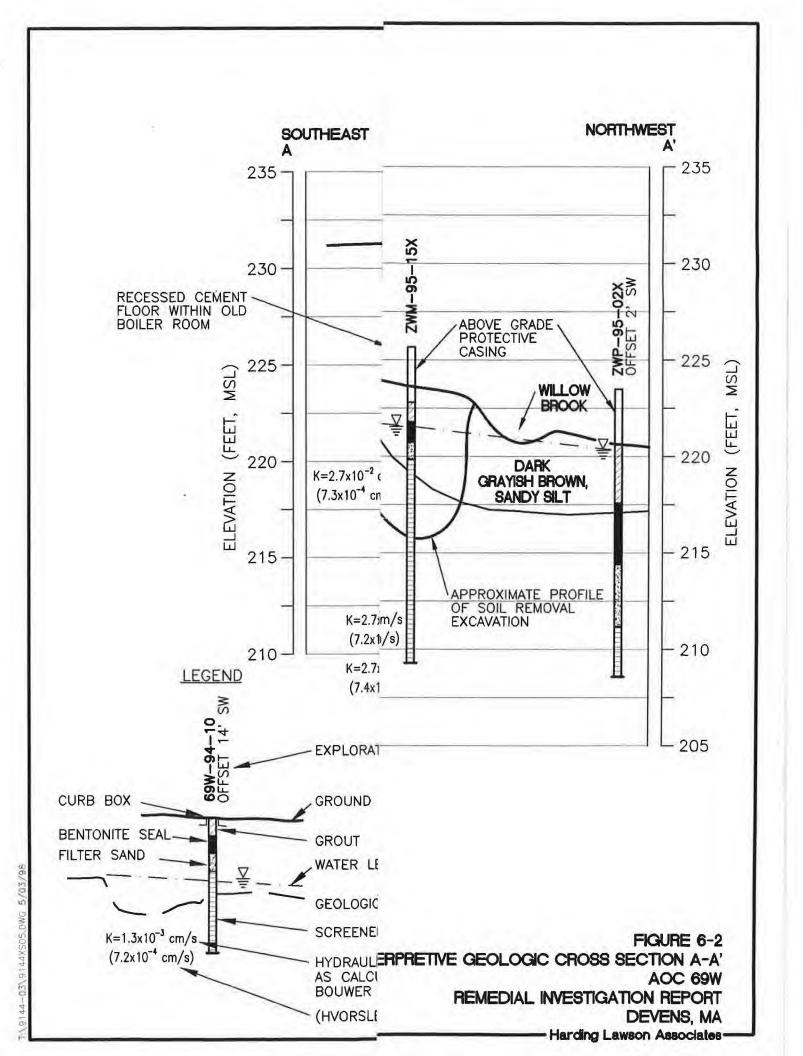
♦ 69W-94-11 (REMOVED)

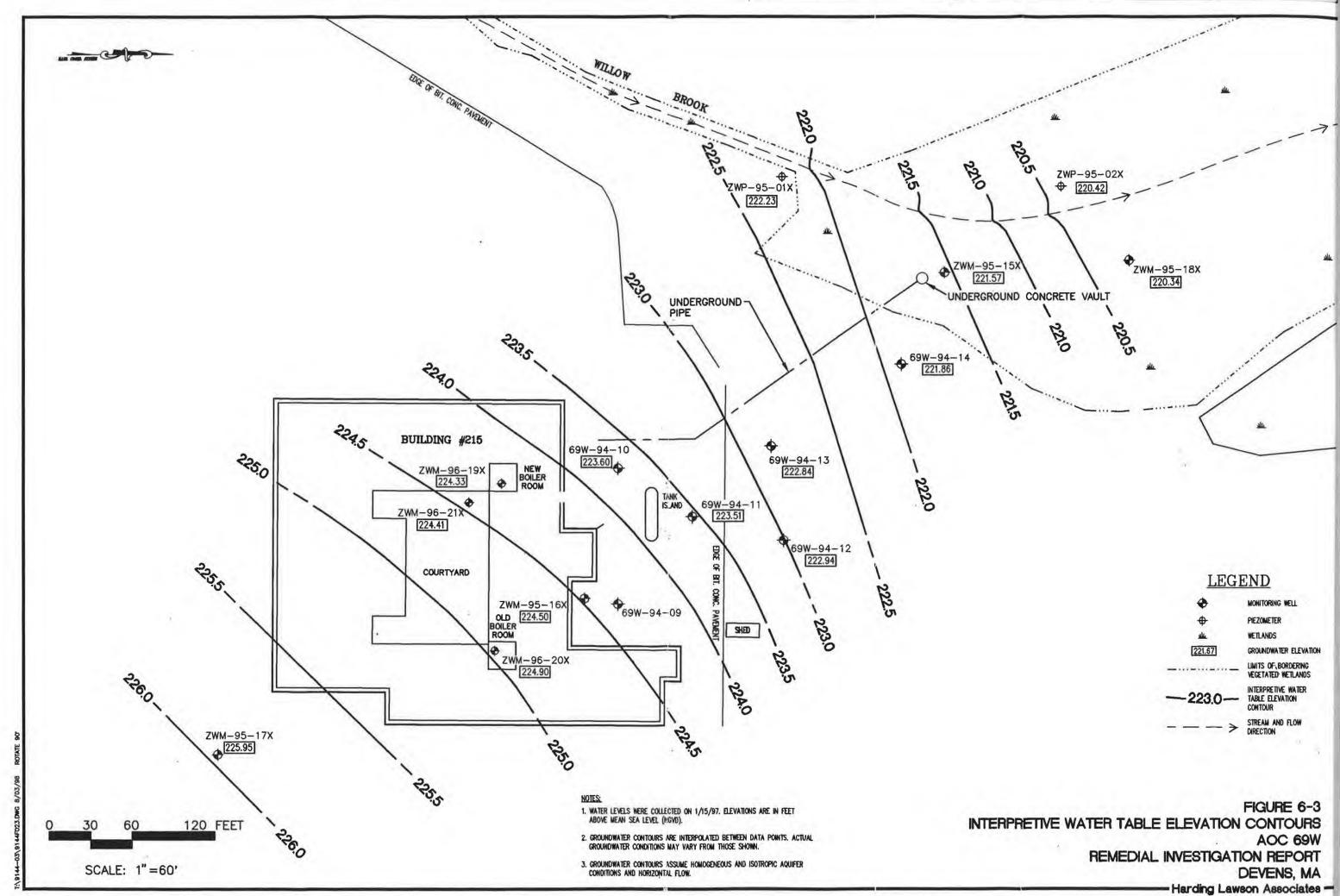
> FIGURE 5-4 AIR SAMPLING LOCATIONS AOC 69W REMEDIAL INVESTIGATION REPORT DEVENS, MA Harding Lawson Associates



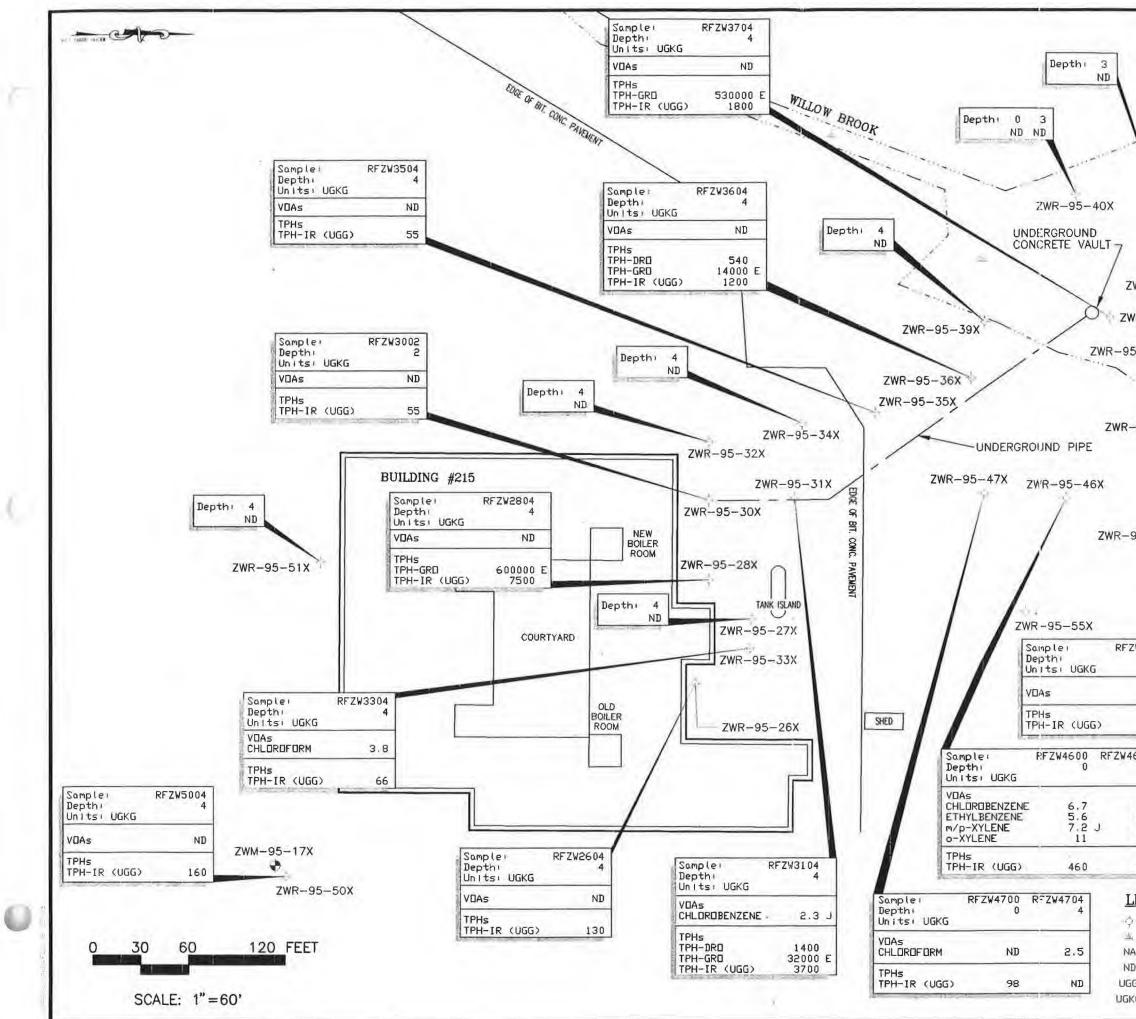
- Harding Lawson Associates -





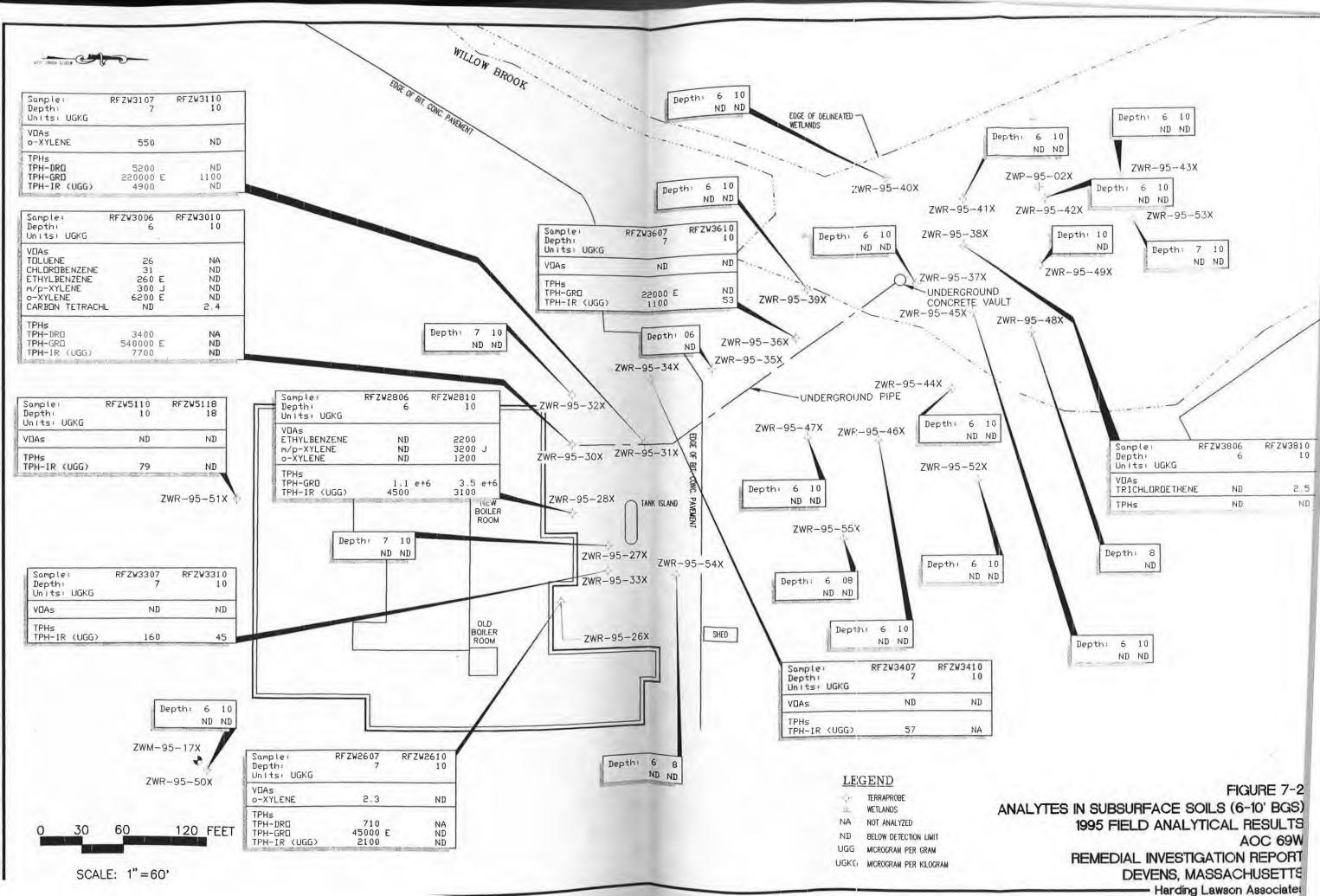


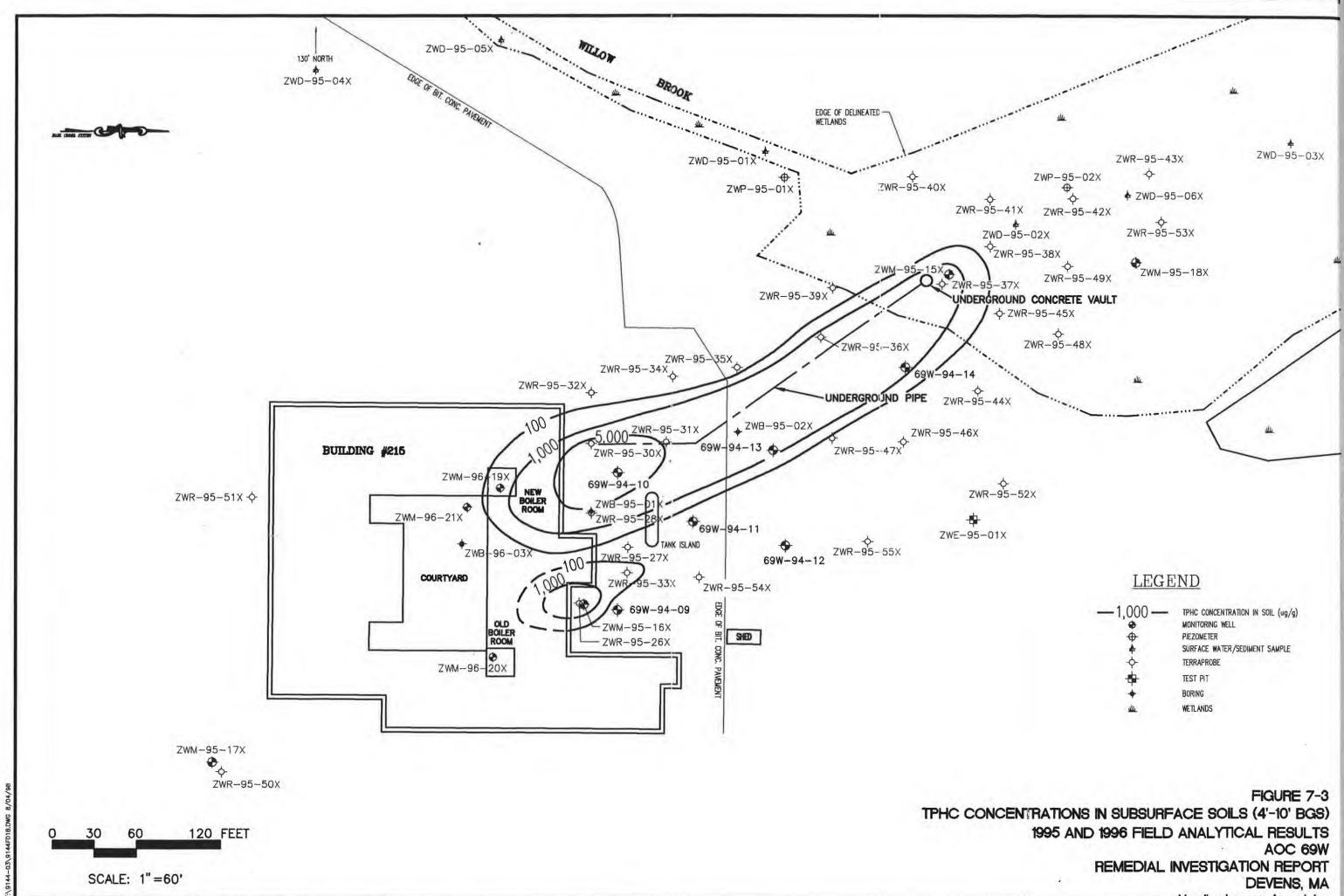
•	MONITORING WELL
	PIEZOMETER
stle	WETLANDS
221.67	GROUNDWATER ELE
	LIMITS OF, BORDER
-223.0-	INTERPRETIVE WAT TABLE ELEVATION CONTOUR
>	STREAM AND FLOW



	epthi 3			
ZWR-95-41X	ND ZWR-9	Same	olei	RFZW53
ZWR-95-42) Depthi	3	Dept Un 11	ts: UGKG	
	ZWR-95-53X			
-95-38X ZWR-	95–49X		IR (UGG)	-
95-37X	Sample: RF Depth: Units: UGKG	ZW4900 0	RFZW4902 2	RFZW49
15X ZWR-95-4		ND	ND	
1 1	TPHs TPH-GRD	ND	2500	ε
1.	TPH-IR (UGG)	140	ND	
5-44X	Sample, RF2 Depthi Units: UGKG	2W4800 1	RFZW4801	RFZW480
	VDAs	ND	ND	1
1	TPHS TPH-IR (UGG)	ND	ND	13
-52X	Sample: RFZV Depth: Units: UGKG	4500 RI 0	FZW4501 F 1	FZW450
1	VDAs TOLUENE CHLOROBENZENE ETHYLBENZENE m/p-XYLENE o-XYLENE	ND ND ND ND ND	2.4 4.1 12 13 12	
204	TPHs TPH-DRD	NA	2400	N
204 4	TPH-GRD	ND ND	83000 E 3400	1100
	TPH-IR (UGG)			
4		400 RF2 0	W4404 4	
4 ND 59	TPH-IR (UGG) Sample: RFZW4 Depth: Units: UGKG VDAs TOLUENE			
4 ND 59	TPH-IR (UGG) Sample: RFZW4 Depth: Units: UGKG VDAs TOLUENE	0	4 ND	

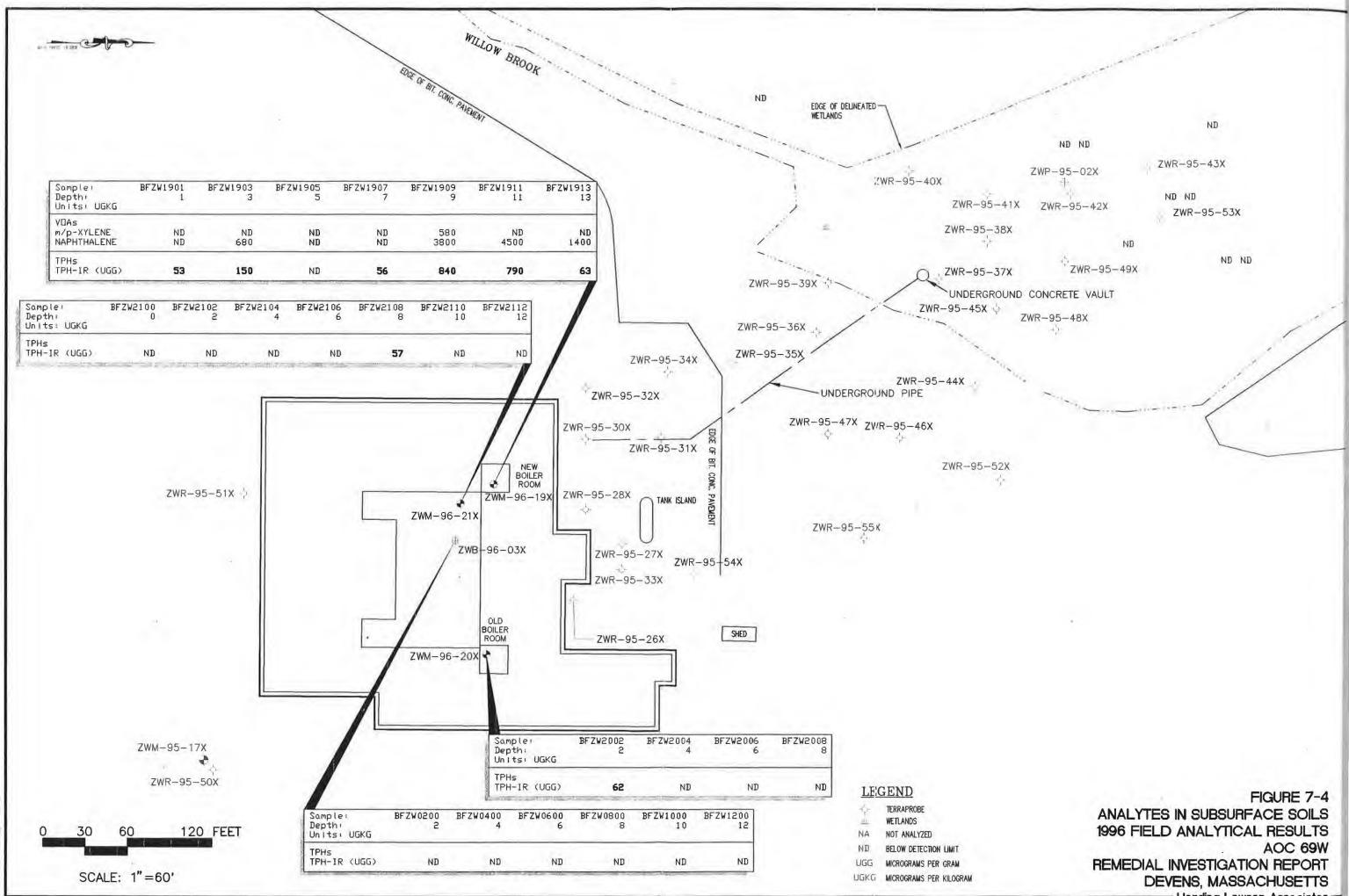
- Harding Lawson Associates



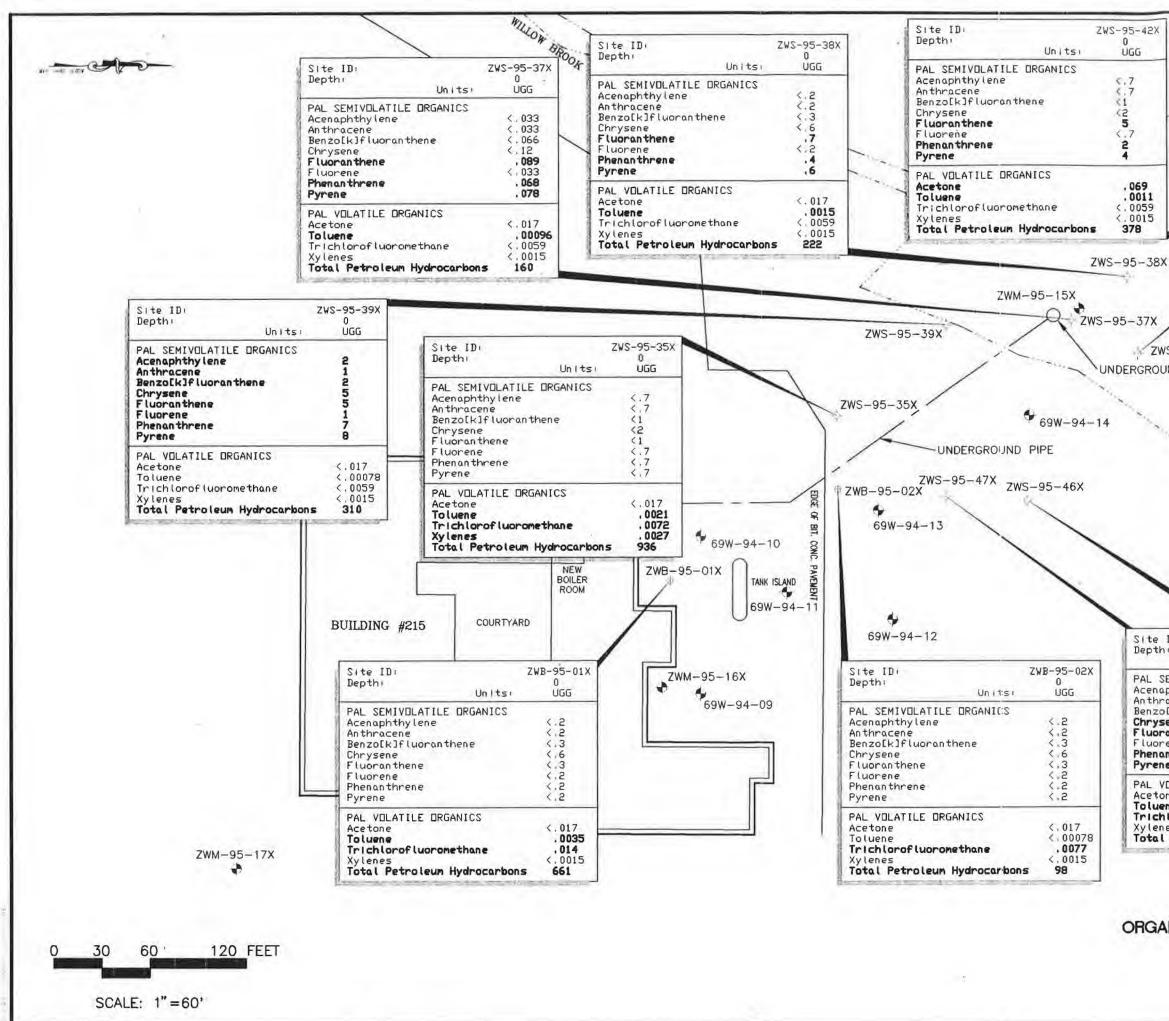


-1,000 -	TPHC CONCENTRATION IN SOIL (ug/g) MONITORING WELL	
•	PIEZOMETER	
*	SURFACE WATER/SEDIMENT SAMPLE	
\$	TERRAPROBE	
+	TEST PIT	
+	BORING	
Alle	WETLANDS	

- Harding Lawson Associates -

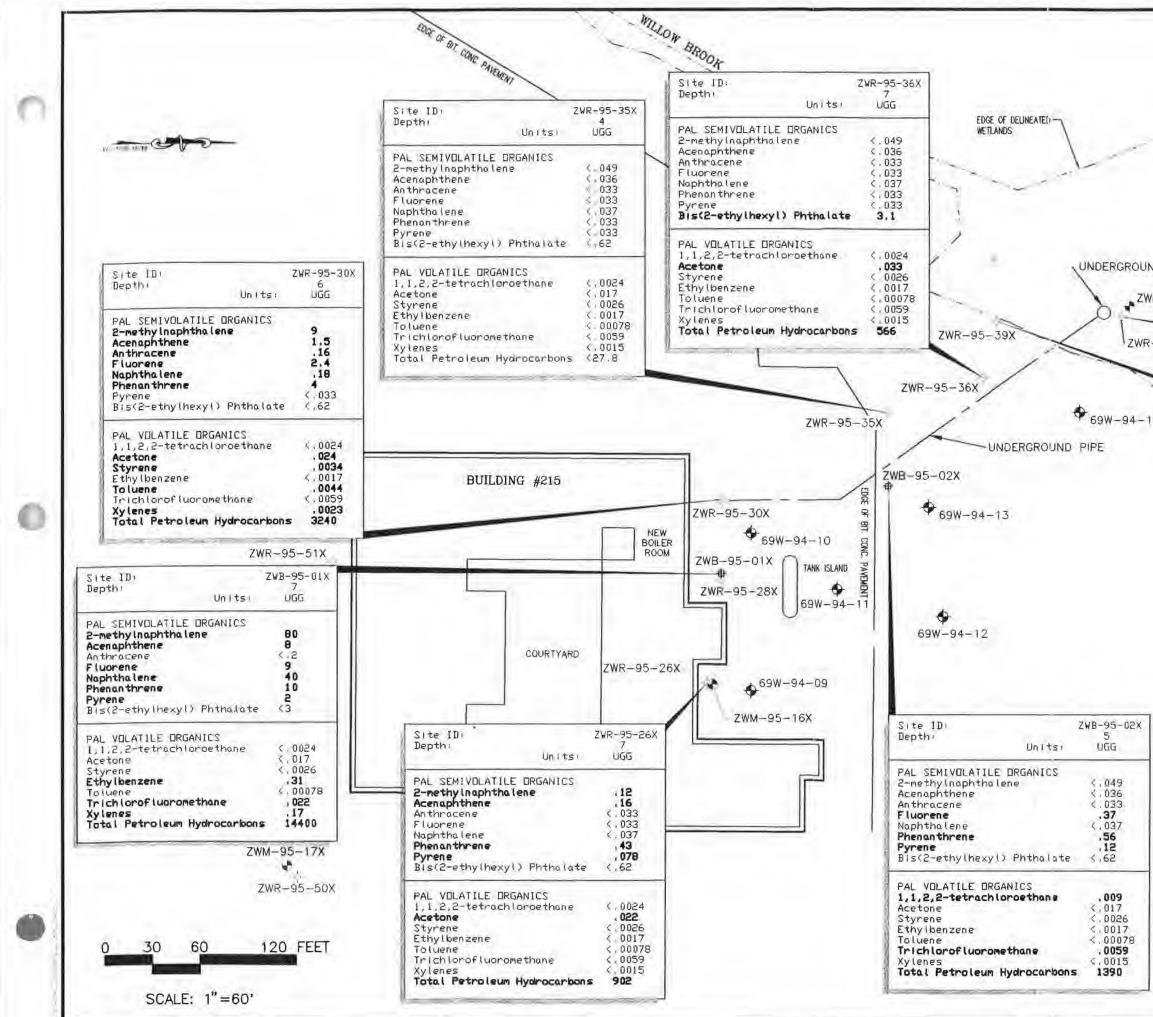


Harding Lawson Associates



ZWS-95-45X Site ID: Depth: UGG Units PAL SEMIVOLATILE ORGANICS <.033 Acenaphthylene <.033 Anthracene <.066 Benzo[k]fluoranthene <.12 Chrysene Fluoranthene <.033 Fluorene .065 Phenanthrene .075 Pyrene PAL VOLATILE DRGANICS <.017 Acetone <.00078 Toluene <.0059 Trichlorofluoromethane <.0015 Xylenes ZWS-95-42X <27.5 Total Petroleum Hydrocarbons ZWS-95-45X UNDERGROUND CONCRETE VAULT ZWS-95-46X Site ID: Depth Unitsi UGG PAL SEMIVOLATILE ORGANICS <3 Acenaphthylene <3 <7 <10 Anthracene Benzo[k]fluoranthene Chrysene Fluoranthene 9 <3 luorene 9 Phenanthrene 10 Pyrene PAL VOLATILE DRGANICS C.017 Acetone Toluene < 00078 Trichlorofluoromethane <.0059 .0015 Xylenes Total Petroleum Hydrocarbons 652 Site ID: ZWS-95-47X Depthi 0 Unitsi UGG PAL SEMIVOLATILE ORGANICS <.033 Acenaphthy lene <.033 <.066 Anthracene Benzo[k]fluoranthene .17 Chrysene LEGEND Fluoranthene .19 Fluorene <.033 MONITORING WELL ÷. Phenanthrene .14 PIEZOMETER .22 Pyrene TERRAPROBE PAL VOLATILE DRGANICS <.017 BORING 111 Acetone .001 Toluene WETLANDS Trichlorofluoromethane .0055 UGG MICROGRAMS PER GRAM <.0015 Total Petroleum Hydrocarbons 52.5 FIGURE 7-5 ORGANIC ANALYTES IN SURFACE SOILS (0-2' BGS) 1995 OFF-SITE ANALYTICAL RESULTS

AOC 69W REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS Harding Lawson Associates



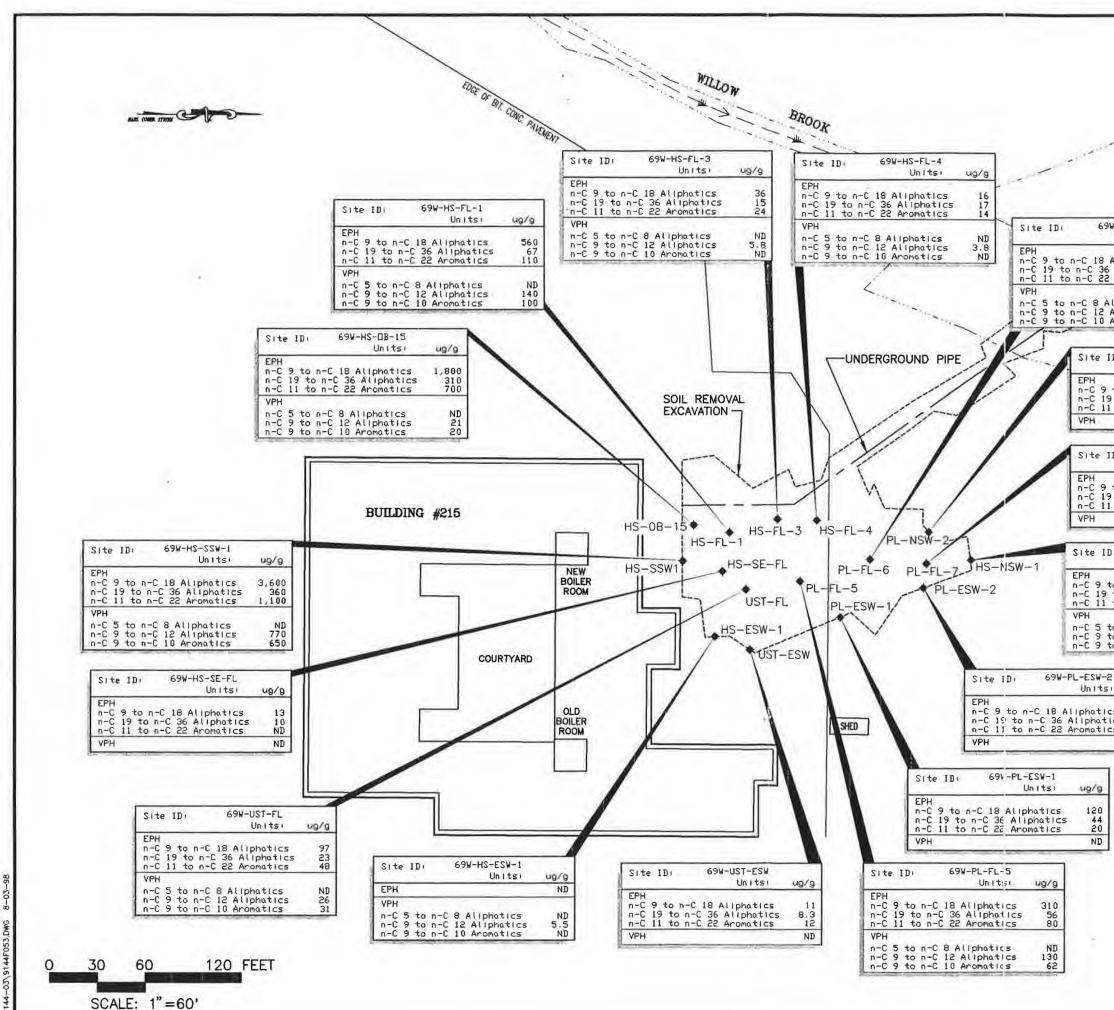
	Site ID: Depth:	Unitsi	ZWR-95-37X 4 UGG
	PAL SEMIVOLATILE DRO 2-methylnaphthalene Acenaphthene Anthracene Fluorene Naphthalene Phenanthrene Pyrene Bis(2-ethylhexyl) Ph		<.049 .2 .053 .53 .037 .89 .096 <.62
CONCRETE VAULT -95-15X 95-37X	PAL VOLATILE ORGANIC 1,1,2,2-tetrachloros Acetone Styrene Ethylbenzene Toluene Trichlorofluorometha Xylenes Total Petroleum Hydr	thane ine	<pre>\$.0024 .03 \$.0026 \$.0017 .0024 \$.0059 \$.0015 \$.0015</pre>
Site I Depthi		WR-95-39 4 UGG	×
2-meth	ene	<.049 <.036 <.033 <.033 <.033	
Naphth Phenar Pyrene	threne	<.033 <.033 <.62	

PAL VOLATILE DRGANICS 1,1,2,2-tetrachloroethane <.0024 Acetone <.017 Styrene <.0026 Ethylbenzene <.0017 Toluene <.00078 Trichlorofluoromethane <.0059 Xylenes <.0015 Total Petroleum Hydrocarbons <27.5

LEGEND

2	MONITORING WELL
÷.	PIEZOMETER
÷.	TERRAPROBE
-ip-	BORING
100	WETLANDS
UGG	MICROGRAMS PER GRA

TABLE 7-6 ORGANIC ANALYTES IN SUBSURFACE SOILS 1995 OFF-SITE ANALYTICAL RESULTS AOC 69W REMEDIAL INVESTIGATION REPORT DEVENS, MASSACHUSETTS ABB Environmental Services, Inc.*



69W-PL-FL-6		
Unitsi	ug/9	-
n-C 18 Aliphatics	310	
n-C 36 Aliphatics n-C 22 Aromatics	36 120	
n-C 22 Aromatics	120	
-C 8 Aliphatics	ND	
n-C 12 Aliphatics n-C 10 Aromatics	52	
n-C IU Aromatics	33	
Site ID: 69W-	PL-NSW-2	5.0
	Units	ug/9
EPH	linhadir-	2.2
n-C 9 to n-C 18 A n-C 19 to n-C 36	Aliphatics	3.3
n-C 11 to n-C 22	Aromatics	9
VPH		ND
Site ID: 69W	-PL-FL-7	-
20,00,020	Units	ug/9
EPH	And An Arts of	
n-C 9 to n-C 18 A	Allohatics	370
n-C 19 to n-C 36 n-C 11 to n-C 22	Aromatics	23
VPH		ND
	al the second	
Site ID: 69W-H	IS-NSW-1	
0,100,10.	Unitsi	ug/g
EPH	a de la seco	
n-C 9 to n-C 18 Al n-C 19 to n-C 36 A	Iphatics	410 68
1-C 19 10 1-C 36 A	inpractics	120
n-C 11 to n-C 22 A	romatics	
n-C 11 to n-C 22 A	romatics	
VPH		ND
VPH n-C 5 to n-C 8 All n-C 9 to n-C 12 Al		ND 7.4 ND

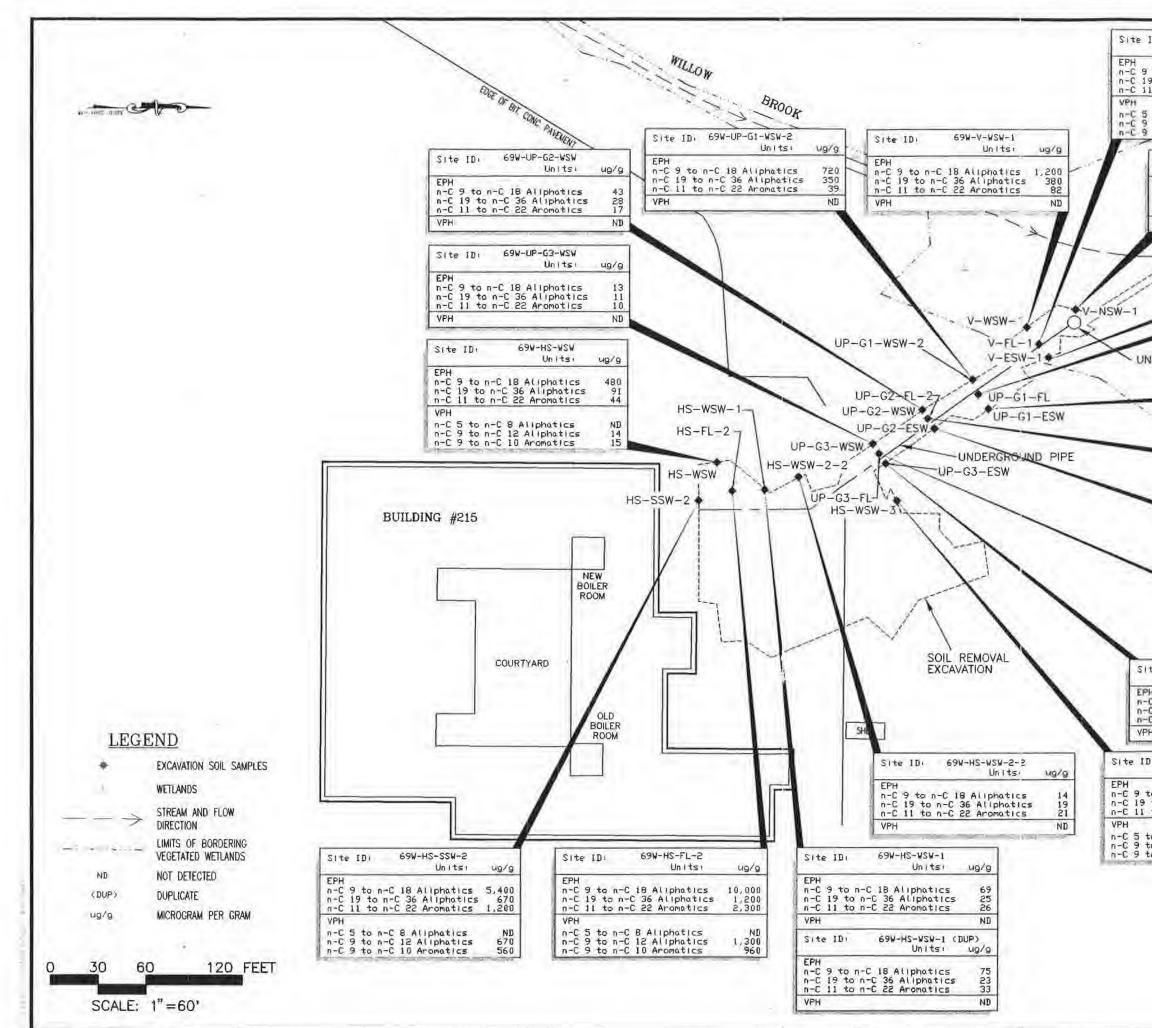
sle

2	
	ug/9
:5	5,6
CS	
s	ND
	ND
_	



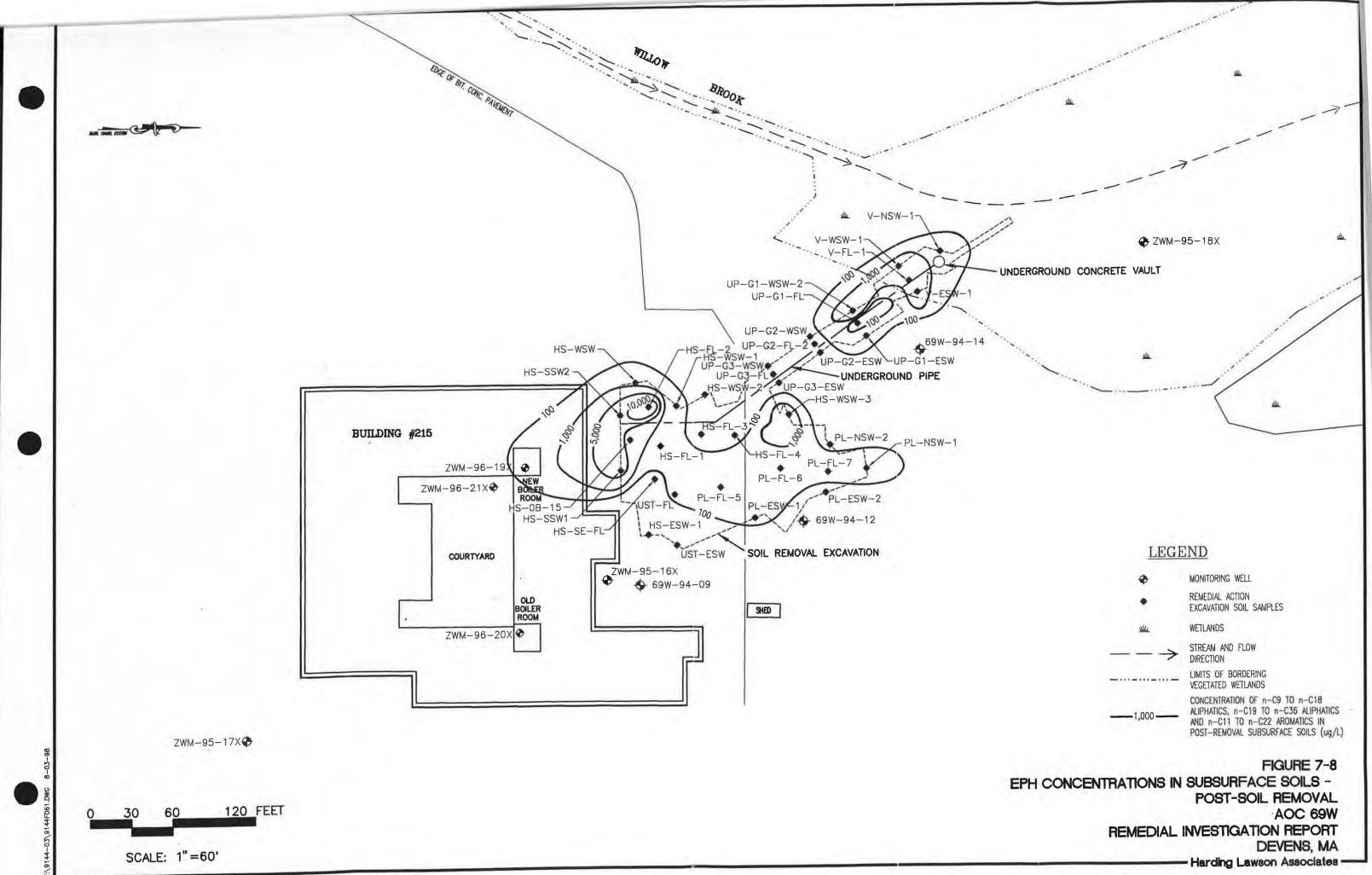
+	EXCAVATION SOIL SAMPLES
Alle	WETLANDS
$ \rightarrow$	STREAM AND FLOW DIRECTION
	LIMITS OF BORDERING VEGETATED WETLANDS
ND	NOT DETECTED
(DUP)	DUPLICATE
ug/g	MICROGRAMS PER GRAM

FIGURE 7-7 (EAST) EPH-VPH CONCENTRATIONS IN SOIL 1997 OFF-SITE ANALYTICAL RESULTS AOC 69W REMEDIAL INVESTIGATION REPORT DEVENS, MA

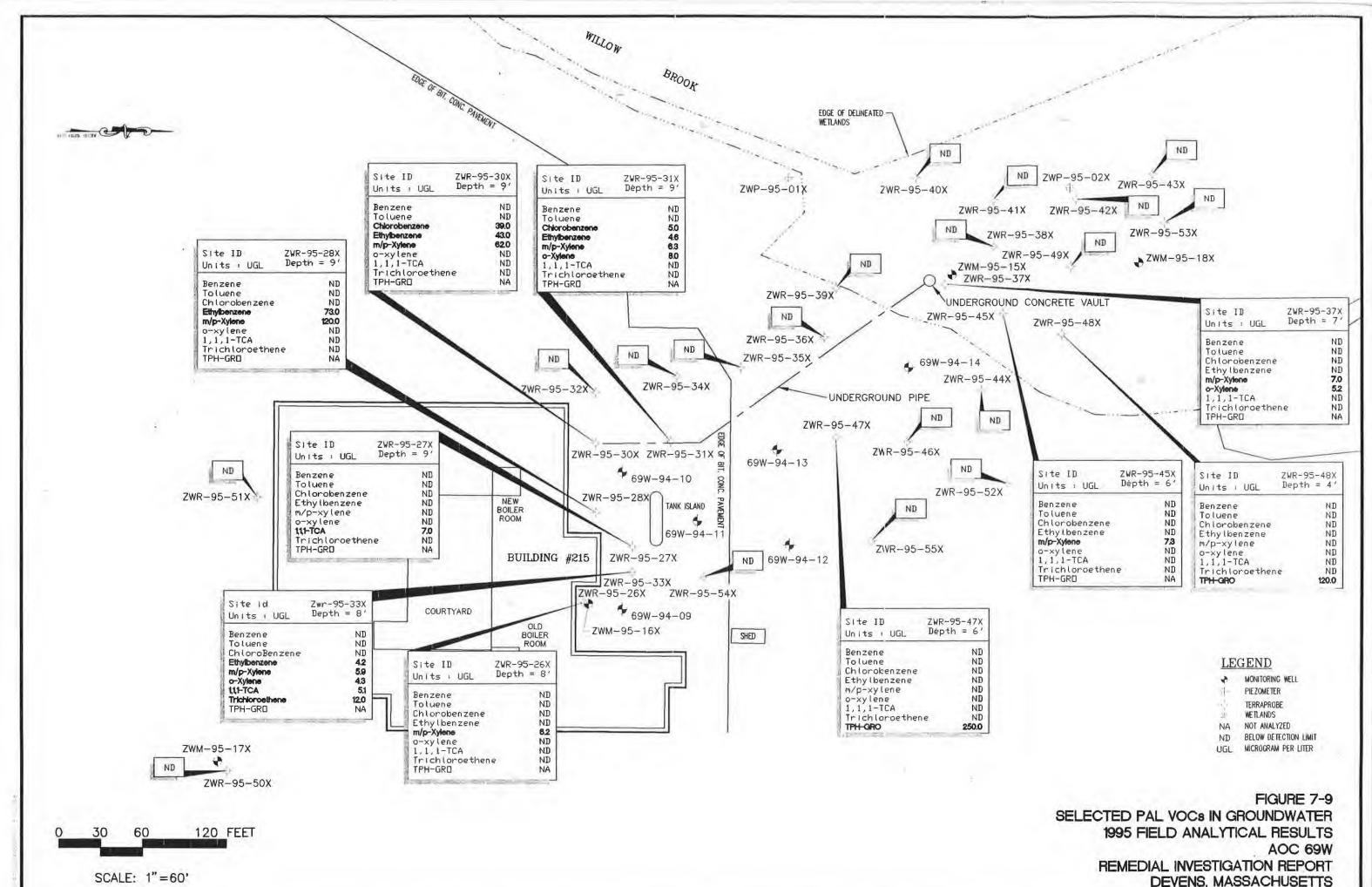


Site ID: 69W-V-FL-1	and the second s
Units: ug/g CPH C 9 to n-C 18 Aliphatics 600	
n-C 19 to n-C 36 Aliphatics 130 n-C 11 to n-C 22 Aromatics 77	
/PH -C 5 to n-C 8 Aliphatics ND -C 9 to n-C 12 Aliphotics 43 -C 9 to n-C 10 Aromatics 37	
	Site ID: 69W-V-ESW-1 Units: 40/0
Site ID: 69W-V-NSW-1 Units: ug/g	EPH
EPH n-C 9 to n-C 18 Aliphatics 74 n-C 19 to n-C 36 Aliphatics 120 n-C 11 to n-C 22 Aromatics 22	n-C 9 to n-C 18 Aliphatics 1,700 n-C 19 to n-C 36 Aliphatics 480 n-C 11 to n-C 22 Aromatics 190 VPH
VPH ND	n=C 5 to n=C 8 Aliphatics ND n=C 9 to n=C 12 Aliphatics 36 n=C 9 to n=C 10 Aromatics 35
1	Site ID: 69W-UP-GI-FL Units: ug/g
	EPH n-C 9 to n-C 18 Allphatics 14 n-C 19 to n-C 36 Allphatics 10 n-C 11 to n-C 22 Aromatics 12
UNDERGROUND CONCRETE VAULT	VPH n=E 5 to n=E 8 Allphatics NE n=E 9 to n=E 12 Allphatics 43 n=E 9 to n=E 10 Aromatics 24
	Site ID: 69W-UP-GI-ESW Units: ug/g
·	EPH n-C 9 to n-C 18 Aliphatics 120 n-C 19 to n-C 36 Aliphatics 68 n-C 11 to n-C 22 Aromatics 28
	ND
	Site ID: 69W-UP-G2-FL-2 Units: ug/c
	EPH n=C 9 to n=C 18 Aliphatics 4.3 n=C 19 to n=C 36 Aliphatics 6.1
	NI VPH
	Site ID: 69W-UP-62-ESW Units: ug/g
	EPH n-C 9 to n-C 18 Aliphotics 3.4 n-C 19 to n-C 36 Aliphotics ND n-C 11 to n-C 22 Aromatics ND
	VPH ND
Site 1D: 69W-UP-G3-ESV Units: ug/g	Site ID: 69V-UP-G3-FL Units: ug/g
EPH n-C 9 to n-C 18 Aliphatics ND n-C 19 to n-C 36 Aliphatics 5.4 n-C 11 to n-C 22 Aromatics ND	EPH n-C 9 to n-C 18 Aliphatics 10 n-C 19 to n-C 36 Aliphatics 19 n-C 11 to n-C 28 Aromatics 27
VPH ND	VPH ND
te ID: 69W-HS-WSW-3 Units: ug/g	
н	
C 9 to n-C 18 Allphatics 860 C 19 to n-C 36 Allphatics 130 C 11 to n-C 22 Aromatics 220	

FIGURE 7-7 (WEST) EPH-VPH CONCENTRATIONS IN SOIL 1997 OFF-SITE ANALYTICAL RESULTS AOC 69W REMEDIAL INVESTIGATION REPORT DEVENS, MA

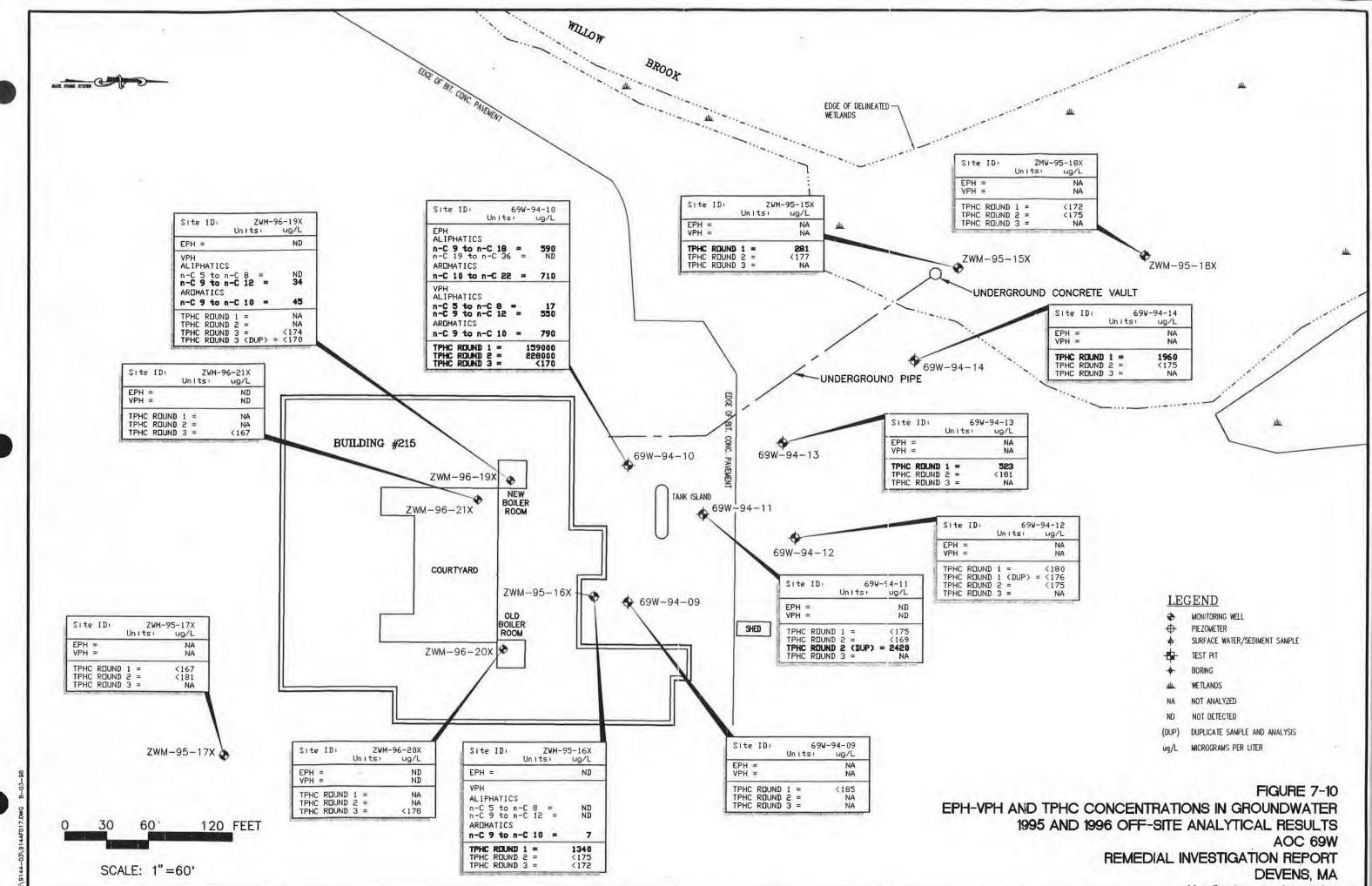


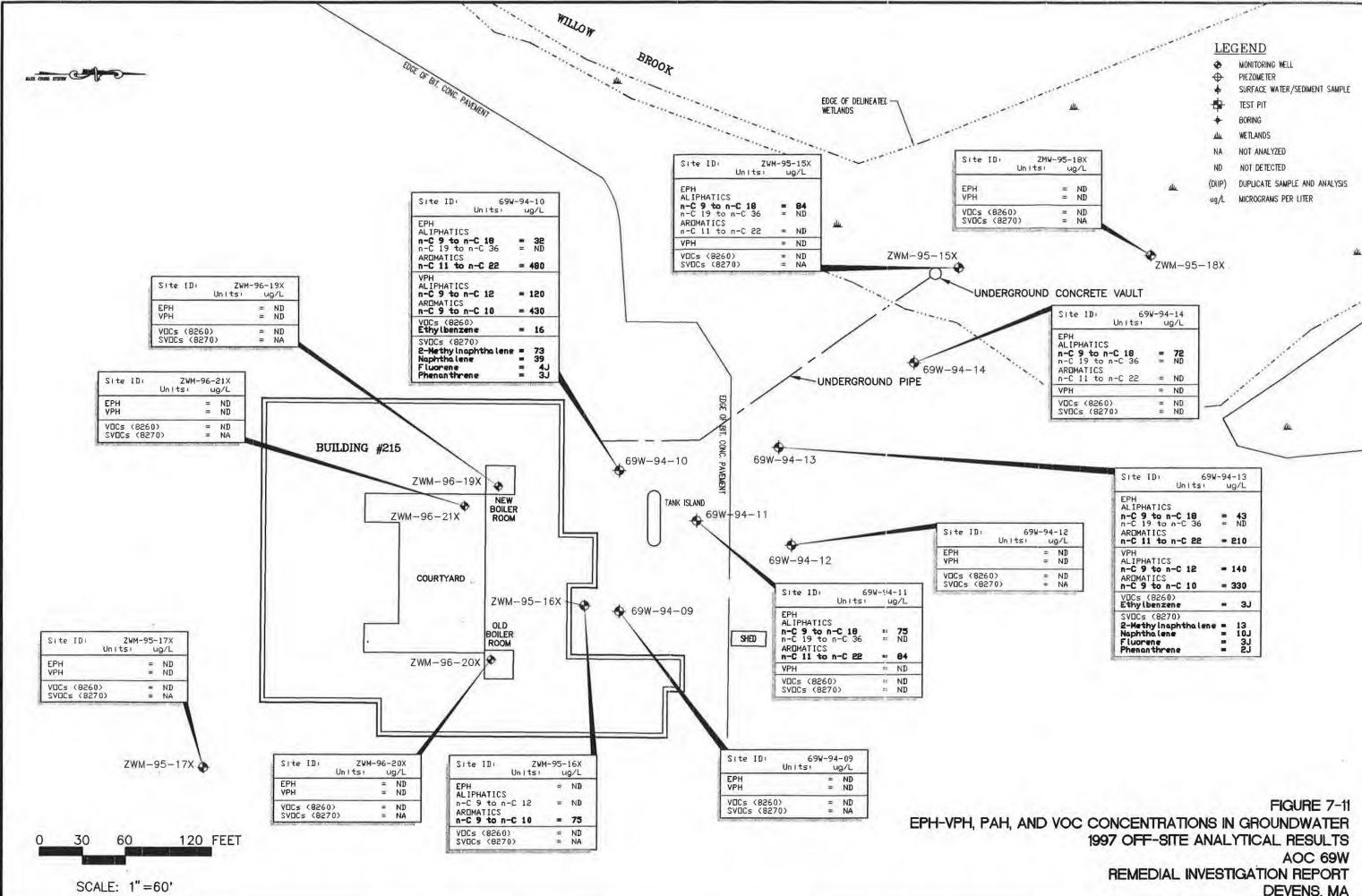
۲	MONITORING WELL
. +	REMEDIAL ACTION EXCAVATION SOIL SAMPLES
يملك	WETLANDS
$ \rightarrow$	STREAM AND FLOW DIRECTION
	LIMITS OF BORDERING VEGETATED WETLANDS



*	MONITORING WELL
do	PIEZOMETER
	TERRAPROBE WETLANDS
NA	NOT ANALYZED
ND	BELOW DETECTION LIMIT
UGL	MICROGRAM PER LITER

DEVENS, MASSACHUSETTS - Harding Lawson Associates -

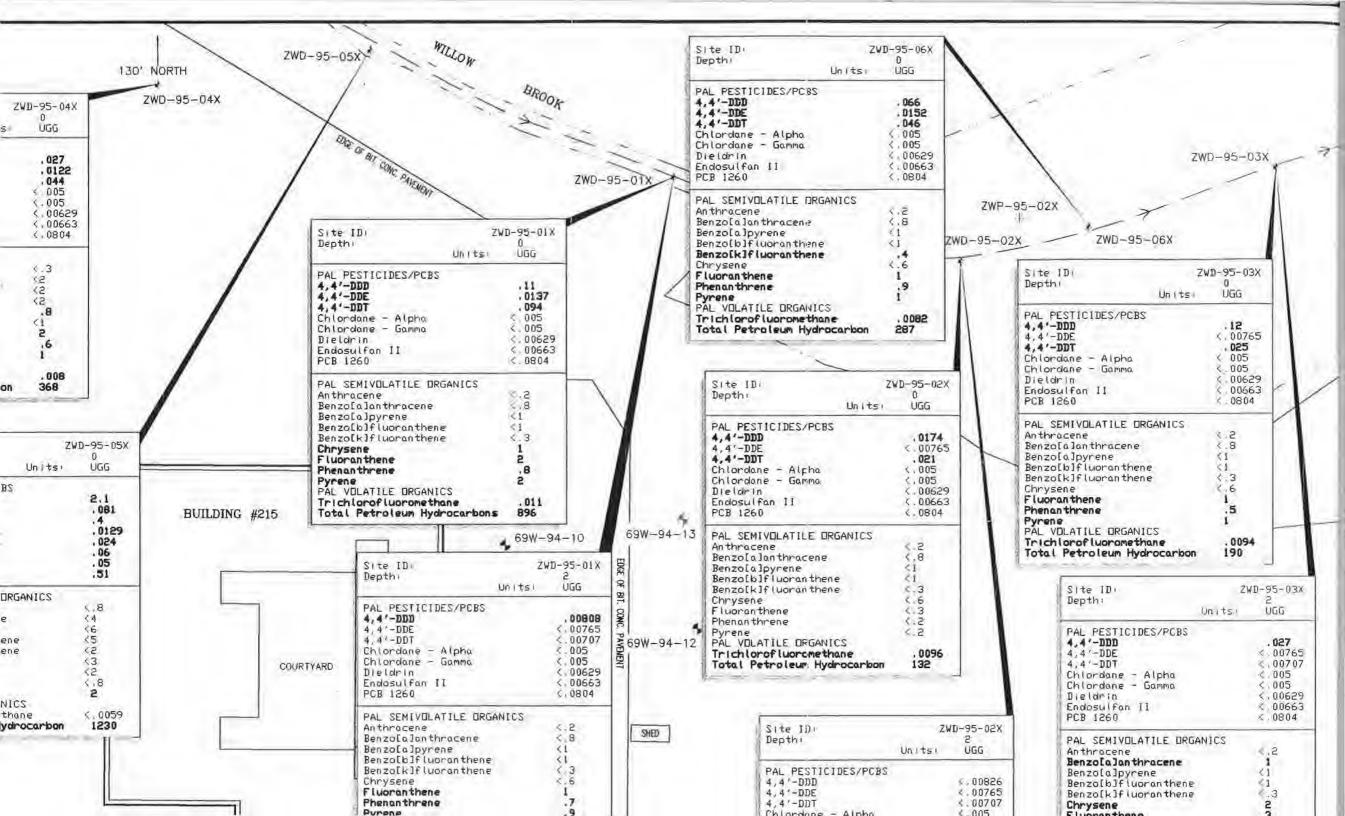


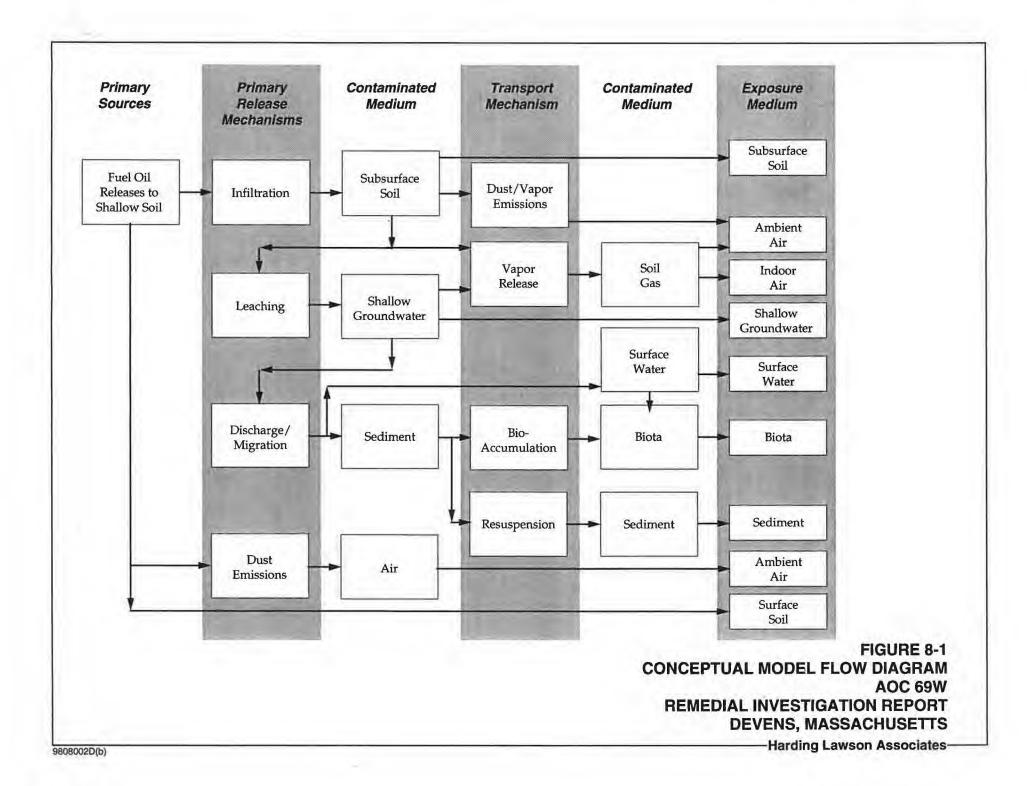


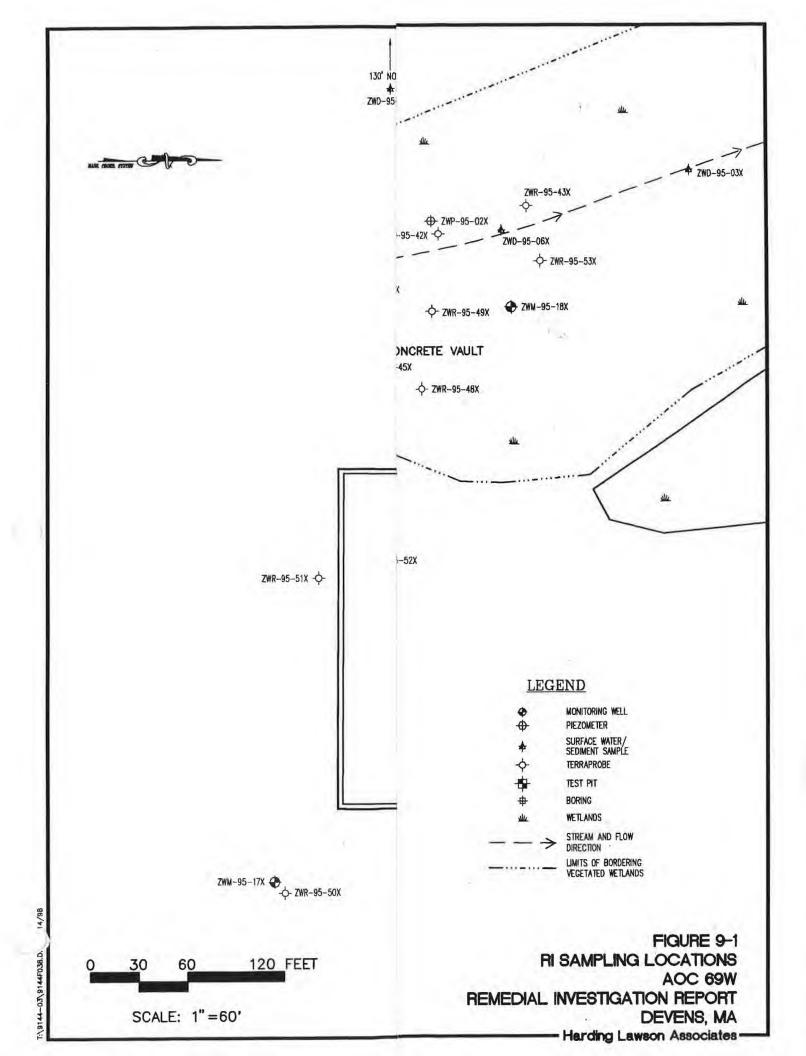
EPH	ug/L	1
ALIPHATICS n-C 9 to n-C 18 n-C 19 to n-C 36 ARDMATICS	= 72 = ND	as a second
n-C 11 to n-C 22	= ND	1 1
VPH	= ND	
VDCs (8260) SVDCs (8270)	= ND = ND	-/

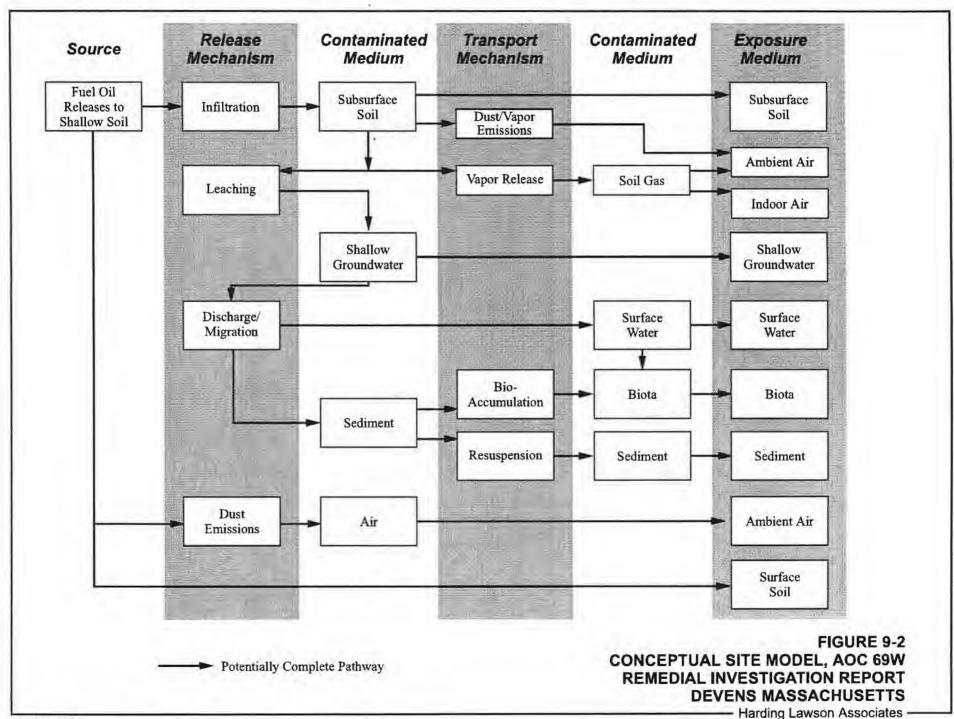
		Site ID: 69W- Units:	-94-13 ug/L
	W-94-12 ug/L	EPH ALIPHATICS n-C 9 to n-C 18 n-C 19 to n-C 36 ARDMATICS n-C 11 to n-C 22	= 43 = ND = 210
	= ND = ND	VPH ALIPHATICS n-C 9 to n-C 12	- 140
,	= ND = NA	ARDMATICS n-C 9 to n-C 10	= 330
		VDCs (8260) Ethylbenzene	- 3J
		SVDCs (8270) 2-Methy Inaphtha lene Naphtha lene Fluorene Phenanthrene	= 13 = 10J = 3J = 2J

DEVENS, MA ABB Environmental Services, Inc.-









W9804011(g)

